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Antonin Kusbach
Utah State University

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TERRESTRIAL ECOSYSTEM CLASSIFICATION IN THE ROCKY MOUNTAINS,
NORTHERN UTAH

by

Antonin Kusbach

A dissertation submitted in partial fulfillment
of the requirements for the degree

of

DOCTOR OF PHILOSOPHY

in

Ecology

Approved:

James N. Long
Co-Major Professor

Helga Van Miegroet
Co-Major Professor

Karel Klinka
Committee Member

Leila M. Shultz
Committee Member

Janis L. Boettinger
Committee Member

Byron R. Burnham
Dean of Graduate Studies

UTAH STATE UNIVERSITY
Logan, Utah

2010

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ABSTRACT

Terrestrial Ecosystem Classification in the Rocky Mountains, Northern Utah

by

Antonin Kusbach, Doctor of Philosophy

Utah State University, 2010

Major Professors: Dr. James N. Long, Dr. Helga Van Miegroet
Department: Wildland Resources

Currently, there is no comprehensive terrestrial ecosystem classification for the central Rocky Mountains of the United States. A comprehensive classification of terrestrial ecosystems in a mountainous study area in northern Utah was developed incorporating direct gradient analysis, spatial hierarchy theory, the zonal concept, and concepts of diagnostic species and fidelity, together with the biogeoclimatic ecosystem classification approach used in British Columbia, Canada.

This classification was derived from vegetation and environmental sampling of both forest and non-forest ecosystems. The SNOwpack TELemetry (SNOTEL) and The National Weather Service (NWS) Cooperative Observer Program (COOP) weather station network were used to approximate climate of 163 sample plots.

Within the large environmental diversity of the study area, three levels of ecosystem organization were distinguished: (1) macroclimatic – regional climate; (2) mesoclimatic, accounting for local climate and moisture distribution; and (3) edaphic - soil fertility. These three levels represent, in order, the L+1, L, and L-1 levels in a spatial hierarchy.

Based on vegetation physiognomy, climatic data, and taxonomic classification of zonal soils, two vegetation geo-climatic zones were identified at the macroclimatic (L+1) level: (1) montane zone with Rocky Mountain juniper and Douglas-fir; and (2) subalpine zone with Engelmann spruce and subalpine fir as climatic climax species.

A vegetation classification was developed by combining vegetation samples (relevés) into meaningful vegetation units.

A site classification was developed, based on dominant environmental gradients within the subalpine vegetation geo-climatic zone. Site classes were specified and a site grid was constructed. This site classification was coupled with the vegetation classification. Each plant community was associated with its environmental space within the site grid. This vegetation-site overlay allowed ecosystems to be differentiated environmentally and a structure, combining zonal, vegetation, and site classifications, forms a comprehensive ecosystem classification.

Based on assessment of plant communities' environmental demands and site vegetation potential, the comprehensive classification system enables inferences about site history and successional status of ecosystems. This classification is consistent with the recent USDA, Forest Service ECOMAP and Terrestrial Ecological Unit Inventory structure and may serve as a valuable tool not only in vegetation, climatic, or soil studies but also in practical ecosystem management.

(221 pages)

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Antonin Kusbach

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CHAPTER 1

INTRODUCTION

Increasingly, society has to deal with complex challenges in management of natural resources, including changes in land use and climate. A particular challenge is that management of natural resources needs to be sustainable. In light of these challenges, it is desirable to adopt an ecosystem approach to natural resource management rather than attempting to manage individual resources (Bailey 2002). This approach requires classification of units of land into ecologically meaningful and relatively uniform segments.

The need for land classification has long been recognized. Many systems have categorized units of land based on a few, and often one, important ecosystem components, e.g., vegetation, animals, microorganisms or some aspect of the physical environment, such as climate or soils (Pojar et al. 1987). For example, the Köppen classification of climates modified by Trewartha (1968) is entirely a climatic classification. The habitat types (Daubenmire 1952), community types (Mueggler 1988) as well as the potential natural vegetation classifications of Küchler (1969) are classifications based strictly on vegetation. Some systems have combined two or more ecosystems components, e.g., ecoregions (Bailey 1998a, b) and the biogeoclimatic ecosystem classification developed in British Columbia, Canada (Pojar et al. 1987). Thus, for classification and mapping purposes, the ecosystem emphasizes vegetation, geomorphology and soils; its meaning is generally geographic (D. Roberts 2008, personal

communication). Currently, however, there is no comprehensive ecosystem classification for the central Rocky Mountains of the United States.

One of the tests of ecosystem classification usefulness is the extent to which it provides insight into factors influencing the distribution of vegetation. In addition to disturbances and biotic interactions (e.g., competition); the physical environment, conceptually represented by climate, soil moisture and soil nutrients, exerts considerable influence on vegetation distribution (e.g., Larsen 1930, Daubenmire 1943, Major 1951, Pojar et al. 1987). It follows, therefore, that a comprehensive ecosystem classification will incorporate vegetation and attributes of the physical environment.

The habitat type and community type concepts are the only “fine-scale” classifications currently available in the central Rocky Mountains. Because they are entirely based on vegetation (Pfister 1976), an explicit link is missing between vegetation and physical environment. An effective organizational structure of terrestrial ecosystems should organize ecosystems hierarchically based on fundamental spatial and functional differences. A comprehensive terrestrial ecosystem classification developed within this hierarchical framework provides that missing link; it integrates three independent classifications: zonal (climatic); vegetation; and site classification (Meidinger and Pojar 1991).

Such a comprehensive ecosystem classification can provide insight into a number of fundamental ecological questions, for example: (1) are there identifiable vegetation-climatic zones as firm altitudinal belts (Daubenmire 1943) in the study area and in the central Rocky Mountains? (2) What are vegetation patterns and important species assemblages in the study area? (3) In addition to climate, what are the environmental

factors associated with the distribution of vegetation in the study area and in the central Rocky Mountains? (4) Is it possible to detect ecosystem dynamics i.e., disturbance history and successional trends based on physical environment? These and other insights have important implications for ecosystem science, outreach, education, and management (e.g., Kotar 1988, Bailey 2006, Sharik et al. 2010).

In this dissertation, a mountaineous study area in northern Utah was selected as representative of a general area of the Rocky Mountains. General goals were: (1) to construct three independent classifications: zonal (climatic); vegetation; and site classification; and (2) integrate these classifications into a comprehensive terrestrial ecosystem classification based on an ecosystem hierarchical framework.

The framework of ecosystem hierarchical organization (the levels L+1, L, and L-1) is introduced in Chapter 2. It explains ecosystem complexity using a bottom-up and top-down approach. Using multivariate ordination, environmental gradients are found and then used to construct ecosystem organizational hierarchy.

Broad vegetation-environmental relationships based on understanding of ecosystem organization, particularly at the highest level L+1, are introduced in Chapter 3. Vegetation patterns together with important environmental factors operating at the L+1 level were assessed in zonal context and zonal (climatic) structure was suggested.

In Chapter 4, units of existing vegetation were identified independently of the physical environment. A more objective and statistical approach using fidelity and diagnostic species concept was applied to vegetation classification.

Site classification revealed important environmental factors at the levels L and L-1 in Chapter 5. This classification was overlayed with independent vegetation classification to

detect relationship between physical environment and existing vegetation, resulting in comprehensive ecosystem classification for 136 sample plots within subalpine climatic zone.

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CHAPTER 2

ORGANIZATION OF TERRESTRIAL ECOSYSTEMS IN THE ROCKY MOUNTAINS, NORTHERN UTAH ¹

Abstract

Currently, there is no comprehensive terrestrial ecosystem classification for the central Rocky Mountains of the United States. Fundamentals of direct gradient analysis and spatial hierarchy theory were used to develop an organizational structure of terrestrial ecosystems in a mountainous study area in northern Utah.

This structure was derived from intensive sampling of both forest (spruce-fir, Douglas-fir, aspen, juniper and mahogany woodland) and non-forested (willow-riparian, shrublands, tall-forb meadows and grasslands) ecosystems. One-hundred and sixty-three plots were described by physiographic features and soil properties such as nutrient pools and dynamics. The SNOwpack TELemetry and the National Weather Service Cooperative Observer Program weather station network were used to approximate climate of sample plots. A complex dataset of environmental variables was analyzed using ordination.

Ecologically meaningful gradients accounted for 54% of the total variance in the environmental data. Within the large environmental diversity in the study area three levels of ecosystem organization were distinguished: (1) macroclimatic; (2) mesoclimatic, accounting for local climate and moisture distribution; and (3) edaphic -

¹ Coauthored by Antonin Kusbach, Helga Van Miegroet, and James N. Long

soil fertility. The climatic data were used to document the importance and dominant role of macroclimate.

The proposed structure is consistent with the recent USDA, Forest Service ECOMAP and Terrestrial Ecological Unit Inventory approach and will serve as a conceptual framework for a comprehensive ecosystem classification.

Introduction

Rocky Mountain ecosystems are diverse in terms of vegetation as well as climate, topography, geology, and disturbances. Understanding these ecosystems requires classification accounting for the underlying complexity in important environmental drivers. Because the multi-scale character of these drivers results in multi-scale ecosystem patterns (Bailey 2002, O'Neill 2005, Wu 2007) development of such a classification requires a conceptual framework based on a meaningful hierarchical organization (e.g., O'Neill et al. 1986, King 1997, 2005, Wu 2007).

There are three dimensions of scale with which to characterize natural elements or events: (1) time; (2) space; and (3) organizational hierarchy (Wu 2007). Time and space have been frequently used as scalars in ecological studies; however, until recently, hierarchical system organization has received little attention in landscape ecology (King 1997, Wu 2007). For example, landscape classifications throughout Western North America have been based almost entirely on spatial scale (Merriam 1890, Daubenmire 1943, Krajina 1965, Küchler 1969, Omernik 1987, Viereck et al. 1992). While space and time are quantifiable (e.g., area in square meters, time in years), organizational hierarchy, as a directional ordering of interacting entities with distinctive process rates, is qualitative

(Wu 2007). Organizational hierarchy is represented by levels, and because ecosystems exist in space and time, levels always correspond to spatial and temporal scales (Wu 2007). Levels are definitional, and typically organized using a triadic structure in which the focal level “L” is between the next lower level “L-1”, and the next higher level “L+1” (O’Neill 1989, Bissonette 1997, King 1997).

The focal level (L) is chosen by the investigator as the level of interest (Allen et al. 1984, O’Neill 1989). It might, for example, represent tissues in a conventional biological system (King 1997). The next lower level (L-1) represents cells making up tissues; the next higher level (L+1) represents organs. In this hierarchical structure, mechanistic or quantitative understanding of the tissue level (L) can be found at the cell level (L-1) because tissue consists of elements (cells) of the lower level (Beckner 1974, O’Neill et al. 1986, Bissonette 1997). But we cannot explain tissue function just by looking at cells (from the bottom-up). We also need a top-down view; to know a tissue’s biological role requires the context of the next higher level (L+1) represented by organs inside an organism (O’Neill et al. 1986, Urban et al. 1987, King 1997). Similarly in landscape ecology, systems are too complex to be explained just from the bottom-up by mechanistic addition (O’Neill and King 1998, Turner et al. 2001, King 2005). Rather, both bottom-up and top-down approaches are necessary to account for complexity of ecosystems (Rowe and Barnes 1994, Chen et al. 2008).

Life or vegetation zones were the foundation of the earliest ecological land classifications in the Rocky Mountains (Merriam 1890, Shreve 1915, Daubenmire 1943). In this approach, the zones are characterized by major plant species (Whittaker 1972, Long 1994). Life zones provide a useful overview of vegetation pattern across large

geographic areas (thousands of km²) such as San Francisco Peaks, Arizona (Fig. 2.1a, b); however, these zones are too broad to capture important details of vegetation pattern along complex environmental gradients such as elevation, topography, and soils (Fig. 2.1c) (Daubenmire 1943). Peet's (2000) vegetation zonation, based on Whittaker and Niering (1965), incorporates topographical features as a potential second level of organization and makes explicit the influence of elevation and topography-moisture gradients on vegetation zonation. While it is an enhancement on simple life zones (altitudinal belts *sensu* Merriam) this two-level approach to vegetation zonation still does not completely capture important details of ecosystem complexity.

Climate, soil moisture, and soil nutrients are generally believed to be fundamental environmental factors influencing the distribution of vegetation and have been frequently invoked in ecosystem classifications (e.g., Pogrebnjak 1930, Hills 1952, Krajina 1965, Pojar et al. 1987, Meidinger and Pojar 1991). We used direct gradient analysis and spatial hierarchy theory (in the form of a triadic structure) to explicitly assess the roles of the physical environment (Pojar et al. 1987) in the distribution of vegetation within a complex landscape typical of the central Rocky Mountains to improve our conceptualization of these semi-arid ecosystems.

Our overarching goal is to develop a conceptual framework, based on hierarchical organization, for ecosystem classification of the northern Wasatch Range. Our specific objectives are: (1) to determine the important environmental gradients influencing ecosystem patterns; and (2) use these gradients to characterize ecologically meaningful levels of ecosystem organization.

Methods

Study area

Franklin Basin (FB) is a montane-subalpine area, approximately 15,000 ha in size, situated between the Bear River Range and the Wasatch Range in the central Rocky Mountains on the Utah and Idaho border. Smaller in size (ca 1000 ha), the T.W. Daniel Experimental Forest (TWDEF) is situated on the high ridge plateau of the Wasatch Range (10 km to the southeast of FB). Logan Canyon is lower in elevation and together with FB and TWDEF makes up the study area (ca 20,000 ha, and ca 1,400 m of vertical extent) (Fig. 2.2).

According to Bailey (1998a) and McNab et al. (2007), the study area occurs within M331 Southern Rocky Mountains Steppe-Open Woodland-Coniferous Forest-Alpine Meadow Province, “D” Overthrust Mountain Section, “n” Northern Wasatch Range, and “o” Bear River Front Range Subsections. The mean total annual precipitation ranges from about 720 to 1250 mm and mean annual air temperature ranges from 2.4 to 5.7 °C for Temple Fork, Tony Grove Lake, Franklin Basin, and Utah State University (USU) Doc Daniel weather stations (<http://www.wcc.nrcs.usda.gov/snow/>).

According to Köppen’s climatic classification modified by Trewartha (1968), the study area is characterized by the microthermal/cold snowy-forest climate; while the lower portion occurs in the cool temperate climate, and the highest portion of the study area occurs in the subalpine-boreal climatic type.

The terrain is mountainous, rocky and steep with occasional flat to gently sloping high ridge-plateaus and benches. The elevation ranges from 1590-3060 m across the three study sites. The highest area of the Bear River Range was glaciated during the

Pleistocene as manifested by glacial geomorphologic features like moraines, U-shaped valleys, erratics, and irregular glacial deposits (Young 1939, Degraff 1976). The study area is mostly built from calcareous sedimentary rocks (limestone, dolomite) with inter-layered quartzite, and from Tertiary sediments consisting of grit, conglomerate, and siltstone of the Wasatch Formation at the TWDEF. The soils are formed in residuum, colluvium, alluvium, glacial till and outwash, and occur on diverse landforms such as cliffs, talus slope, moraines, karst valleys, mountain slopes, landslides, plains, valleys, depressions, ravines, and wetlands (Schoeneberger et al. 2002).

Over half of the study area is occupied by forest ecosystems including Engelmann spruce (*Picea engelmannii*), subalpine fir (*Abies lasiocarpa*), Douglas-fir (*Pseudotsuga menziesii*), aspen (*Populus tremuloides*); woodland ecosystems including mountain mahogany (*Cercocarpus ledifolius*) and Rocky Mountain juniper (*Juniperus scopulorum*); and riparian, mostly willow (*Salix spp.*) ecosystems. Areal extent and basic types of forest ecosystems are very similar to those of the mid-1800s at the beginning of European settlement. Substantial changes in fire regimes, often in combination with cutting and grazing, have led to dramatic changes in the structure and the age-class distribution of forest stands. In many places, 100- to 140-year-old stands are now predominant (Long 1994). Forests in the study area are thus characterized by mid- and late-seral stages where potential climax tree species are easily recognizable and forest understory is usually well developed (Pfister and Arno 1980). Non-forested ecosystems include willow-riparian communities, shrublands (*Artemisia spp.*), meadows and grasslands, which may represent stable or temporary communities. Despite human

impacts in last 120 years, the study area is considered as relatively natural in terms of plant species composition (Bird 1964).

Data collection

Sampling was intensive enough to capture as much ecosystem variation as possible focusing on all major existing plant communities (Brohman and Bryant 2005) occupying all major landforms (Schoeneberger et al. 2002), but avoiding ecotones and recently disturbed (burned, logged, damaged by insects) areas. One-hundred-sixty-three sample plots were established within the study area in the summers of 2006 and 2007. A stratified (based on vegetation physiognomy) fixed (subjective selection) sampling design was used with sample plot size of 1000 m² for forest and 100 m² for non-forested ecosystems (Brohman and Bryant 2005). Three replicate plots were considered the minimum for defining a vegetation unit. We anticipated 50-60 units based on a preliminary reconnaissance of the study area. To distinguish among units, we followed recommendations of Grossman et al. (1998).

As the study focused on environmental gradients; we did not examine historical gradients and biotic interactions (e.g., plant competition, allelopathy, mycorrhizae). To characterize the environmental gradients we described each sample plot in terms of relatively static or constant attributes i.e., physiographic variables such as elevation, slope aspect, slope gradient, topographic position, and slope shape (Lotspeich 1980); dynamic attributes such as O and A horizon thickness, humus form, pH, nutrient pools characterizing relatively slow processes; and attributes such as nutrient supply rates describing relatively fast processes (Table 2.1). One soil pit was dug in each sample plot

to the unweathered parent material or permanent water table. Description and sampling of pedons followed practices and terminology of the National Cooperative Soil Survey (Soil Survey Staff 1999, 2006, Schoeneberger et al. 2002). Humus form was identified following Green et al. (1993).

One composite soil sample from 0-30 cm was collected from a pedon face in each plot, air dried and sieved (< 2 mm), and the fine fraction analyzed for texture classes (sandy, loamy, clayey) using the feel-method (Thien 1979), for pH (1:1 soil in water) using a Corning pH analyzer, and CaCO_3 content by treating the samples with HCl using a mercury manometer (Loeppert and Suarez 1996). The total C and N concentration were determined by dry combustion using a LECO CN analyzer (LECO TrueSpec C/N, Leco Corp., St. Joseph, MI). A static-absolute nutrient availability index (SNAI), i.e., cation pool “snapshot”, was determined by extracting exchangeable cations with 1 M HN_4Cl at pH 7.0 using a mechanical vacuum extractor (Holmgren et al. 1977) and cation analysis of extractant using an inductively coupled plasma spectrophotometer (ICP) (Iris Advantage, Thermo Electron, Madison, WI). Extractable P (PO_4) was determined by the Olsen P method (Olsen et al. 1954) using sodium bicarbonate extraction, followed by P analysis using Thermo Electron Spectronic 20 Genesys spectrophotometer. Total mineralizable nitrogen was determined from 7-day anaerobic incubation and extraction with 2 M KCl (Keeney and Bremner 1966) followed by NH_4 analysis (Lachat Quickchem 8000, Flow Injection Analyzer).

To determine a dynamic-relative nutrient availability index (DNAI) (Qian and Schoenau 2002), plant root simulators (PRSTM-probes; Western Ag Innovations, Inc., Saskatoon, Canada), a combination of anion and cation exchange membranes, were

buried vertically into the mineral soil at each site for 6 weeks (during September and November). PRSTM-probes were cleaned and sent to Western Ag Innovations for extraction and chemical analysis including Ca, Mg, K, S, Fe, Mn, Zn, Cu, Pb, Al, NH₄ cations, and NO₃ and PO₄ anions (Table 2.1). One of 163 exchange membrane sets was not recovered.

Climatic data such as air temperature, precipitation, soil temperature (depth of 50 cm), and soil moisture (depth of 20 cm and 50 cm) for the northern Wasatch Range (corresponding with M331D Section in McNab et al. (2007)) were obtained from nearby weather stations to approximate ambient and soil climate of the sample plots. Both, the Natural Resources Conservation Service (NRCS) SNOwpack TELemetry (SNOTEL; <http://www.wcc.nrcs.usda.gov/snow/>) and the National Weather Service Cooperative Observer Program (COOP; <http://www.nws.noaa.gov/om/coop/>) station networks provide long term observations for air temperatures and precipitation (>10 years). Soil temperature and moisture at SNOTEL sites were available for 6 years (2003-2008) and there were two COOP stations with soil temperature measurements for the northern Wasatch Range. Accuracy of the analysis may therefore be limited by the short data record. Nevertheless, these data were an important source of information in this analysis.

Data analysis

We used direct gradient analysis (e.g., Whittaker 1973, Austin et al. 1984) to order the sample plots according to environmental attributes (McCune and Grace 2002).

We analyzed a dataset consisting of 163 sample plots and 43 variables. Because of the complex environmental character of the dataset we used Principal components

analysis (PCA) (Pearson 1901) in order to: (1) reduce dimensionality of the dataset, i.e. to represent a large dataset with a smaller number of composite variables; (2) assess the composite variables as significant principal components (PCs); and (3) interpret the significant PCs as environmental gradients (McCune and Grace 2002). Orthogonal rotations and correlation type of a cross-products matrix were used to get independent, mutually uncorrelated PCs (Lattin et al. 2003).

Slope aspect, measured in azimuth degrees and corrected for magnetic declination, was converted from a circular to a linear form representing "coolness" (Roberts and Cooper 1989). The maximum aspect value ($av = 1.0$, the coolest conditions) occurs at 30 degrees aspect, and the minimum ($av = 0.0$, the warmest conditions) at 210 degrees aspect. We transformed the variables with $|\text{skewness}| > 1$ to be close to multivariate normality, standardized the data by adjustment to standard deviate (z -scores), and checked the dataset for outliers using a cutoff of 2.0 standard deviations from the grand mean (McCune and Grace 2002).

Significance of PCs was tested by a graphical approach (Cattell 1966) and Monte Carlo randomization test (McCune and Mefford 2006). We calculated correlation coefficients (loadings) with each ordination axis, the linear (parametric Pearson's r) and rank (nonparametric Kendall's τ) relationships between the ordination scores and the observed variables. Our use of r and τ is suggested to be, even in relatively small datasets, more conservative than p -values for the null hypothesis of no relationship between ordination scores and variables (McCune and Grace 2002). We set the threshold for r and $\tau > 0.35$.

Additionally, linear regressions were used to assess: (1) the role of elevation in the PCs; (2) the relationship of climatic variables with elevation; (3) the relationship of soil moisture with elevation and precipitation; and (4) the relationship of understory vegetation cover with soil development represented by soil depth, coarse rock fragments content, and A horizon thickness. The variables were transformed for normality by power or logarithmic transformation when necessary.

R software, ver. 2. 7. 2. (<http://www.r-project.org/>) and PC-ORD (McCune and Mefford 2006) were used in the analysis.

Results

Principal components analysis was run three times; the first run was based on the entire original dataset (163 sample plots, 43 variables). The second PCA was run on the reduced dataset (163 sample plots, 5 variables) retaining only the physiographic variables (elevation, topographic position, slope gradient, slope aspect, and slope shape). Assuming that elevation is a surrogate for the highest level of organizational hierarchy (L+1 level in the triadic structure) and may affect additional analysis of the lower levels (L and L-1), we dropped it from the original dataset in the third PCA run.

In the first PCA run, Cattell's approach with a pronounced elbow just after the fifth PC and variances of remaining PCs (proportions $\leq 4\%$ for each) declining approximately linearly, suggested five statistically significant PCs (Fig. 2.3). Also the Monte Carlo test with 5000 randomizations showed significant *p*-values for the first five PCs (Table 2.2).

Elevation was the only physiographic variable whose loadings were significant and stable i.e., they did not fluctuate highly for the four significant PCs (e.g., max. $r = 0.40$

for PC1 and min. $r = 0.28$ for PC2, PC4; and max. $\tau = 0.28$ for PC1 and min. $\tau = 0.20$ for PC2) in contrast with e.g., topographic position where loadings ranged from $r = 0.78$ for PC2 to $r = 0.07$ for PC3, PC4 (Table 2.3). Multiple regression also demonstrated a relationship between elevation and the first four PCs ($F = 4.98$, $R^2 = 0.52$, PC1 - PC4: $p < 0.001$).

The second PCA run provided physiographic insight. A single significant principal component (PC1 as the axis 1) accounted for 45% of the total variance (Table 2.2). A joint plot (Fig. 2.4), a diagram of radiating lines (vectors), indicated strength (vector length) and direction (vector angle) of relationships of the variables with ordination axes (McCune and Grace 2002). Topographic position, slope gradient, and slope shape together reflected strong topographical/morphometric meaning of the axis 1. Elevation had a weaker but nevertheless important influence on this axis (Table 2.3).

General importance of elevation was also documented by regressions with the climatic data. Elevation is a predictor for mean annual air temperature (MAAT: $p < 0.001$, $R^2 = 0.67$), mean total annual precipitation (MTAP: $p < 0.001$, $R^2 = 0.39$), and soil temperatures (mean annual, MAST: $p < 0.001$, $R^2 = 0.61$; mean summer, MSST: $p < 0.001$, $R^2 = 0.43$). Mean winter soil temperature (MWST) was poorly correlated with elevation ($p < 0.8$, $R^2 = -0.03$) most likely because of snowpack insulation (Fig. 2.5a, b). Soil moisture at 20 and 50 cm depths was little affected either by elevation or precipitation (Fig. 2.5c, d).

To gain ecological insight into the lower hierarchical levels (L and L-1) and discriminate among them, we reran PCA without elevation. In this third PCA run (163 sample plots, 42 variables), five statistically significant PCs accounted for 60% of the

total variance in the dataset (Table 2.2). We thus reduced 42 variables to five statistically significant PCs, thereby reducing dimensionality of the problem by 88%. When comparing the first and third PCA runs, we obtained the same number of statistically significant PCs but the numerical summary was better for the third PCA run (Table 2.2). There were also differences among loadings between the first and third PCA runs (e.g., pH, NO₃ DNAI, Fe and Mn SNAI, Table 2.3). Because of improved results, the third PCA run was used for the interpretation of principal components.

In our analysis the first two PCs or gradients (collectively explaining 35% of the variance in the dataset) might not be adequate to explain the complex environment. On the other hand, there is the risk of interpretation of weak patterns with more dimensions (McCune and Grace 2002). The fifth PC, despite its statistical significance, is not readily interpretable as there are just five significant loadings for this PC (Table 2.3). Therefore, instead of five statistically significant PCs, we interpreted four ecologically meaningful PCs (Fig. 2.3). Based on significance of the loadings, the first principal component (PC1) was associated with parent material and soil properties (soil color, pH, CaCO₃ content); total N and C; Ca, Mg and K DNAIs; and N, Ca, Mg and metals SNAIs (Table 2.3). We interpreted PC1 as an indicator of **soil fertility**, differentiating between rich (sufficiency of macronutrients) and poor (deficiency of macronutrients) ecosystems.

The second principal component (PC2) was related to topographic position, slope gradient, slope shape with soil properties (soil depth, coarse rock fragment content, CaCO₃ content) affecting soil moisture conditions (indicated by water table and mottles). It was interpreted as **local topography** accounting for soil moisture. Because we did not measure soil moisture directly we did not consider it as an explicit environmental

gradient. The interpretation was consistent with interpretation of PC1 in the second PCA run. PC2 generally differentiates between “caplands” and “cuplands” (Devlin et al. undated), and steepness and gentleness of the terrain. For instance, one may find completely different plant communities within a short distance (e.g., meters) where a moist, concave pocket of aspen changes abruptly into a dry, rocky, convex island of mountain mahogany.

The third principal component (PC3) was associated with soil organic matter (SOM) (O and A horizon thickness, humus form), mineralizable N DNAI, NO_3 and NH_4 DNAIs (Table 2.3). We interpreted this PC as indicative of **microbial activity**, influencing SOM decomposition rate. This rate is conditioned by overall climate (as represented by elevation in the first PCA run), and expressed by surface soil horizon properties (O and A horizon thickness, humus form) as well as by indicators of N availability. It influences the nutrient/chemical environment by either NH_4 mineralization or nitrification.

The last interpretable principal component (PC4) was associated with coarse rock fragment content and soil depth, with DNAIs of mobile nutrients (N, S) and relatively immobile metals (Fe, Mn, Zn), and K SNAI. It was interpreted as **soil development**, reflecting the amount of parent material residuum (soil “rockiness”) and soil permeability. More developed and older soils are deeper, contain less coarse rock fragments. These soils have high total porosity and high water holding capacity because water movement in soil micropores is slow. In contrast, less developed soils are younger and shallower, and contain high amounts of coarse rock fragments. These soils have low total porosity but high macroporosity, resulting in low water holding capacity, but fast

water infiltration and percolation via macropores. Rapid water flow in these soils creates pulses of plant-available nutrients in soil solution.

These less developed soils, on the other hand, had less understory vegetation cover. The high DNAIs for N, S and metals captured by PRSTM-probes combined with low nutrient uptake potential associated with low understory vegetation cover, thus imply possible greater nutrient leaching loss from less developed, young, skeletal soils.

Discussion

Development of a hierarchical organization from the complex dataset represented a compromise between complexity, understandability, and applicability. We synthesized functionally similar principal components to environmental gradients and used them as the definitional levels of ecosystem organization (Table 2.4).

That elevation is strongly associated with climate has long been observed (e.g., Merriam 1890, Larsen 1930, Daubenmire 1943). Our data revealed that elevation: (1) is superimposed over all the meaningful principal components; (2) differs from the other physiographic variables, which have topographical or morphometric meaning (Lotspeich 1980); and (3) is strongly correlated with climate. Therefore, we consider elevation an appropriate surrogate for **regional climate** as the macroclimatic level (Major 1951, Bailey 2002).

Analyzed importance of elevation was consistent with altitudinal zonation of mountains (e.g., Krajina 1965, Bailey 1998a, b). In mountains, ecosystems change with elevation because of the change of climate; the use of the term regional climate/macrocclimate in an elevation sense is therefore appropriate. Elevation or

macroclimatic zones should be seen as relatively narrow vertically stacked belts with abrupt climatic transition rather than gradual changes over vast horizontally extended areas.

The other physiographic variables, topographic position, slope gradient and slope shape were significant predictors of PC2 (local topography). This gradient was an important local topo-climatic attribute because it is modified by local topography (Major 1951, Thornthwaite 1954) finally accounting for soil moisture conditions. We considered it a surrogate for **local climate** as the mesoclimatic level (Major 1951).

Soil properties were significant predictors of PC1, PC3 and PC4. Because of the orthogonality of ordination axes, soil fertility was not correlated with soil development as well as with microbial activity. However, these gradients were in fact functionally similar in their impact on plant nutrition; they affect presence or absence of plant-available nutrients. We considered them a surrogate for **soil fertility** as the edaphic level.

We differentiated among the levels of ecosystem organization: regional climate (L+1); local climate (the focal level, L); and soil fertility (L-1). We superimposed local climate over fertility because of their cause-effect relationship; soils are more dynamic (i.e., they are “faster”) and act at a finer temporal and spatial scale (Urban et al. 1987, King 1997). At the focal level, ecosystems are affected by meso- or local climate. Ecosystems in the context of the level L-1 are differentiated by fertility and together with the focal level ecosystems they build up the L+1 level (bottom-up/inductive approach). To understand these lower level (L, L-1) ecosystems, they must also be placed in the context of macroclimate (L+1) (top-down/deductive approach).

For example, the Engelmann spruce-subalpine fir (S-F) zone stretches across the levels L+1, L, and L-1 (Fig. 2.6). The ecosystem organization enabled stratification of this extremely environmentally diverse S-F zone. We know now that poor and fertile S-F communities from the fertility level (L-1) can be found in different conditions of the local climate (L) (e.g., on warm/dry vs. cool/moist slopes or valleys). As a result, S-F communities appear in very different elevations and local climates because of the compensating effects of environmental factors on plants (Pojar et al. 1987). Thus, S-F communities occur as a high-elevation regional ecosystem (Klinka and Chourmouzis 1999, K. Klinka 2009, personal communication); however, they may also occur in lower elevations because local environmental conditions modify the influence of low-elevation regional climate (Major 1969). For example, S-F communities descend on cool shady slopes or as riparian/wetland communities along valley bottoms (Fig. 2.1c).

Our hierarchical ecosystem organization corresponds well to lower elements of the Hierarchical Framework of Ecological Units (Cleland et al. 1997, ECOMAP 2007) and the Terrestrial Ecological Unit Inventory (TEUI) Technical Guide (Winthers et al. 2005). Regional climate represented by elevation matches landscape and even subregion, local climate and soil fertility match land unit of the ECOMAP/TEUI standard (Table 2.4).

We made no *a priori* assumptions about environmental factors associated with the various levels in our organizational structure. Rather, the apparent roles of climate, topography, and soil nutrients within the landscape hierarchy were derived from the data (O'Neill and King 1998).

We anticipate the organizational structure will contribute to implementation of spatial hierarchy theory (O'Neill 2005) to the practice of classification as well as to its

application into ecosystem management (e.g., by landscape managers or foresters) (Sharik et al. 2010).

Summary and conclusions

We found ordination to be a valuable technique to organize ecosystems and detect multiple scales of pattern based on interpretation of the meaningful principal components. By using PCA we were able to confirm the environmental heterogeneity of the study area, derive important environmental gradients influencing ecosystem patterns within that area, and synthesize these gradients into ecologically meaningful levels of ecosystem organization. These are: (1) macroclimate as the level L+1; (2) mesoclimate as the focal level (L); and (3) soil fertility as the level L-1.

Our hierarchical ecosystem organization represents an improvement on earlier land classifications in the Rocky Mountains by adding another dimension. This structure is consistent with the lower elements of the ECOMAP/TEUI standard.

We conclude that this organization can be used as a framework for additional structuring e.g., site (ecotope, habitat) classification. We intend to use this hierarchy in building a comprehensive ecosystem classification system that will serve as a tool not only for communication in ecosystem research but also for practical ecosystem management.

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Table 2.1. Research variables. Dynamic nutrient availability index (DNAI) indicated by ‘d’, static nutrient availability index (SNAI) indicated by ‘s’ in abbreviations. NA-not applicable.

Variable	Abbreviation	Units/Values
Elevation	elev	meters
Topographic position	topos	1-crest,shoulder, 2-back slope, 3-foot slope, 4-flat (<5%), 5-toeslope, 6-depression
Slope gradient	sl	%
Slope aspect	av	aspect values 0-1 (Roberts and Cooper 1989)
Slope shape	shape	1-convex, 2-linear, 3-concave
Parent material	parmat	1-quartzite, 2-Wasatch Formation, 3-till, 4-limestone or dolomite, 5-colluvium, 6-alluvium
Soil O-horizon depth	Ohor	centimeters
Soil A horizon depth	Ahor	centimeters
Humus form	hum	values 1-17; e.g., 1-fibrimor, 10-mormoder, 14-rhizomull, 17-no humus (Green et al. 1993)
Soil depth	sdepth	centimeters
Coarse rock fragment content	RF	% volumetric
Soil water table	wtable	1- up to 30cm depth, 2- 30-80cm, 3- 80-150cm, 4- no water table
Soil mottles	mottles	1- up to 30cm depth, 2- 30-80cm, 3- 80-150cm, 4- no mottles
Soil color value	cvalue	1-7 according to Munsell® notation
Soil texture	text	1-sandy, 2-loamy, 3-clayey
Soil pH	pH	1-14 pH scale
Calcium carbonate content	CaCO ₃	%
Total nitrogen	Nox	%
Total carbon	Cox	%
Carbon nitrogen ratio	C.N	NA
Mineralizable nitrogen DNAI	Nmin_d	microgram/10 cm ² /6weeks
Nitrate DNAI	NO ₃ _d	microgram/10 cm ² /6weeks

Ammonium DNAI	NH4_d	microgram/10 cm ² /6weeks
Calcium DNAI	Ca_d	microgram/10 cm ² /6weeks
Magnesium DNAI	Mg_d	microgram/10 cm ² /6weeks
Potassium DNAI	K_d	microgram/10 cm ² /6weeks
Phosphorus DNAI	P_d	microgram/10 cm ² /6weeks
Iron DNAI	Fe_d	microgram/10 cm ² /6weeks
Manganese DNAI	Mn_d	microgram/10 cm ² /6weeks
Zinc DNAI	Zn_d	microgram/10 cm ² /6weeks
Sulphur DNAI	S_d	microgram/10 cm ² /6weeks
Aluminum DNAI	Al_d	microgram/10 cm ² /6weeks
Mineralizable nitrogen SNAI	Nmin_s	milligram/kilogram
Ammonium SNAI	NH4_s	milligram/kilogram
Calcium SNAI	Ca_s	milligram/kilogram
Magnesium SNAI	Mg_s	milligram/kilogram
Potassium SNAI	K_s	milligram/kilogram
Phosphorus SNAI	P_s	milligram/kilogram
Iron SNAI	Fe_s	milligram/kilogram
Manganese SNAI	Mn_s	milligram/kilogram
Zinc SNAI	Zn_s	milligram/kilogram
Sulphur SNAI	S_s	milligram/kilogram
Aluminum SNAI	Al_s	milligram/kilogram

Table 2.2. PCA summary for ten principal components. Ecologically meaningful PCs are in bold.

1st run	PC1	PC2	PC3	PC4	PC5	PC6	PC7	PC8	PC9	PC10
Eigenvalue	8.18	6.88	4.78	3.2170	2.33	1.60	1.53	1.23	1.20	1.16
% of Variance	19.03	15.99	11.11	7.48	5.41	3.73	3.56	2.87	2.78	2.71
Cumulative % of Var.	19.03	35.02	46.13	53.61	59.03	62.75	66.31	69.17	71.96	74.66
<i>p</i> - value	0.0002	0.0002	0.0002	0.0002	0.0002	0.9832	0.9926	1.0000	1.0000	1.0000
3rd run										
Eigenvalue	8.05	6.80	4.66	3.15	2.33	1.54	1.48	1.23	1.19	1.15
% of Variance	19.17	16.20	11.10	7.49	5.54	3.66	3.52	2.93	2.82	2.74
Cumulative % of Var.	19.17	35.37	46.47	53.96	59.50	63.16	66.68	69.61	72.43	75.17
<i>p</i> - value	0.0002	0.0002	0.0002	0.0002	0.0002	1.0000	1.0000	1.0000	1.0000	1.0000
2nd run										
Eigenvalue	2.26	1.15	0.68	0.61	0.30					
% of Variance	45.14	23.05	13.64	12.16	6.01					
Cumulative % of Var.	45.14	68.19	81.83	93.99	100.00					
<i>p</i> - value	0.0002	0.0804	1.0000	1.0000	1.0000					

Table 2.3. PCA loadings. Significant Pearson's (r), and Kendall's (τ) coefficients are in bold. Both significant r and τ express a significant variable for the particular PC (shaded). Variables are defined in Table 2.1. NA - not applicable.

Variable	PCA first run										PCA third run					PCA second run		
	PC1		PC2		PC3		PC4	PC5	PC1		PC2		PC3		PC4	PC5	PC1	
	r	tau	r	tau	r	tau	r	r	r	tau	r	tau	r	tau	r	r	r	tau
elev	-0.40	-0.28^a	0.28^a	0.20^a	0.36	0.25^a	-0.28^a	-0.01	NA	NA	NA	NA	NA	NA	NA	NA	0.48	0.36
topos	0.20	0.10	-0.78	-0.62	0.07	-0.05	0.07	-0.36	0.14	0.04	-0.79	-0.63	0.03	-0.05	0.08	-0.36	-0.89	-0.76
sl	-0.20	-0.03	0.57	0.45	0.18	0.12	-0.05	0.15	-0.15	0.01	0.58	0.45	0.21	0.13	-0.08	0.15	0.75	0.54
av	-0.23	-0.19	-0.16	-0.13	0.29	0.20	0.18	-0.28	-0.23	-0.19	-0.16	-0.13	0.30	0.20	0.16	-0.28	-0.23	-0.15
shape	0.16	0.08	-0.62	-0.50	-0.02	-0.05	0.23	-0.37	0.11	0.04	-0.63	-0.51	-0.05	-0.06	0.26	-0.37	-0.79	-0.66
Ohor	0.13	0.09	-0.30	-0.21	0.75	0.42	0.17	0.16	0.13	0.08	-0.34	-0.23	0.75	0.43	0.13	0.16		
Ahor	0.27	0.21	-0.21	-0.11	-0.58	-0.45	0.09	-0.30	0.23	0.19	-0.21	-0.12	-0.60	-0.46	0.15	-0.30		
hum	-0.05	-0.06	0.25	0.19	-0.69	-0.42	-0.22	-0.25	-0.05	-0.05	0.27	0.20	-0.70	-0.43	-0.17	-0.25		
sdepth	0.02	-0.03	-0.71	-0.53	-0.20	-0.15	0.36	-0.20	-0.05	-0.08	-0.70	-0.53	-0.21	-0.15	0.39	-0.20		
RF	-0.26	-0.13	0.61	0.45	0.18	0.15	-0.43	0.30	-0.20	-0.09	0.62	0.46	0.19	0.14	-0.46	0.30		
parmat	0.63	0.41	-0.33	-0.26	-0.20	-0.16	0.21	-0.20	0.60	0.37	-0.36	-0.28	-0.23	-0.17	0.25	-0.20		
wtable	-0.36	-0.20	0.54	0.40	-0.42	-0.24	0.29	0.25	-0.33	-0.18	0.59	0.41	-0.37	-0.22	0.30	0.25		
mottles	-0.31	-0.18	0.62	0.47	-0.37	-0.20	0.23	0.33	-0.28	-0.15	0.66	0.49	-0.32	-0.17	0.24	0.33		
cvalue	-0.77	-0.63	-0.12	-0.16	0.02	0.10	0.11	-0.08	-0.78	-0.63	-0.06	-0.13	0.05	0.11	0.09	-0.08		
text	0.57	0.34	-0.23	-0.13	0.05	-0.05	0.23	0.06	0.54	0.33	-0.27	-0.15	0.04	-0.03	0.24	0.06		
pH	0.67	0.54	0.56	0.36	-0.02	-0.03	0.04	-0.01	0.70	0.57	0.51	0.33	-0.01	-0.03	0.04	-0.01		
CaCO3	0.38	0.43	0.64	0.57	0.18	0.12	-0.07	-0.12	0.43	0.48	0.60	0.55	0.19	0.12	-0.08	-0.11		
Nmin_d	0.03	0.04	0.19	0.12	-0.56	-0.39	-0.63	-0.08	0.04	0.04	0.20	0.13	-0.59	-0.42	-0.60	-0.08		
Nox	0.79	0.55	-0.15	-0.05	-0.09	-0.21	-0.11	0.29	0.77	0.52	-0.21	-0.07	-0.13	-0.22	-0.07	0.29		
NO3_d	0.08	0.08	0.20	0.12	-0.65	-0.45	-0.55	-0.07	0.08	0.08	0.21	0.13	-0.68	-0.47	-0.51	-0.07		
NH4_d	-0.27	-0.18	-0.01	-0.02	0.39	0.27 ^b	-0.26	-0.10	-0.25	-0.17	-0.01	-0.02	0.38	0.26 ^b	-0.29	-0.10		

Cox	0.78	0.60	0.12	0.16	0.17	0.02	-0.07	0.24	0.80	0.63	0.04	0.13	0.13	0.01	-0.05	0.24
C/N	0.09	0.04	0.45	0.24	0.46	0.38	0.07	-0.10	0.15	0.08	0.42	0.22	0.46	0.38	0.06	-0.10
Ca_d	0.57	0.40	-0.01	0.01	-0.33	-0.24	-0.18	0.07	0.56	0.38	-0.04	0.00	-0.35	-0.25	-0.17	0.07
Mg_d	0.51	0.36	0.24	0.11	0.20	0.08	-0.29	-0.26	0.55	0.38	0.18	0.09	0.17	0.07	-0.28	-0.26
K_d	-0.55	-0.38	-0.20	-0.22	-0.43	-0.23	0.18	0.18	-0.58	-0.41	-0.14	-0.19	-0.40	-0.22	0.19	0.18
P_d	-0.20	-0.17	-0.39	-0.29	-0.47	-0.34	-0.05	0.26	-0.25	-0.20	-0.36	-0.27	-0.48	-0.34	-0.02	0.26
Fe_d	0.12	-0.10	-0.66	-0.44	0.09	-0.09	-0.54	-0.03	0.08	-0.13	-0.68	-0.43	0.03	-0.12	-0.53	-0.03
Mn_d	-0.20	-0.23	-0.51	-0.32	0.21	0.05	-0.49	0.01	-0.23	-0.25	-0.51	-0.30	0.18	0.04	-0.51	0.00
Zn_d	0.16	0.09	-0.31	-0.15	-0.36	-0.27	-0.55	0.02	0.13	0.07	-0.31	-0.15	-0.41	-0.29	-0.52	0.02
S_d	-0.08	-0.12	-0.45	-0.27	0.20	0.08	-0.54	-0.13	-0.10	-0.13	-0.46	-0.26	0.15	0.06	-0.55	-0.13
Al_d	-0.11	-0.06	0.14	0.10	-0.25	-0.15	-0.17	-0.15	-0.11	-0.06	0.16	0.11	-0.24	-0.14	-0.18	-0.15
Ca_s	0.87	0.65	0.04	0.10	0.01	-0.07	0.13	0.28	0.87	0.65	-0.02	0.07	-0.01	-0.06	0.14	0.28
Mg_s	0.81	0.65	0.22	0.12	0.29	0.08	-0.02	-0.03	0.84	0.68	0.14	0.09	0.27	0.08	-0.02	-0.03
K_s	0.25	0.18	-0.19	-0.13	-0.44	-0.32	0.37	0.21	0.21	0.15	-0.17	-0.13	-0.42	-0.31	0.38	0.21
NH4_s	0.31	0.16	-0.45	-0.28	0.18	0.06	-0.13	0.34	0.29	0.14	-0.49	-0.30	0.14	0.04	-0.11	0.34
Nmin_s	0.79	0.57	-0.16	-0.02	0.10	-0.06	-0.15	0.25	0.79	0.56	-0.23	-0.05	0.05	-0.08	-0.12	0.25
P_s	-0.09	-0.13	-0.43	-0.33	-0.38	-0.31	0.18	0.41	-0.14	-0.15	-0.41	-0.33	-0.39	-0.32	0.22	0.42
Al_s	-0.70	-0.55	-0.24	-0.20	0.25	0.22	-0.07	-0.03	-0.70	-0.56	-0.20	-0.17	0.26	0.22	-0.09	-0.03
Fe_s	-0.32	-0.24	-0.22	-0.16	0.43	0.34	-0.07	0.32	-0.33	-0.25	-0.21	-0.14	0.44	0.35	-0.12	0.32
S_s	0.23	0.14	-0.50	-0.30	0.01	-0.02	-0.16	0.52	0.19	0.11	-0.52	-0.32	-0.03	-0.04	-0.15	0.52
Mn_s	-0.40	-0.31	-0.54	-0.43	-0.04	-0.04	0.01	0.37	-0.44	-0.35	-0.52	-0.42	-0.06	-0.06	0.04	0.37
Zn_s	-0.47	-0.36	-0.27	-0.20	0.05	0.06	-0.20	0.25	-0.48	-0.37	-0.25	-0.19	0.03	0.04	-0.19	0.25

^a significant loadings relatively to *elev* maximum (0.40)

^b important contrast of ammonium with nitrate nutrient supply rate

Table 2.4. Comparison of USDA Forest Service ecological land classification with ecosystem organization and the environmental gradients.

USDA Forest Service, ECOMAP 2007; TEUI 2005		Ecosystem organization	
Scale	Ecological unit	Definitional levels (Environmental gradients)	Environmental principal components
Ecoregion	Domain		
	Division		
	Province		
Subregion	Section		
	Subsection	REGIONAL CLIMATE	Elevation
Landscape	Landtype association		
	Landtype	LOCAL CLIMATE	Topography
Land unit	Landtype phase	SOIL FERTILITY	Soil fertility, soil development, microbial activity

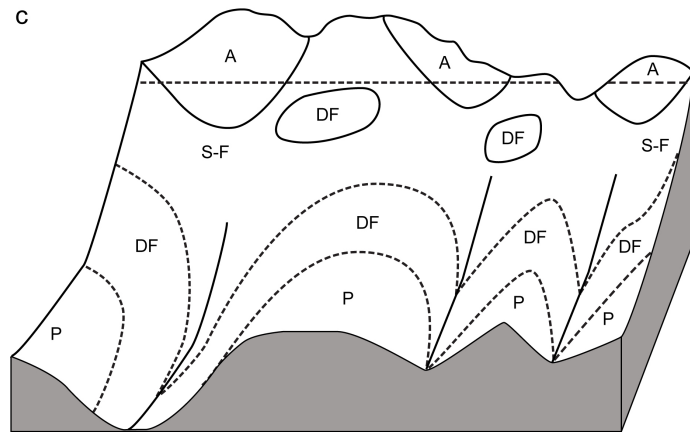
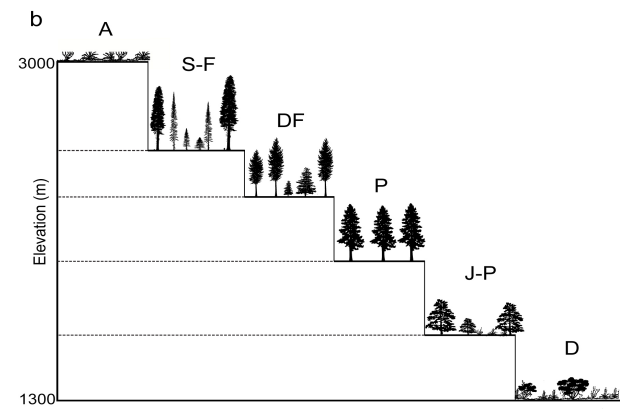
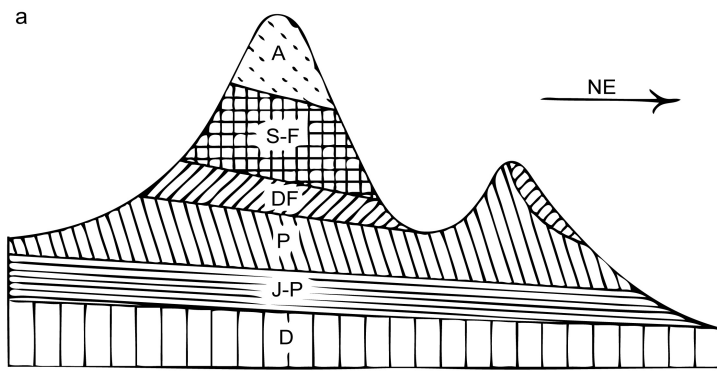


Figure 2.1. Earlier land classifications: (a) life zones as simple elevation belts *sensu* Merriam, (1890) as (b) idealized, stair-like landscape; (c) details of vegetation pattern along complex environmental gradients such as elevation, topography, and soils presented as “telescoping, discontinuity and interfingering” of plant communities leading to inversion of vegetation zones (Daubenmire 1943). Alpine (A), Engelmann spruce-subalpine fir (S-F), Douglas-fir (DF), ponderosa pine (P), juniper-pinyon (J-P), desert (D)

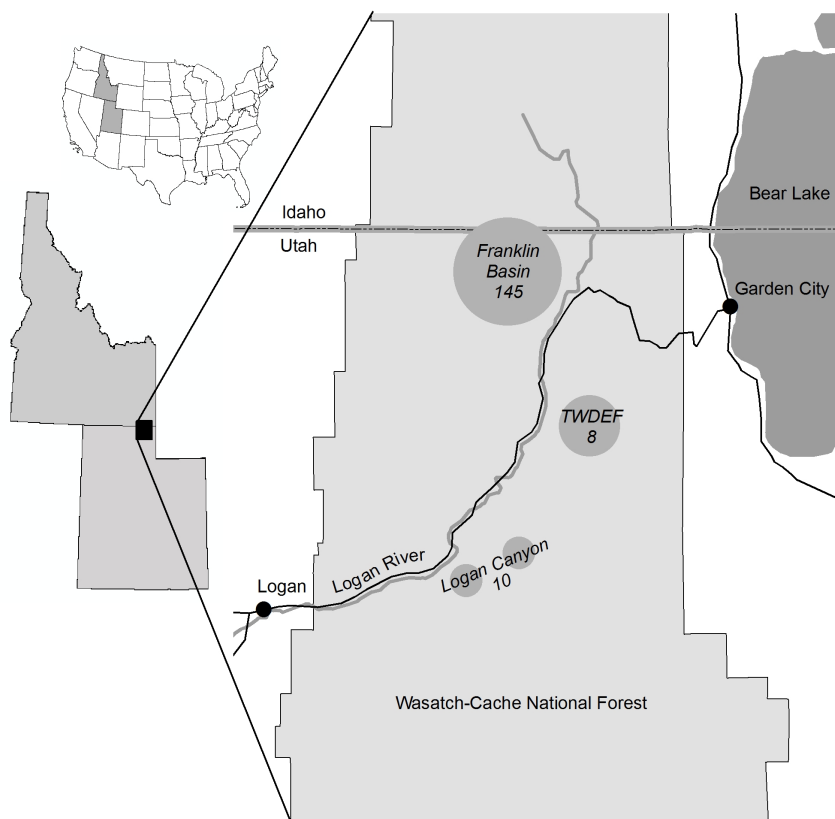


Figure 2.2. The study area on the Utah-Idaho border with subareas; digits represent numbers of the sample plots.

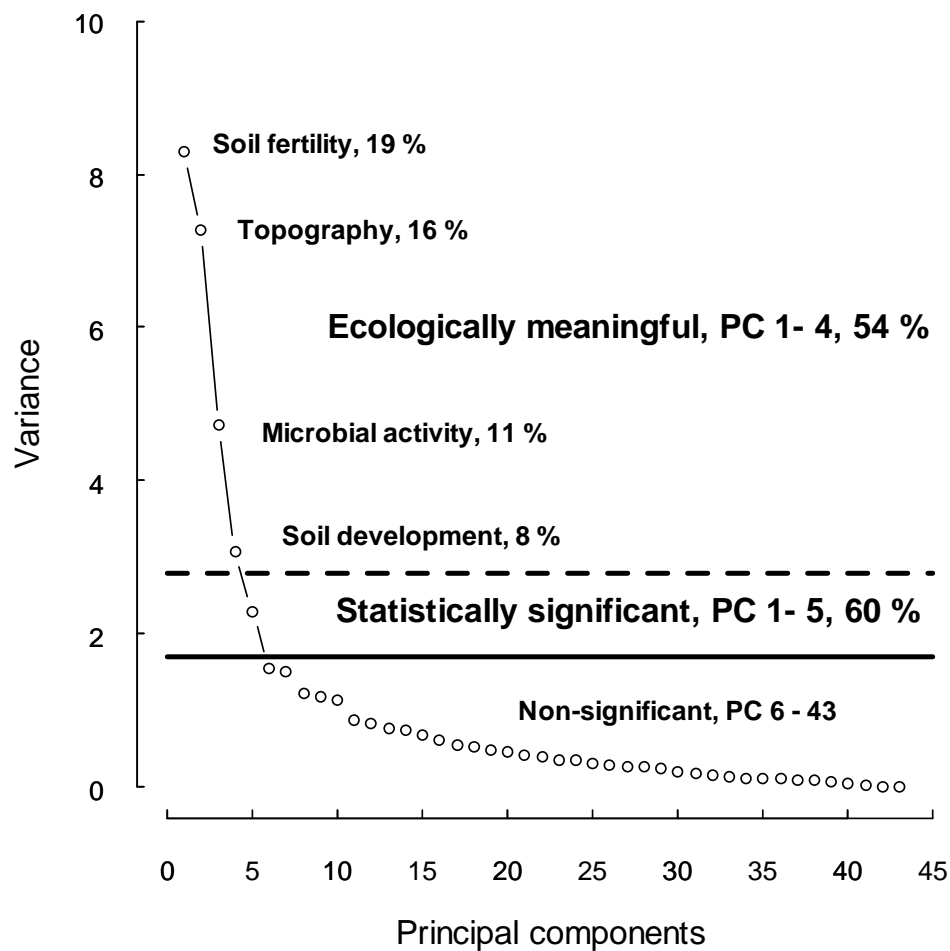


Figure 2.3. PCA scree plot. Ecologically meaningful, statistically significant and non-significant principal components and interpretation of the PCs with proportion of variance (%) explained. See text for explanation of interpretation.

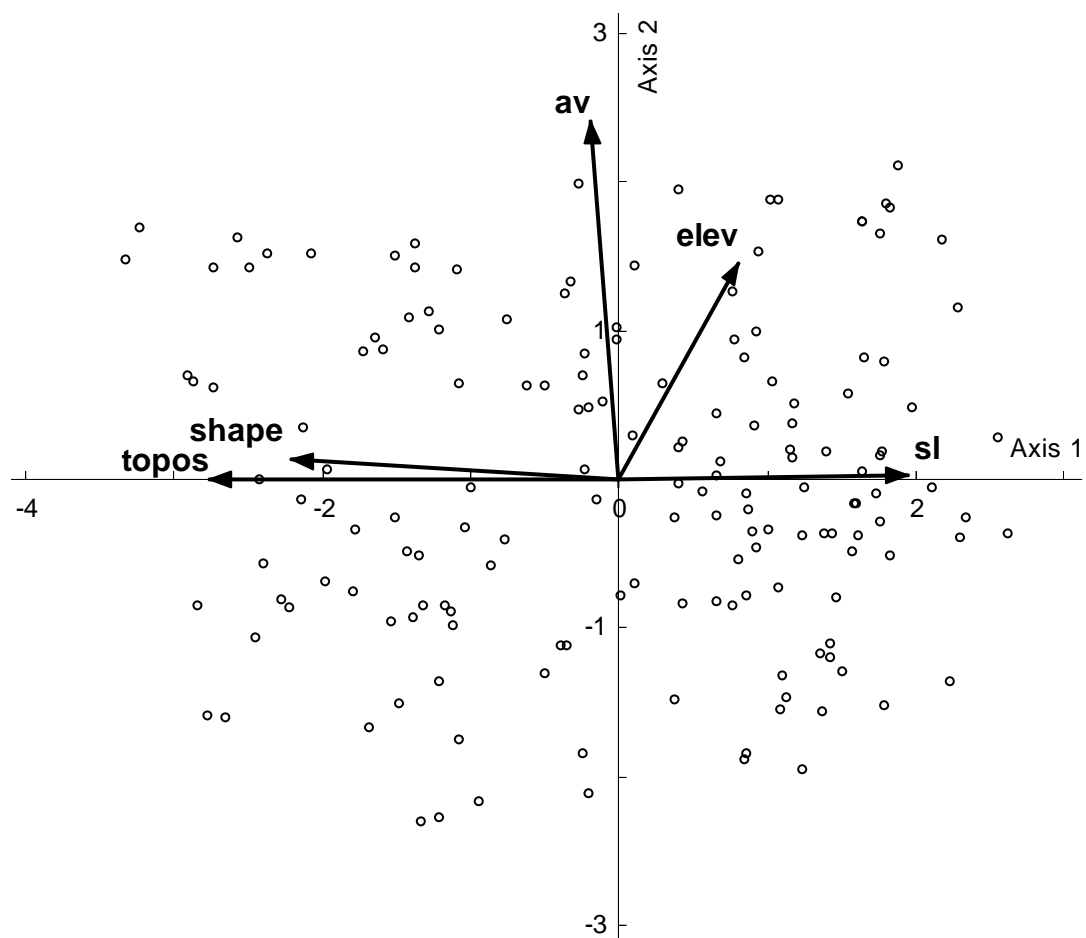


Figure 2.4. Joint plot of the second PCA run. Length (strength) and direction of the vectors show correlation of the physiographic variables in the ordination space. See text for explanation.

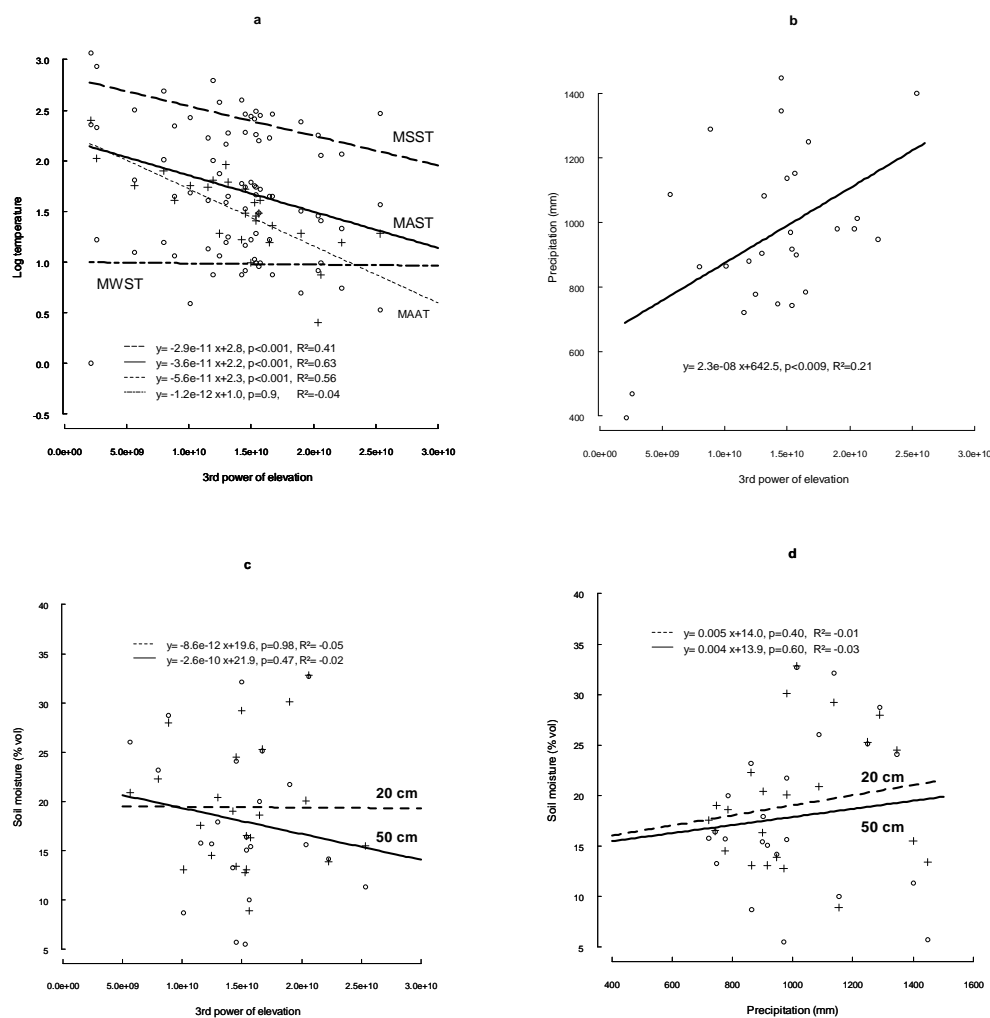


Figure 2.5. Relationship of: (a) air and soil temperatures, (b) precipitation, and (c) soil moisture with elevation; and (d) soil moisture with precipitation for the northern Wasatch Range, UT. Mean annual air temperature (MAAT), mean annual soil temperature (MAST), mean summer soil temperature (MSST), mean winter soil temperature (MWST). Soil moisture in 20 and 50 cm of soil depth.

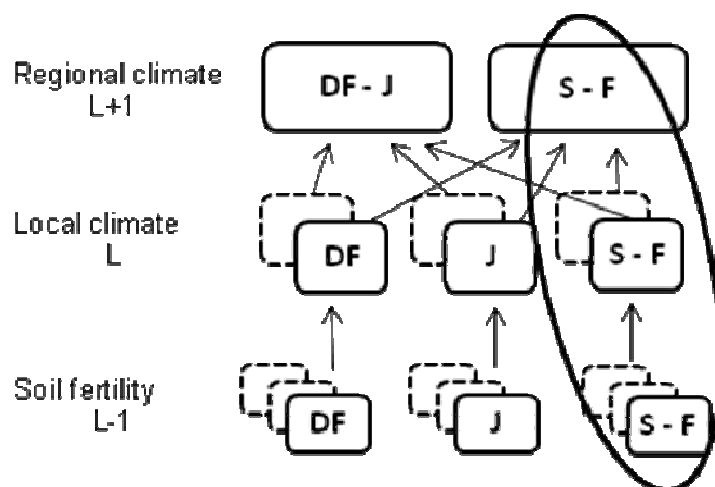


Figure 2.6. Hierarchical organization of forest ecosystems represented by spruce-fir (S-F), Douglas-fir-juniper (DF-J), Douglas-fir (DF), and juniper (J) communities. The oval represents the S-F zone in the earlier classifications. Arrows demonstrate agglomerative (bottom-up) approach; because of environmental compensation, e.g., S-F from the local climate (L) can be found either along valley bottoms inside the DF-J unit or on shaded slopes inside the S-F unit of the regional climate (L+1). The same S-F from the L level consists of either poor or rich S-F from the fertility level (L-1).

CHAPTER 3
VEGETATION GEO-CLIMATIC ZONATION IN THE ROCKY
MOUNTAINS, NORTHERN UTAH ²

Abstract

Fundamentals of the direct gradient analysis, hierarchical organization of terrestrial ecosystems together with approach of the biogeoclimatic classification used in British Columbia were used to develop a vegetation geo-climatic zonation in a mountainous study area in the northern Utah.

This classification was derived from sampling of forest (spruce-fir, Douglas-fir, aspen, juniper woodland) ecosystems on zonal sites, i.e., sites with mature vegetation, moderate topographic and intermediate soil characteristics. Thirty-five plots were described by vegetation, physiographic features and soil properties such as nutrient pools and dynamics. The SNOwpack TELemetry (SNOTEL) and The National Weather Service (NWS) Cooperative Observer Program (COOP) weather station network were used to approximate climate of sample plots.

Principal component analysis revealed the dominant role of regional climate in affecting broad vegetation pattern. We discriminated inside this pattern using cluster analysis, RandomForest classification and ANOVA and based on vegetation physiognomy, climatic data and taxonomic classification of zonal soils we identified two vegetation geo-climatic zones: (1) montane with Rocky Mountain juniper and Douglas-

² Coauthored by Antonin Kusbach, Helga Van Miegroet, James N. Long, and Janis L. Boettinger

fir; and (2) subalpine zone with Engelmann spruce and subalpine fir as climatic climax species.

This zonation will provide the framework for building a comprehensive ecosystem classification.

Introduction

“Without classification there is no science of ecosystems and ecology. And indeed, no science” - V. J. Krajina

Rocky Mountain ecosystems are complex in vegetation. A vegetation pattern i.e., species distribution, is affected by climate, topography, and geology as well as other abiotic and biotic factors such as disturbances and plant interactions. Environmental gradients are steep and the legacies of natural and anthropogenetic disturbances are pervasive (e.g., Gannet 1882, Barnes et al. 1982, McCune and Grace 2002, Shaw and Long 2007). Understanding these diverse ecosystems requires classification that accounts for the underlying complexity in important environmental drivers. Development of such a classification, in turn, requires a meaningful ecosystem organization using e.g., a triadic structure of the definitional levels in which the focal level “L” is between the next lower level “L-1”, and the next higher level “L+1” (e.g., O’Neill et al. 1986, Urban et al. 1987, King 2005).

Such a hierarchical organization was created for the Rocky Mountains, northern Utah represented by the study area (Chapter 2). Regional climate as the highest level (L+1) was superimposed over local climate (the focal level, L) and soil fertility (the lowest level, L-1). We suggest the overarching influence of regional climate in this

organizational hierarchy is strongly related to broad vegetation pattern or regional ecosystems (Klinka and Chourmouzis 1999, K. Klinka 2009, personal communication) such as e.g., life zones (Merriam 1890) or biogeoclimatic zones of mountains (Krajina 1965).

The level (L+1) in the ecosystem organization (regional climate) is suggestive of **the zonal (climatic climax) concept**. Originally formulated in terms of soil zonality (Dokuchaev around 1870), it expresses the relationship between climate, associated vegetation, and soils (e.g., White 1997). In other words, stable, i.e., late-seral or old growth (climatic climax) plant communities with intermediate edaphic conditions (relative to the extremes of a region) best reflect the influence of regional climate (Hills 1952, Krajina 1965, Pojar et al. 1987, Bailey 2002). Thus, local climatic, topographical and edaphic extremes such as those found on warm south-facing slopes, cool north-facing slopes, cold depressions or skeletal soils, are eliminated and only intermediate environmental conditions should be considered in the zonal concept application (Fig. 3.1).

Our overarching goal is to better understand broad vegetation patterns in a Rocky Mountain landscape. Specifically, we examine the relationships between vegetation and environmental variables for **zonal sites** (*sensu* Krajina 1965, Bailey 2002), sites with mature vegetation, moderate topographic and intermediate soil characteristics (i.e., intermediate site conditions) (Pojar et al. 1987, Klinka and Chormousis 1999). Our objective is to assess and classify the response of the complex vegetation to those environmental factors operating at the highest level of our ecosystem organization.

Methods

Study area

Franklin Basin (FB) is a montane-subalpine area, approximately 15,000 ha in size, situated between the Bear River Range and the Wasatch Range in the central Rocky Mountains on the Utah and Idaho border. The T.W. Daniel Experimental Forest (TWDEF), approximately 1,000 ha in size, is situated on the high ridge plateau of the Wasatch Range (10 km to the southeast of FB). Logan Canyon is lower in elevation and its upper part together with FB and TWDEF makes up the study area (ca 20,000 ha, and ca 1,400 m of vertical extent) (Fig. 2.2).

According to Bailey (1998) and McNab et al. (2007), the study area occurs within M331 Southern Rocky Mountains Steppe-Open Woodland-Coniferous Forest-Alpine Meadow Province, “D” Overthrust Mountain Section, “n” Northern Wasatch Range, and “o” Bear River Front Range Subsections. The mean annual precipitation ranges from about 720 to 1250 mm and mean annual air temperature ranges from 2.4 to 5.7 °C for Temple Fork, Tony Grove Lake, Franklin Basin, and Utah State University (USU) Doc Daniel weather stations (<http://www.wcc.nrcs.usda.gov/snow/>).

The terrain is mountainous, rocky and steep with occasional flat to gently sloping high ridge-plateaus and benches. The elevation ranges from 1590 to 3060 m across the three study sites. The highest area of the Bear River Range was glaciated during the Pleistocene as manifested by glacial geomorphologic features like moraines, U-shaped valleys, erratics, and irregular glacial deposits (Young 1939, Degraff 1976). The study area is mostly built from calcareous sedimentary rocks (limestone, dolomite) with interlayered quartzite, and from Tertiary sediments (grit, conglomerate, and siltstone of

Wasatch Formation) at the TWDEF site. The soils are formed in residuum, colluvium, alluvium, glacial till and outwash, and occur on diverse landforms such as cliffs, moraines, karst valleys, slopes, landslides, plains, valleys, depressions, ravines, and wetlands (Schoeneberger et al. 2002).

Over half of the study area is occupied by forest ecosystems including Engelmann spruce (*Picea engelmannii*), subalpine fir (*Abies lasiocarpa*), Douglas-fir (*Pseudotsuga menziesii*), aspen (*Populus tremuloides*), woodland ecosystems including mountain mahogany (*Cercocarpus ledifolius*) and Rocky Mountain juniper (*Juniperus scopulorum*), and riparian, mostly willow (*Salix spp.*) ecosystems. Substantial changes in fire regimes, often in combination with cutting and grazing, have led to dramatic changes in the structure and the age-class distribution of forest stands. In many places, 100- to 140-year-old stands are now predominant (Long 1994). Forests in the study area are thus characterized by mid- and late-seral stages where forest understory is usually well developed (Pfister and Arno 1980).

Data collection

Ecosystems and their components change along an altitudinal gradient (e.g., Gannett 1882, Daubenmire 1943a, b, Whittaker and Niering 1965, Peet 2000, Shaw and Long 2007). We established 35 sample plots in the summers of 2006 and 2007 along a broad elevation range in order to capture broad climatic variation e.g., in temperature and precipitation. It was a part of a larger sampling design, but for this analysis we selected zonal sites i.e., mature forest stands with intermediate site characteristics such as mid-slope position, gentle to moderate slope (< 30 %), loamy soils > 50 cm deep with coarse

rock fragment content < 50 % by volume and no growing-season water table (Pojar et al., 1987) (Fig. 3.1). In other words, slope position, gradient, aspect and shape do not strongly modify overall climate as in e.g., frost pockets, cold air drainages and on steep south/north-facing slopes. As “mature” we considered vegetation with relatively stable stand composition in which potential climax tree species are recognizable, and where a clear successional trajectory is discernible e.g., from advance regeneration of climax species (Pfister and Arno 1980, Pojar et al. 1987).

A stratified (based on vegetation physiognomy) fixed (subjective selection) sampling design was used with circular zonal plots size of 1000 m² (Brohman and Bryant 2005). We described each sample plot by species abundances (cover percentage) and by environmental variables such as relatively static or constant attributes i.e., physiographic variables (elevation, slope aspect, slope gradient, topographic position and slope shape (Lotspeich 1980); dynamic attributes such as O and A horizon thickness, humus form, pH, nutrient pools, attributes describing relatively slow processes; and attributes such as nutrient supply rates describing relatively fast processes (Table 2.1). Parent material observed on the sites was checked by geologic map (Dover 1995). One soil pit was dug in each plot to the unweathered parent material and described using the National Cooperative Soil Survey protocols (Soil Survey Staff 1999, 2006, Schoeneberger et al. 2002). Humus form was identified following Green et al. (1993).

A detailed description of the various site characterization methods was provided in Chapter 2. Briefly, one composite soil sample from 0-30 cm was collected from a pedon face in each plot, air dried and sieved (< 2 mm), and the fine fraction analyzed for texture classes (sandy, loamy, clayey) using the feel-method (Thien 1979). Samples were then

analyzed for pH (1:1 soil in water, Corning pH analyzer) and total C and N (LECO CN analyzer, Leco Corp., St. Joseph, MI). Exchangeable cations using a mechanical vacuum extractor (Holmgren et al. 1977), followed by extractant analysis on inductively-coupled plasma spectrophotometer (ICP) (Iris Advantage, Thermo Electron, Madison, WI); extractable P [the Olsen P method (Olsen et al. 1954), sodium bicarbonate extraction, Thermo Electron Spectronic 20 Genesys spectrophotometer]; and mineralizable N [7-day anaerobic incubation and extraction (Keeney and Bremner 1966), NH_4 analysis (Lachat Quickchem 8000 Flow Injection Analyzer)] were determined as a static-absolute nutrient availability index (SNAI).

To determine a dynamic-relative nutrient availability index (DNAI) (Qian and Schoenau 2002), plant root simulators (PRSTM-probes; Western Ag Innovations, Inc., Saskatoon, Canada), a combination of anion and cation exchange membranes, were buried vertically into the mineral soil at each site for six weeks (during September and November). PRSTM-probes were cleaned and sent to Western Ag Innovations for extraction and chemical analysis including Ca, Mg, K, S, Fe, Mn, Zn, Cu, Pb, Al, NH_4 cations, and NO_3 and PO_4 anions (Table 2.1).

Climatic data such as air temperature, precipitation, soil temperature, and soil moisture for the northern Wasatch Range [corresponding with M331D Section in McNab et al. (2007)] were obtained from nearby weather stations to approximate ambient and soil climate of the zonal sites. Soil temperature (50 cm depth) and moisture (20 cm and 50 cm depth) were measured daily on an open area, i.e., no tree canopy (R. Julander 2009, personal communication). Both the Natural Resources Conservation Service (NRCS) SNOwpack TELemetry (SNOTEL; <http://www.wcc.nrcs.usda.gov/snow/>) and

the National Weather Service Cooperative Observer Program (COOP; <http://www.nws.noaa.gov/om/coop/>) station networks provide long term observations for air temperatures and precipitation (>10 years). Soil temperature and moisture at SNOTEL sites were available for six years (2003-2008, USU Doc Daniel for 2008-2009) and there were two COOP stations with soil temperature measurements for the northern Wasatch Range (Utah Climate Center; <http://climate.usurf.usu.edu/>). Accuracy of the analysis may therefore be limited by the short data record. Also no soil property information is available for the monitoring sites such as organic horizon, texture, and coarse rock fragment content. Nevertheless, these data were an important source of information in this analysis (Table 3.1).

Concepts of vegetation zones

Zones in earlier classifications i.e., life zones (Merriam 1890) or vegetation zones (Daubenmire 1943a, Peet 2000 based on Whittaker and Niering 1965) were named based on general physiognomy, i.e., dominant tree species (Whittaker 1972, Long 1994). For example, within the elevation range of the study area (1500 - 3000 m) there are juniper, Douglas-fir and Engelmann spruce-subalpine fir zones. Within those broad vegetation zones, environmental variation is large.

Aspen-dominated communities cover extensive areas in the Western U.S. (e.g., Rogers et al. 2010) and aspen is considered an extremely important component of many Rocky Mountain landscapes. There is continuous discussion about character of aspen in the Western U.S.; aspen is considered a pioneer, shade-intolerant species that may create either stable (persistent, climax) and unstable (seral) stands (e.g., Mueggler 1985, Kay

1997, Kulakowski et al. 2004, Shepperd et al. 2006, Kashian et al. 2007, Rogers et al. 2010) or even old-growth ancient forests (Peterson et al. 1995). Successional status of aspen communities, especially stable aspen, as well as the environmental conditions within the community is still ill-defined (Mueggler 1988). Typically, because of its successional status (pioneer tree species), aspen is not included within earlier vegetation zonations and has been classified separately (Mueggler 1988). To better understand the role of aspen in these landscapes and its response to environmental factors which might be influencing the distribution of conifer-dominated communities, aspen was included in this analysis despite to its pioneer and because of its putative stable/climax character. We sampled both mature conifer sites (< 15 % of aspen canopy cover) and aspen-dominated sites (> 85 % canopy cover, little or no conifer regeneration) (e.g., Mueggler 1985, Rogers et al. 2010).

Data analysis

In this data analysis, we performed the following analytical steps in order to reveal broad vegetation-site relationships: (1) grouping of vegetation data; (2) ordination of environmental data; (3) cluster analysis on the important environmental variables; (4) discriminant analysis by RandomForests of clusters base on important environmental variables; (5) ANOVA of vegetation groups; and (6) classification of zonal soils based on climatic data. The dataset was comprised of 35 zonal sites, 41 environmental variables and 18 tree and shrub species.

Vegetation grouping was performed as unsupervised cluster analysis of species abundances represented by cover percentage [flexible beta with $\beta = -0.25$, Sorensen

distance (Bray and Curtis 1957)]. Identification of the vegetation groups was based on species constancy/frequency and dominance (e.g., Brohman and Bryant 2005, Winthers et al. 2005, Jennings et al. 2008). Species covers were subjected to square root transformation to approximate a normal distribution.

We used Principal components analysis (PCA) ordination (Pearson 1901) to determine the relative importance of the environmental variables and interpret principal components (PC) associated with zonal sites. Orthogonal rotations and correlation type of a cross-products matrix were used to get independent, mutually uncorrelated PCs (Lattin et al. 2003). Significance of PCs was tested by a Monte Carlo randomization test (based on proportion-based p -values for each PC). In order to document the relationship of the variables with the PCs and interpret PCs, we calculated correlation coefficients (loadings) with each ordination axis, and the linear (parametric Pearson's r) and rank (nonparametric Kendall's τ) relationships between the ordination scores and the observed variables. Our use of r and τ is suggested to be, even in relatively small datasets, more conservative than p -values for the null hypothesis of no relationship between ordination scores and variables (McCune and Grace 2002). We set the threshold for r and $\tau > 0.35$. For variables conversion and transformation see Chapter 2.

To associate the vegetation groups with important environmental factors obtained in the PCA and to distinguish among them, we ran cluster analysis. Ward's linkage method with compatible Euclidian distance matrix has been often used as effective method of multivariate ecological data clustering (Ward 1963, McCune and Grace 2002). However, the best results were accomplished with Sorensen distance (Bray-Curtis coefficient) as suggested by McCune and Grace (2002). We transformed the variables with

$|\text{skewness}| > 1$ to be close to multivariate normality and standardized the data by adjustment to standard deviate (z -scores). We checked the dataset for outliers given cutoff of 2.0 standard deviations from the grand mean (McCune and Mefford 2006). A clustering dendrogram was scaled by a distance objective function (Wishart 1969) and resulting height was used to decide how many clusters to retain (McCune and Grace 2002). The solution was verified by pseudo F function (Calinski and Harabasz 1974).

RandomForests analysis (Breiman 2001) was used to identify the most important environmental variables associated with meaningful zonal site clustering to highlight cluster differences. This machine-learning method is advantageous for classification of ecological data; it is accurate, combines many classification trees, and determines variable importance (e.g., Chen et al. 2004, Cutler et al. 2007).

Using the most important factors obtained from RandomForests classification and PCA, we assessed differences between the vegetation groups by 1-way ANOVA using Student-Newman-Keuls and Tukey's multiple comparison tests. The variables were transformed for normality by power or logarithmic transformation when necessary.

We used the climatic data (Table 3.1) such as air and soil temperature, precipitation, and soil moisture (obtained from SNOTEL and COOP weather stations) as approximation of ambient and soil climate of the zonal sites to classify soils following the National Cooperative Soil Survey (Soil Survey Staff 1999, 2006, Schoeneberger et al 2002). Based on daily soil temperature measurements at a depth of 50 cm from the soil surface, we calculated mean annual soil temperature (MAST), mean summer soil temperature (MSST) (June, July, and August in the Northern Hemisphere), and mean winter soil temperature (MWST) (December, January, February in the Northern

Hemisphere) (Soil Survey Staff 2006). We used daily soil moisture measurements at a depth of 20 and 50 cm (consistent with conditions for the soil moisture control section extent) for calculation of the mean soil moisture, mean number of dry consecutive days in the 4 months following the summer solstice, and mean number of moist consecutive days in the 4 months following the winter solstice (Soil Survey Staff 2006). We considered volumetric soil moisture content of 12 % as a general threshold between a dry and moist soil moisture control section for loamy soils (Brady and Weil 1999).

JUICE software ver. 7. 0. 41. (Tichý 2002), R software, ver. 2. 7. 2. (<http://www.r-project.org/>), SAS 9.1.3 Service Pack 4 software (<http://www.sas.com/software/sas9/>), and PC-ORD 5 (McCune and Mefford 2006) were used in the analysis.

Results

Vegetation grouping

Cluster analysis of species abundances resulted in the identification of five major vegetation groups based on both the most constant and dominant species. These groups are: Engelmann spruce, subalpine fir, Douglas-fir, Rocky Mountain juniper and aspen (Table 3.2).

Environmental ordination

We used the environmental data (Table 2.1) to run two PCAs: the first included all 35 sites (i.e., both conifer- and aspen-dominated sites); the second included only the 18 conifer-dominated sites. The most important principal component (PC1) in both PCA runs was associated with elevation, and soil properties such as A horizon thickness, humus form, rock fragment content, parent material, soil color, pH and nutrients (Table

3.3, 3.4). PC1 was interpreted as a **climate/geomorphology** gradient for both PCA cases, driven by elevation as the surrogate for regional climate (Chapter 2) and by geomorphology, which is reflected by parent material and soil properties (e.g., rock fragment content, color, pH and nutrients) (Fig. 3.2). Climate/geomorphology gradient explained 24 % of total variance in the entire dataset (with aspen) but explanatory power increased to 31 % in the reduced dataset (conifers), suggesting that inclusion of aspen-dominated sites in the analysis masked the importance of this major gradient. When the aspen-dominated sites were excluded, there was a dramatic increase in elevation loadings (Table 3.4). Therefore, the reduced conifer dataset results and twenty-three important environmental variables indicated by significant loadings in PC1 were used further to characterize the climate/geomorphology gradient.

The second principal component (PC2) in the reduced dataset (conifer) was associated with aspect value, organic horizon thickness, soil texture, and nutrients such as N, Ca, Mg, K, Al and P (Table 3.4). We interpreted this PC as indicative of **microbial activity** (Fig. 3.2). This activity influenced soil organic matter decomposition rate influencing organic horizon thickness as well as the nutrient/chemical environment. Warm south-facing slopes experienced enhanced nitrification as indicated by high nitrate DNAs (Table 3.4). Interpretation of the PC2 and also PC3, which is not relevant to this chapter, is consistent with Chapter 2.

Cluster analysis

Twenty-three environmental variables with significant loadings in PC1 were used in cluster analysis to identify environmentally similar sites and their associations to

internally homogeneous and mutually different clusters. In both the entire dataset and conifer analysis there were clearly a two-cluster solution based on the distance objective function and information retained (stability of the two-cluster solution was indicated by the longest horizontal distances of clusters' branches) (Fig. 3.3a, b), and confirmed by calculation of the Pseudo F function (Calinski and Harabasz 1974) with the highest pF value for the two-cluster solution. In the conifer clustering, this solution cleanly discriminated spruce-fir and juniper-Douglas-fir vegetation groups (Fig. 3.3b). In the entire dataset including aspen-dominated sites, clustering did not discriminate these aspen sites (Fig. 3.3a).

Discriminant analysis

The clustering revealed two important, approximately balanced (similar number of observation) clusters/classes (Breiman and Cutler 2005, Chen et al. 2004) that were internally homogeneous and mutually different. RandomForests classification identified those environmental variables most strongly associated with this two-cluster solution. The clusters/classes represented different environments based on twenty-three important environmental variables. As a RandomForests result, there was a distinct break in the ranked variables discriminating more important from less important variables. Four environmental variables were identified as the most important factors discriminating the clustering of sites. In order of apparent importance, they were: Ca, K, elevation and Mn (Fig. 3.4). Except for elevation, all referred to soil nutrient status. "Out-of-bag" estimate of error rate as a measure of RandomForests misclassification was 6 %.

ANOVA

1-way ANOVA revealed overall significant differences between the vegetation groups represented by the conifer-dominated sites in the important variables identified by RandomForests using F -test and verified by p values (Ca SNAI: $F = 4.57$, $p = 0.0053$; K SNAI: $F = 7.02$, $p = 0.0006$; elevation: $F = 15.32$, $p < 0.0001$; Mn DNAI: $F = 6.53$, $p = 0.0007$). Based on Student-Newman-Keuls and Tukey's multiple comparison tests, the juniper and Douglas-fir groups were not significantly different in terms of the important variables; neither were the subalpine fir and Engelmann spruce groups. The aspen group was not significantly different from the spruce and fir group, but was different from the juniper and Douglas-fir group (no aspen-dominated sites sampled in the lowest elevations) (Fig. 3.5, Table 3.5).

Therefore, we combined the four conifer-dominated groups into two final physiognomic groups: the composite juniper/Douglas-fir group we referred to as **montane**; and the composite Engelmann spruce/subalpine fir group we referred to as **subalpine**. The two resulting physiognomic groups differed from each other in all important environmental variables (Ca SNAI: $F = 7.39$, $p = 0.0023$; K SNAI: $F = 9.99$, $p = 0.0005$; elevation: $F = 24.50$, $p < 0.0001$; Mn DNAI: $F = 9.19$, $p = 0.0007$). Student-Newman-Keuls and Tukey's pairwise tests showed that the four important variables behaved consistently across the physiognomic groups; these are significantly different.

Based on clustering, classification and ANOVA the subalpine and montane physiognomic groups differed from each other based on climate/geomorphology. Aspen-dominated sites did not differ from the subalpine physiognomic group but differed from the montane physiognomic group (Fig. 3.5, Table 3.5).

Microbial activity (PC2) was associated with pronounced nitrification on gentle warm (i.e., south-facing) slopes. We focused on significant indicators of N availability i.e., mineralizable nitrogen and nitrate DNAI (Table 3.4). Once again, the vegetation groups were associated with distinctly different microbial activity (PC2) based on slope aspect, nitrogen DNAIs; (slope aspect: $F = 9.26$, $p = 0.0002$; Nmin DNAI: $F = 12.61$, $p < 0.0001$; NO_3 DNAI: $F = 14.36$, $p < 0.0001$). Student-Newman-Keuls pairwise test showed significant difference between the Douglas-fir and juniper group in slope aspect, mineralizable nitrogen and nitrate DNAI ($\alpha = 0.05$). Tukey's pairwise test showed significant difference between the Douglas-fir and juniper groups in nitrate DNAI (Table 3.5). These results implied that different aspects even on mild slopes may contribute to differences in nitrification resulting differences in the Douglas-fir and juniper group occurrence.

Classification of zonal soils

In the mountains of northern Utah, elevation is a good predictor of air temperature, soil temperature and also precipitation (Chapter 2). Despite the significant change of mean soil temperature with elevation, (except mean winter soil temperature - MWST due to snowpack insulation, Van Miegroet et al. 2000), regression results of the SNOTEL and COOP climatic data (Chapter 2) cannot be used to estimate temperature regime of zonal soils because of: (1) absence of climate station data from lower elevation forested soils; and (2) absence of tree cover type and O horizon information at the monitoring sites. Tree cover may have an influence on soil temperature regime as it transforms air temperature fundamentally. We estimated soil temperature regime based on Munk's (1988)

measurements of soil temperatures under different tree canopy types in Logan Canyon, northern Utah. The soil temperature regime of soils under spruce-fir and aspen was classified as **cryic**, and under Douglas-fir and Rocky Mountain juniper as **frigid** (Soil Survey Staff 2006).

Soil moisture was not significantly related to either elevation or precipitation (Chapter 2); soil physical properties such as texture, depth, and coarse fragment content may be superimposed over the effect of overall climate represented by elevation. Because these soil physical properties were not available for the climate stations and calculation of the number of dry and moist consecutive days was inconclusive (Table 3.1), we used earlier measurements of soil moistures under different tree canopy types again supported by the data of nearby weather stations to estimate soil moisture regime. The soil moisture regime of soils under spruce-fir and aspen was classified as **udic**, and under Douglas-fir and Rocky Mountain juniper as **xeric** (Munk 1988, Soil Survey Staff 2006).

Soils in the montane group (juniper and Douglas-fir communities) were classified as Pachic and Typic Argixerolls in the soil order of **Mollisols**. The majority of soils in the subalpine group (subalpine fir and Engelmann spruce communities) were classified as Typic and Eutric Haplocryalfs in the soil order of **Alfisols** (Appendix A) (Soil Survey Staff 2006). Soil classifications of both physiognomic groups was consistent with e.g., Burns and Honkala (1990) and Erikson and Mortensen (1974). The majority of soils under aspen community were classified as Pachic and Typic Palecryolls, Typic Argicryolls, and Typic Haplocryolls (**Mollisols**) (Appendix A, Table 3.6). As in the montane zone, Mollisols indicate thick A horizon, soil organic matter accumulation and huge potential for C storage.

The parent material of subalpine zonal soils was mostly Pleistocene glacial deposits (moraine till) and Eocene sediments (grit, conglomerate, siltstone). Montane zonal soils have been formed in late Pleistocene - Holocene fluvio-colluvial deposits derived from Ordovician and Cambrian calcareous sediments.

Significant difference between the physiognomic groups in elevation (Fig. 3.5) together with strong relationships between elevation and climate suggested important climatic difference between these groups. This climatic difference combined with differences in geomorphology contributes to substantial soil differences among the physiognomic groups and distinguishes zonal soils at the level of soil order (Table 3.6).

Discussion

Our analysis, based on a broad range of data (vegetation, climate, and geomorphology) revealed a strong altitudinal pattern within the highest level (L+1) of our hierarchical ecosystem organization. At this level of organization, we distinguished two firm vegetation geo-climatic zones: the montane or lower mountain and the subalpine or upper mountain zone. These zones occur as stacked broad vertical belts with dramatic climatic, vegetation and geomorphologic differences.

The montane zone is characterized by Rocky Mountain juniper and Douglas-fir as the potential climatic climax species. The zone is warmer and drier than the subalpine zone; these climatic properties together with different parent material and rich understory vegetation are reflected by fertile Mollisols, which are also younger than the subalpine zonal soils. Higher potential productivity of the montane Mollisols is indicated not just by thick A horizons, but also by significantly higher concentrations of important

macronutrients Ca and K contributing to higher soil alkalinity (Table 3.6). Because there was no significant difference between Douglas-fir and juniper communities in important environmental factors, and because there were relatively few zonal Douglas-fir-dominated stands ($> ca\ 1000\ m^2$), we did not separate a Douglas-fir vegetation geoclimatic zone in the northern Wasatch Range.

Rocky Mountain juniper and Douglas-fir can form mixed-species stands on zonal sites. Floristic differences between Douglas-fir and juniper-dominated sites within the montane zone may result from: (1) different shade tolerance of these two species; and potentially (2) differences in N availability and form (either ammonium mineralization or nitrification). For example, increased insolation on south-facing slopes may limit Douglas-fir reproduction via failed seed germination or low seedling survival (Bates 1923, Burns and Honkala 1990). The shade-tolerance and requirement of a species may be increased on warm dry sites (Krajina 1965, 1969, Klinka and Chormousis 1999). Whereas Douglas-fir is intermediate in shade-tolerance compared to many of its associates (Burns and Honkala 1990), it may require more shade protection for establishment in semiarid conditions (Krajina 1965, 1969). Juniper, on the other hand, is a very shade-intolerant species (i.e., requires light exposure particularly light from above) in the later life stages (e.g., Burns and Honkala 1990). If Douglas-fir is able to establish, it has the potential to overtop and outcompete juniper (e.g., in thicker secondary growth).

Generally, conifers such as spruces (*Picea*), firs (*Abies*) and pines (*Pinus*) are physiologically adapted to high ammonium levels in soils; they take up ammonium preferentially (Yanai et al. 2009, Hangs et al. 2003, Bedell et al. 1999, Olsthoorn et al. 1991). Similarly, for Douglas-fir in dry conditions of south-facing slopes, uptake of

ammonium is limited by its relative immobility (Gijsman 1991), but it is still preferred to nitrate (Kamminga-van Wijk and Prins 1993). Junipers appear to prefer nitrate to ammonium (Miller et al. 1991, Stark and Hart 1997) and may therefore be more competitive in high nitrification environments. Also allelopathic properties of Rocky Mountain juniper can inhibit establishment of other plants including Douglas-fir (e.g., Peterson 1972, Horman and Anderson 2003). Our ANOVA PC2 results are consistent with pronounced nitrification, and possibly low availability of ammonium, restricting Douglas-fir regeneration on south-facing slopes.

Engelmann spruce and subalpine fir are the climatic climax species for **the subalpine zone** at higher elevation. The zone is cooler and moister than the montane zone. A major portion of the zone in the study area has a glacial history (Young 1939, Degraff 1976) and soils have experienced frequent climatic changes during the Pleistocene (Buol et al. 2003). Lower productivity and higher acidity of the subalpine Alfisols is indicated by significantly lower Ca and K concentrations and higher Mn supply rate (Table 3.6). The higher amount of this metal is associated with more humid subalpine conditions likely facilitating release of this metal from parent material.

Engelmann spruce and subalpine fir are both very shade-tolerant tree species. They are commonly found in mixtures with spruce dominance in old growth and late-seral stands (e.g., Aplet et al. 1988, Peet 2000). In lower, more accessible parts of the subalpine zone within the study area, subalpine fir tends to be more abundant than spruce, probably the result of pioneer logging which favored removal of spruce. As a result of the logging history, we suspect Engelmann spruce is still somewhat underrepresented in second growth mid-seral stands.

The ecological amplitude of aspen is extremely broad in comparison with the conifers. This amplitude is climatic, as represented by aspen's large elevation range, and geomorphologic, as indicated by its occurrence on diverse soil parent materials (Fig. 3.5, Table 3.6). The wide range of climate and geomorphology is associated with large differences in nutrient availability among soils in aspen-dominated sites. Aspen occurs on rich sites with surpluses of macronutrients such as N, K, Ca, Mg. It also occurs on relatively poor sites where some secondary macronutrients may be deficient (Ca, Mg) and micronutrients such as Mn are in surplus. There is no single environmental factor, important at the level of regional climate, that can discriminate aspen as a discrete vegetation geo-climatic zone; this confirms the exceptionally broad ecological amplitude (e.g., Mueggler 1988, Klinka et al. 1999) and it is consistent with high genetic variability of trembling aspen (Mock et al. 2008).

There is no alpine vegetation geo-climatic zone in the study area and probably in the northern Wasatch Range because of the absence of alpine zonal sites in the highest elevations (over 3000 m). The northern Wasatch Range is not high enough for the alpine zone.

True zonal sites are rare in the central Rocky Mountains, Utah because of: (1) a rough, mountainous landscape (Barnes et al. 1982) and "accidents of topography" (Gannet 1882), meaning that "abruptness of elevation change over the local landscape can appear more important than absolute elevation" (Shaw and Long 2007); and (2) many ecosystems never reach potential climax due to natural disturbances such as fire (e.g., Pojar et al. 1987, Cook 1996) and human-caused disturbances such as past logging. Since zonal sites are scarce, a compromise was necessary in site selection during this study.

Rather than restricting sampling exclusively to old-growth, we included sites with mature, mid- to late-seral stands with well developed understory reflecting an obvious successional trajectory (Pfister and Arno 1980, Pojar et al. 1987).

Our vegetation geo-climatic zonation is explicitly framed by the highest definitional level (L+1), i.e., regional climate and geomorphology, in a hierarchical ecosystem organization. There is general consistency in our approach with the earlier vegetation zonation of Merriam (1890) based on an idea of broad life zones depending on overall climate (Chapter 2). However, there is a substantial difference between our approach and the vegetation zonation of Daubenmire (1943a), in that his approach is entirely based on vegetation without specific environmental information.

There may be a compensating influence of environmental factors on plants, and because of that compensation the same climax vegetation may appear over a broad environmental range (Pojar et al. 1987). For example, Engelmann spruce-subalpine fir (S-F) communities occur in the high-elevation subalpine vegetation geo-climatic zone; however they may also descend into the lower montane vegetation geo-climatic zone on shady slopes or as riparian/wetland communities along valley bottoms. In the subalpine zone, precipitation is higher, affecting S-F occurrence in intermediate conditions (on zonal sites). In the lower montane zone with lower precipitation, this decrease of water input may be compensated by supply and retention of water in locally specific conditions (local topography) such as shady slopes and valley bottoms. Local environmental conditions thus modify the influence of the low-elevation regional climate in the montane vegetation geo-climatic zone (Chapter 2). As a result, S-F communities can occur on lower elevation sites where there is enough moisture available.

Whereas Daubenmire (1943a) described spruce-fir, Douglas-fir, juniper and mountain mahogany communities as representatives of four different zones dependent on “ecologic criteria” represented by climate/elevation, we consider these communities to represent local topo-edaphic variations within a single vegetation geo-climatic zone. Such large environmental variations (accompanied by environmental compensation) are thus reflected by floristic differences within a vegetation geo-climatic zone. By specifying floristic differences between the vegetation geo-climatic zones, we can explain many special cases such as interfingering, telescoping, discontinuity and inversion of earlier vegetation zones (Daubenmire 1943a) that reflect the lower (local) level L (i.e., local topography and soil moisture, Chapter 2). It is apparent that the Daubenmire zonation combined the regional L+1 with the local L (topography-moisture) levels. Peet’s (2000) improved zonation by explicitly differentiating two factors, represented by elevation and topography-moisture, but keeping earlier vegetation zones *sensu* Daubenmire inside that framework.

There is great potential value of vegetation geo-climatic zonation; by keeping the L+1 level separate we contributed to “understanding of the discontinuity of the historical and environmentally broad vegetation zones” (Shaw and Long 2007). We argue that firm altitudinal belts exist in the central Rocky Mountains (Daubenmire 1943a). The vegetation geo-climatic zonation distinguished at the L+1 level will serve as the framework for further detailed ecosystem structuring at the lower levels.

There is consistency between Hierarchical Framework of Ecological Units (Cleland et al. 1997, ECOMAP 2007), the Terrestrial Ecological Unit Inventory (TEUI) Technical Guide (Winthers et al. 2005), and vegetation geo-climatic zonation; as regional climatic

units the zones match sections, subsections and landtype associations units of subregions and landscapes in ECOMAP/TEUI standard (Table 2.4. in Chapter 2).

Summary and conclusions

Based on the zonal concept and the ecosystem organizational hierarchy, we defined two vegetation geo-climatic zones as areas with the same floristic structure in climatic climax. These zones were: **montane** with juniper/Douglas-fir; and **subalpine** with Engelmann spruce/subalpine fir as climatic climax species. We characterized these zones based on regional physical environment (i.e., climate and landform geomorphology/soils); with regional climate represented by elevation, precipitation and air and soil temperatures; and geomorphology by soil types. Aspen was excluded from the zonation due its great ecological amplitude. Even stable or climax aspen was not considered as zonal (climatic climax) vegetation. We suggest both the montane and subalpine zones match the subregion and landscape scale of the ECOMAP/TEUI classification.

The vegetation geo-climatic zonation outlined in this paper is a conceptual improvement on earlier approaches to vegetation zonation in the region. As a next step, the vegetation geo-climatic zonation can be used as a framework for additional structuring e.g., site classification, and in building a comprehensive ecosystem classification.

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Table 3.1. Climatic data for the northern Wasatch Range. SNOwpack TELemetry weather stations (SNOTEL), the National Weather Service Cooperative Observer Program weather stations (COOP), mean annual precipitation (MAP), mean annual air temperature (MAAT), mean annual soil temperature at 50 cm (MAST), mean summer soil temperature at 50 cm (MSST), mean winter soil temperature at 50 cm soil depth (MWST), mean soil moisture at 20 cm (Moist20), and 50 cm soil depth (Moist50), mean number of dry consecutive days at 20 cm (Dcon20), and 50 cm soil depth (Dcon50), mean number of moist consecutive days at 20 cm (Mcon20) and 50 cm soil depth (Mcon50). The stations in the study area and closest vicinity are in bold. NA-not applicable.

Station ID	Station	Elevation (m)	MAP (mm)	MAAT (C)	MAST (C)	MSST (C)	MWST (C)	Moist20 (% vol)	Moist50 (% vol)	Dcon20	Dcon50	Mcon20	Mcon50
COOP427 598	SLC INT AP	1286	394	11.00	10.60	21.50	0.00	NA	NA	NA	NA	NA	NA
COOP425 194	LOGAN EXP. FARM	1369	469	7.60	10.30	18.80	2.40	NA	NA	NA	NA	NA	NA
SNOTEL333	BEN LOMOND TR.	1776	1087	5.80	6.10	12.20	2.00	20.90	26.00	48	0	123	123
SNOTEL582	LITTLE BEAR	1995	862	6.70	7.50	14.70	2.30	22.30	23.20	5	0	123	123
SNOTEL1054	FARMINGTON L.	2066	1289	5.00	5.20	10.40	1.90	28.00	28.70	7	11	123	123
SNOTEL906	DRY FORK	2162	865	5.80	5.40	11.30	0.80	13.10	8.70	92	114	121	85
SNOTEL1013	TEMPLE FORK	2257	721	5.70	5.00	9.30	2.10	17.60	15.80	30	98	109	88
SNOTEL684	PARLEY'S SUM.	2286	879	6.10	7.40	16.40	1.40	NA	NA	NA	NA	NA	NA
SNOTEL763	SMITH & MOREH.	2316	777	3.60	6.50	13.20	1.90	14.50	15.70	50	97	123	106
SNOTEL1039	CASCADE MNT.	2348	903	7.10	4.90	8.70	2.30	20.40	17.90	78	87	122	109
SNOTEL971	PARRISH CREEK	2359	1082	6.00	5.20	9.70	2.50	NA	NA	NA	NA	NA	NA
SNOTEL374	BUG LAKE	2423	748	3.40	5.90	13.50	1.40	19.00	13.30	35	52	107	66

SNOTEL332	BEN LOMOND P.	2438	1448	5.60	5.70	11.70	2.20	13.40	5.70	79	121	122	4
SNOTEL474	FARMINGTON	2438	1347	4.40	4.60	9.80	1.50	24.50	24.10	51	90	122	122
SNOTEL484	FRANKLIN BASIN	2464	1137	2.70	6.00	11.50	2.40	29.20	32.10	1	0	123	123
SNOTEL820	TIMPANOGOS DIV.	2481	970	4.90	5.80	11.20	1.80	12.80	5.50	69	123	122	9
SNOTEL533	HORSE RIDGE	2487	917	4.30	5.70	12.10	1.70	13.10	15.10	103	90	123	115
SNOTEL393	CHALK CREEK #2	2487	742	4.10	5.30	9.60	2.60	16.50	16.40	58	101	107	107
SNOTEL596	LOOKOUT PEAK	2500	1153	4.40	4.40	9.00	1.60	8.90	10.00	97	90	45	95
SNOTEL1056	LIGHTNING RIDGE	2504	899	5.00	5.60	11.60	1.70	16.30	15.40	NA	NA	NA	NA
SNOTEL1098	USU DOC DANIEL	2521	1055	2.40	NA	NA	NA	22.70	16.50	8	39	122	18
SNOTEL455	DRY BREAD POND	2545	785	3.30	5.20	9.30	2.40	18.60	20.00	52	92	123	123
SNOTEL823	TONY GR. LAKE	2556	1250	3.90	5.20	11.70	1.40	25.30	25.10	38	0	122	0
SNOTEL366	BRIGHTON	2667	980	3.60	4.50	10.90	1.00	30.10	21.70	6	48	122	122
SNOTEL634	MONTE CRISTO	2731	981	1.50	4.30	9.50	1.50	20.10	15.60	53	73	123	105
SNOTEL392	CHALK CREEK #1	2741	1013	2.40	4.10	7.80	1.70	32.80	32.70	0	0	122	122
SNOTEL814	THAYNES CYN	2813	947	3.30	3.80	7.90	1.10	13.90	14.20	74	65	56	71
SNOTEL766	SNOWBIRD	2938	1400	3.60	4.80	11.80	0.70	15.50	11.30	62	83	109	15

Table 3.2. Vegetation analysis of tree communities. Bold tree species indicate both the most constant and dominant species and are indicators for the particular tree community. Constant species percentage means the proportion of the plots in which species occurs; all species with the threshold value $\geq 60\%$ are displayed. E.g., *Picea engelmannii* 100 means that this species occurs on 100 % of the sample plots within the vegetation group 1. Dominant species percentage means the proportion of the plots, in which species occurs with cover \geq the threshold value for dominant species; all species with the threshold value $\geq 40\%$; 20 % for thin juniper woodland are displayed. E.g., *Picea engelmannii* 50 means that this species occurs on 50 % of the sample plots with cover $\geq 40\%$ within the vegetation group 1.

Vegetation group 1

Number of plots: 8

Threshold value for constant species: 60

Threshold value for dominant species: 40

Constant species: ***Picea engelmannii* 100**, *Abies lasiocarpa* 100, *Populus tremuloides* 75

Dominant species: ***Picea engelmannii* 50**

Vegetation group 2

Number of plots: 3

Threshold value for constant species: 60

Threshold value for dominant species: 40

Constant species: ***Abies lasiocarpa* 100**, *Picea engelmannii* 100, *Populus tremuloides* 100, *Symphoricarpos oreophilus* 67, *Sambucus cerulea* 67, *Pseudotsuga menziesii* 67, *Amelanchier alnifolia* 67

Dominant species: ***Abies lasiocarpa* 33**

Vegetation group 3

Number of plots: 17

Threshold value for constant species: 60

Threshold value for dominant species: 40

Constant species: *Populus tremuloides* 100, *Symphoricarpos oreophilus* 94, *Abies lasiocarpa* 71

Dominant species: *Populus tremuloides* 94, *Symphoricarpos oreophilus* 6

Vegetation group 4

Number of plots: 4

Threshold value for constant species: 60

Threshold value for dominant species: 20

Constant species: *Juniperus scopulorum* 100, *Pseudotsuga menziesii* 100, *Symphoricarpos oreophilus* 75, *Prunus virginiana* 75

Dominant species: *Juniperus scopulorum* 75

Vegetation group 5

Number of plots: 3

Threshold value for constant species: 60

Threshold value for dominant species: 40

Constant species: *Pseudotsuga menziesii* 100, *Juniperus scopulorum* 100, *Amelanchies alnifolia* 100, *Symphoricarpos oreophilus* 100, *Prunus virginiana* 67, *Abies lasiocarpa* 67

Dominant species: *Pseudotsuga menziesii* 100

Table 3.3. PCA summary. Significant principal components are in bold.

Entire dataset (with aspen)	PC1	PC2	PC3	PC4	PC5	PC6	PC7	PC8	PC9	PC10
Eigenvalue	10.03	6.49	4.43	3.85	2.19	1.73	1.56	1.47	1.18	1.03
% of Variance	24.46	15.82	10.80	9.39	5.35	4.21	3.80	3.57	2.88	2.51
Cumulative % of Var.	24.46	40.28	51.08	60.47	65.81	70.03	73.82	77.39	80.27	82.78
<i>p</i> - value	0.0002	0.0002	0.0002	0.0002	1.0000	1.0000	1.0000	1.0000	1.0000	1.0000
Conifer dataset										
Eigenvalue	12.75	6.29	4.65	3.65	2.69	2.12	1.74	1.29	1.22	1.16
% of Variance	31.09	15.34	11.34	8.90	6.65	5.17	4.24	3.14	2.98	2.83
Cumulative % of Var.	31.09	46.43	57.77	66.68	73.24	78.41	82.65	85.79	88.77	91.59
<i>p</i> - value	0.0002	0.0002	0.0439	0.7752	1.0000	1.0000	1.0000	1.0000	1.0000	1.0000

Table 3.4. PCA loadings. Significant Pearson's (r), and Kendall's (τ) coefficients are in bold; both significant r and τ express a significant variable for the particular PC (shaded). Variables are defined in Table 2.1.

Variable	Entire dataset				Conifer dataset			
	PC1		PC2		PC1		PC2	
	r	τ	r	τ	r	τ	r	τ
elev	-0.52	-0.37	-0.55	-0.31	-0.84	-0.54	0.03	-0.09
topos	0.43	0.38	0.39	0.30	0.34	0.30	-0.24	-0.19
sl	0.12	0.13	-0.29	-0.15	0.25	0.25	0.36	0.29
av	-0.49	-0.31	0.48	0.28	-0.36	-0.13	0.53	0.46
shape	0.28	0.25	0.15	0.11	0.07	0.02	0.08	0.07
Ohor	-0.67	-0.46	0.50	0.32	-0.48	-0.19	0.72	0.64
Ahor	0.76	0.60	-0.21	-0.19	0.62	0.40	-0.15	-0.05
hum	0.73	0.49	-0.43	-0.20	0.65	0.50	-0.50	-0.30
sdepth	0.25	0.18	0.50	0.27	0.24	0.14	0.01	-0.06
RF	-0.55	-0.40	-0.43	-0.28	-0.66	-0.56	-0.19	-0.17
parmat	0.68	0.54	0.20	0.04	0.56	0.45	0.12	0.12
mottles	0.33	0.22	-0.04	-0.07	0.60	0.49	0.14	0.11
cvalue	-0.73	-0.57	-0.05	-0.02	-0.67	-0.58	-0.14	-0.02
text	0.18	0.12	0.42	0.37	0.52	0.43	-0.56	-0.43
pH	0.56	0.37	0.42	0.23	0.69	0.48	0.23	0.07
Nmin_d	0.31	0.24	-0.84	-0.61	-0.11	-0.11	-0.66	-0.46
Nox	0.85	0.64	-0.08	-0.10	0.68	0.48	0.30	0.22
NO3_d	0.38	0.30	-0.82	-0.60	0.08	0.04	-0.61	-0.44
NH4_d	-0.43	-0.34	0.07	0.09	-0.64	-0.44	-0.31	-0.19
Cox	0.69	0.55	-0.01	-0.13	0.62	0.41	0.38	0.24

C/N	-0.62	-0.38	0.15	0.12	-0.62	-0.39	-0.11	-0.03
Ca_d	0.38	0.26	-0.35	-0.26	0.13	0.15	-0.72	-0.46
Mg_d	0.40	0.32	-0.23	-0.16	0.07	0.02	-0.74	-0.49
K_d	-0.16	-0.10	-0.29	-0.24	-0.24	-0.16	0.63	0.50
P_d	0.25	0.19	-0.28	-0.24	0.03	-0.01	-0.16	-0.21
Fe_d	-0.01	0.01	-0.76	-0.60	-0.57	-0.36	-0.19	-0.08
Mn_d	-0.60	-0.37	-0.62	-0.41	-0.89	-0.71	-0.17	-0.20
Zn_d	0.34	0.22	-0.69	-0.55	-0.08	-0.16	-0.19	-0.14
S_d	-0.30	-0.19	-0.52	-0.33	-0.67	-0.56	-0.03	-0.10
Al_d	0.00	0.06	-0.22	-0.15	-0.15	-0.14	-0.87	-0.80
Ca_s	0.82	0.59	0.29	0.21	0.95	0.79	0.10	0.07
Mg_s	0.73	0.59	0.32	0.23	0.77	0.58	-0.33	-0.16
K_s	0.35	0.26	0.41	0.26	0.89	0.67	0.20	0.14
NH4_s	0.21	0.18	-0.45	-0.34	-0.36	-0.28	0.16	0.05
Nmin_s	0.72	0.60	-0.15	-0.12	0.73	0.62	0.21	0.22
P_s	0.36	0.19	-0.06	-0.06	0.28	0.20	0.61	0.45
Al_s	-0.64	-0.47	-0.11	-0.08	-0.75	-0.60	-0.07	-0.08
Fe_s	-0.44	-0.37	0.11	0.04	-0.72	-0.54	0.31	0.26
S_s	0.33	0.20	-0.14	-0.13	-0.23	-0.23	0.59	0.33
Mn_s	-0.05	-0.02	-0.50	-0.30	-0.55	-0.32	0.24	0.23
Zn_s	-0.44	-0.30	-0.19	-0.16	-0.71	-0.58	-0.05	-0.06

Table 3.5. Identification of vegetation groups. Different capital letters following variable values indicate significant differences between vegetation groups ($\alpha = 0.05$). Variables are defined in Table 2.1.

Vegetation group	Elevation	Ca_s	K_s	Mn_d
	Mean (95 % confidence limits)			
	(m)	(mg/kg soil)	(mg/kg soil)	($\mu\text{g}/10 \text{ cm}^2/6\text{weeks}$)
Juniper	1784B (1520, 2013)	3937A (2629, 5509)	41A (36, 48)	0.7BC (0.1, 5)
Douglas-fir	1990B (1719, 2230)	4204A (2660, 6098)	44A (37, 51)	0.2C (0, 1.5)
Subalpine fir	2396A (2207, 2571)	2627B (1582, 3935)	35B (29, 40)	4AB (0, 30)
Engelmann spruce	2625A (2497, 2748)	1528B (925, 2282)	27B (24, 31)	55A (13, 235)
Aspen	2427A (2338, 2512)	3144B (2550, 3799)	35B (32, 38)	8AB (3, 21)

Table 3.6. Identification of vegetation geo-climatic zones. Mean annual precipitation (MAP), mean annual air temperature (MAAT), mean annual soil temperature (MAST), mean summer soil temperature (MSST). Different capital letters following variable values indicate significant differences between physiognomic groups/zones ($\alpha = 0.05$). Variables are defined in Table 2.1.

Zone	Elevation	MAP	MAAT	MAST	MSST	MWST
	Mean (range)					
	(m)	(mm)	(C)	(C)	(C)	(C)
Montane	1875B (1590 ^a -2285)	794B (735-917)	6.9B (5.1-8.0)	7.1B (5.9-7.8)	13.6B (11.6-14.6)	2.7A
Subalpine	2544A (2070-3060 ^b)	1021A (847-1137)	3.9A (3.0-6.1)	5.0A (4.2-6.6)	10.2A (8.8-12.7)	2.7A
Aspen	2426A ^c (1810-2750)	971A ^c (779-1121)	4.5A ^c (3.1-7.2)	5.4A ^c (4.3-7.3)	10.9A ^c (9.0-13.8)	2.7A

Zone	Ca_s	K_s	Mn_d	Parent material	Prevailing soil type ^d
	Mean (95 % confidence limits)				
	(mg/kg soil)	(mg/kg soil)	(µg/10 cm ² /6weeks)		
Montane	4050A (3009, 5247)	42A (38, 47)	0.4B (0.1, 2)	Fluvio-colluvial deposits	Pachic Argixerolls Typic Argixerolls
Subalpine	1894B (1332, 2553)	30B (27, 33)	22A (7, 75)	Glacial deposits Tertiary sediments	Typic Haplocryalfs Eutric Haplocryalfs
Aspen	4050A (2543, 3808)	35B (31, 38)	8A (3, 22)	Glacial deposits, fluvio-colluvial deposits, tertiary sediments, quartzite	Pachic, Typic Palecryolls Typic Argicryolls Typic Haplocryolls

^a Elevation of the lowest point of the research area; the montane zone can spread lower

^b Elevation of the highest point of the research area; the subalpine zone can spread higher

^c Significant difference from the montane zone because no aspen-dominated sites sampled in the lowest elevations

^d See Appendix A

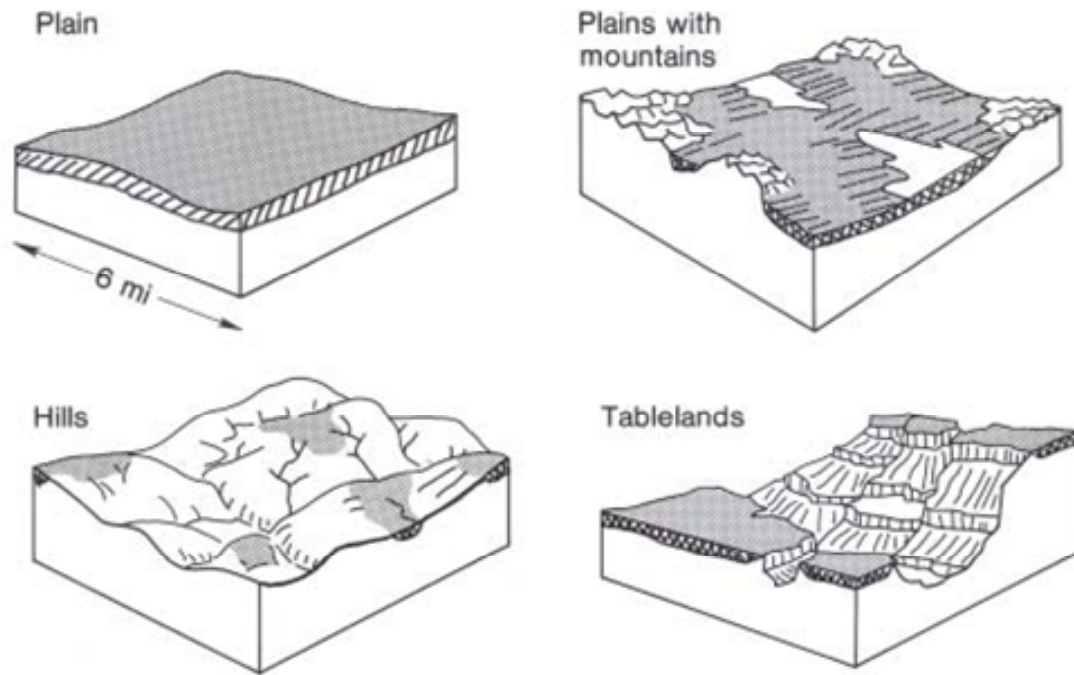


Figure 3.1. The zonal (climatic-climax) concept. Shaded areas, indirectly related to the study area, correspond with zonal sites in different types of a landscape (Bailey 1988).

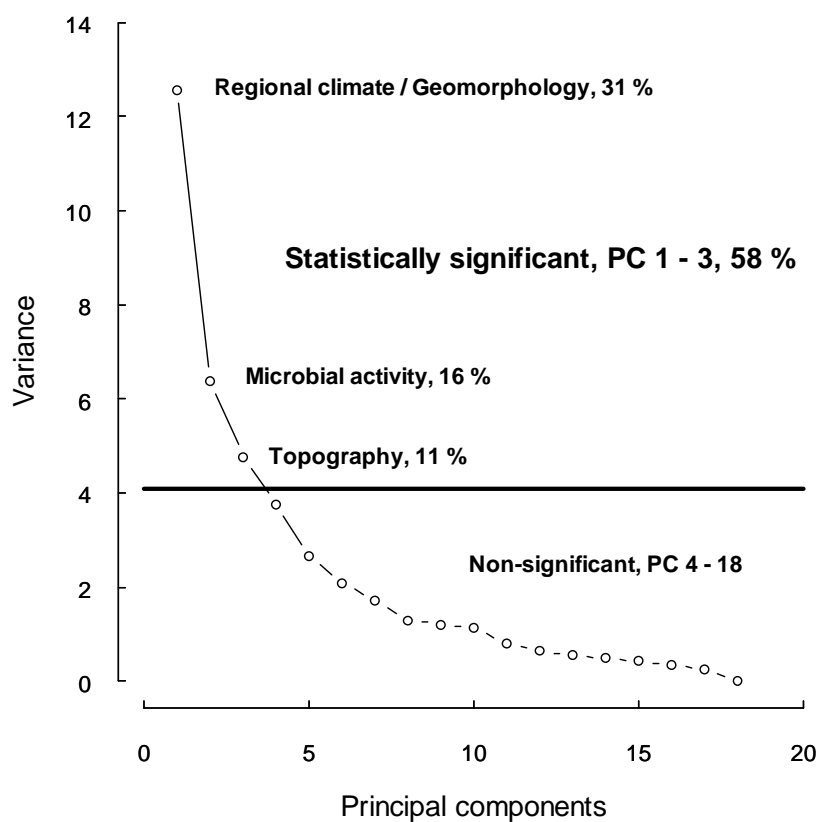


Figure 3.2. PCA scree plot on the conifer dataset. Significance and interpretation of the principal components. See text for explanation.

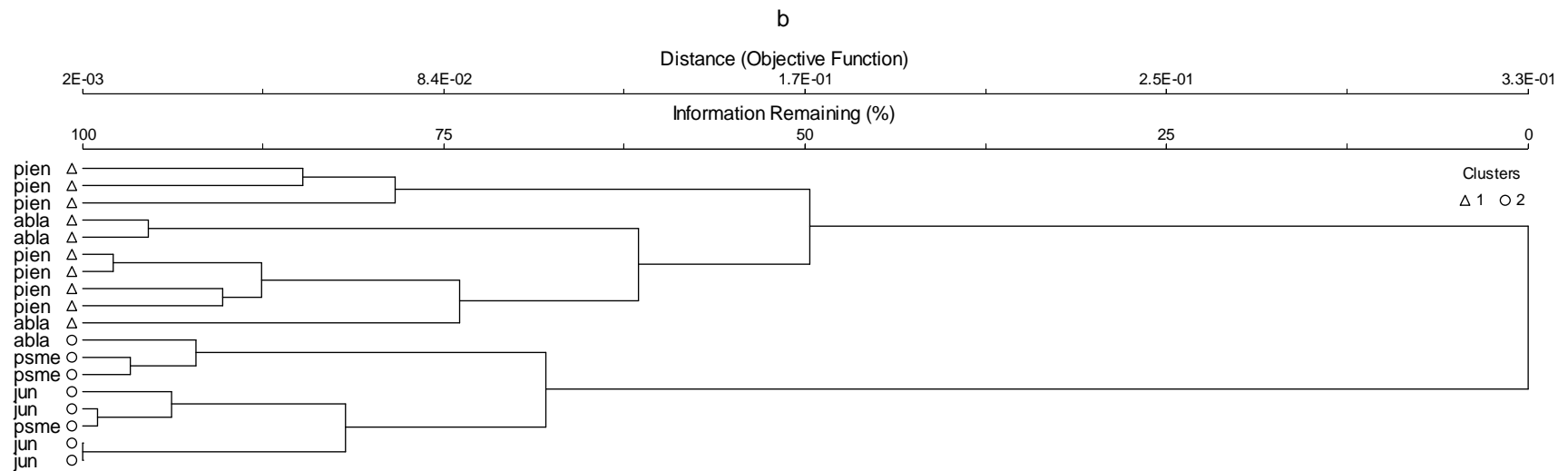


Figure 3.3. Cluster analysis dendrograms with the two cluster solution: a) the entire dataset including aspen-dominated sites; b) the conifer dataset. Rocky Mountain juniper (jun), Douglas fir (psme), aspen (potr), subalpine fir (abla), Engelmann spruce (pien) vegetation group.

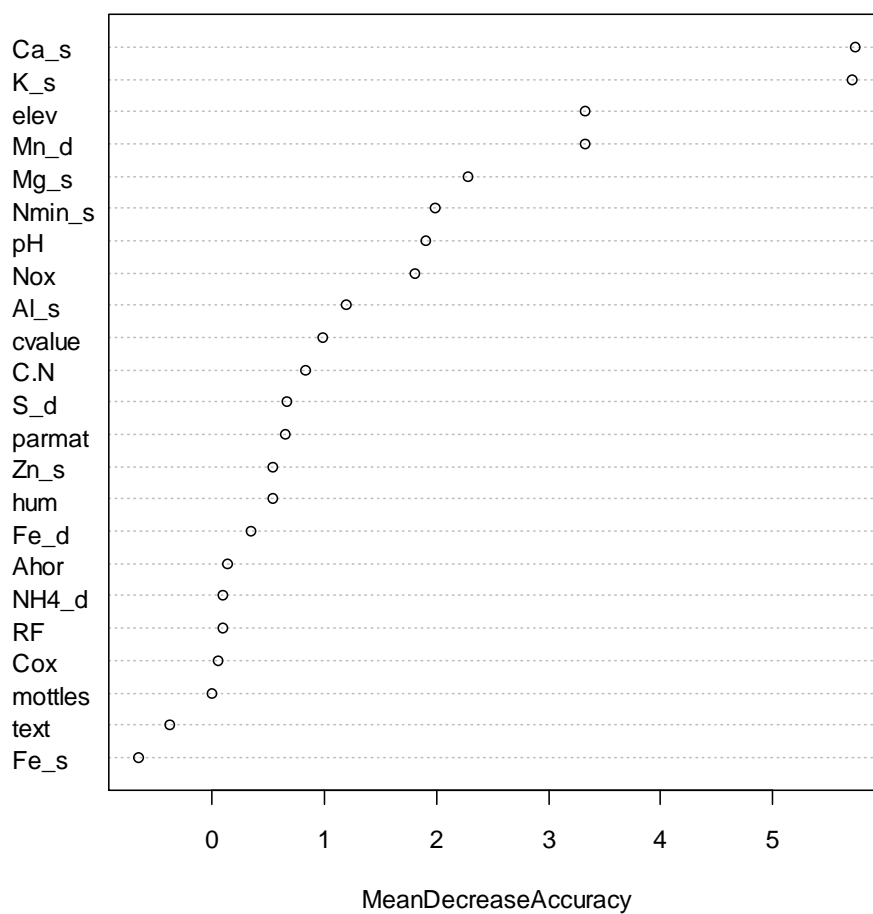


Figure 3.4. RandomForests analysis with variable importance order. Variables are defined in Table 2.1. See text for explanation.

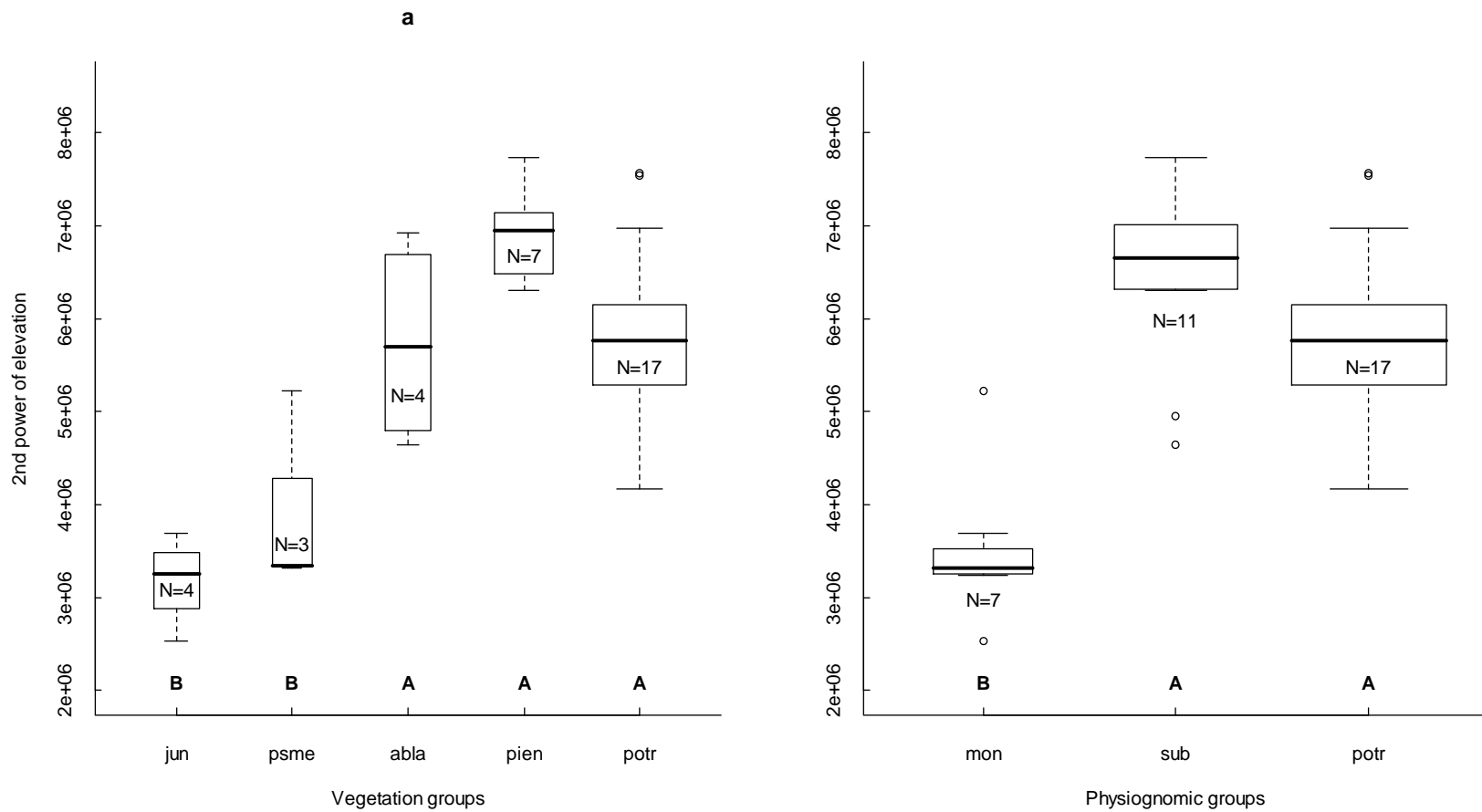


Figure 3.5. Relationship of vegetation and physiognomic groups with elevation. Different letters represent significantly different groups ($\alpha = 0.05$). Vegetation groups are defined in Fig. 3.3. See text for details.

CHAPTER 4

DIAGNOSTIC SPECIES AND FIDELITY CONCEPT IN VEGETATION
CLASSIFICATION IN THE ROCKY MOUNTAINS, NORTHERN UTAH³**Abstract**

The concept of diagnostic species and fidelity has been frequently used in European phytosociology. Based on these concepts, a statistical rather than intuitive approach was used to develop a vegetation classification in a mountainous study area in the Rocky Mountains of northern Utah.

This classification was derived from sampling of forest (spruce-fir, Douglas-fir, aspen, juniper and mahogany woodland) and non-forested (willow-riparian, low shrublands, tall-forb meadows, and sparse vegetation) ecosystems. One-hundred and fifty-seven vegetation sample plots were described by vascular plant species composition and basic habitat physiographic features (elevation, landform, slope, aspect, parent material).

Hierarchical cluster analysis reduced the original number of vegetation samples (relevés) to thirty-four meaningful vegetation units. For each species, fidelity and constancy were calculated and diagnostic species were identified for the vegetation units at the floristic level of alliances and associations, and then compared with indicator species of extensive habitat type classification in the central Rocky Mountains. Diagnostic species may have greater descriptive power than indicator species and potential value for a comprehensive ecosystem classification.

³ Coauthored by Antonin Kusbach, James N. Long, and Helga Van Miegroet

Introduction

“Without classification there can be no science of vegetation” - R. F. Daubenmire

Vegetation is together with animals, microbes and physical environment (i.e., climate and soils referred to hereafter as a habitat or site) a component of ecosystems of whatever size (Fosberg 1967, Pojar et al. 1987). As a distinctive landscape feature, vegetation was a fundamental component of land classifications (Daubenmire 1989). Recently, vegetation classification has become important as a communication tool in ecological research and in application of ecological information in planning, monitoring, conservation and management (Jennings et al. 2009).

Early land and vegetation classifications in the western U.S. such as the concept of potential natural vegetation (Küchler 1969) and habitat types (Daubenmire 1952), were based largely on constancy and dominance of species occurring in potential climax communities using the classic Braun-Blanquet method of sorting of phytocoenological tables. The most frequent and dominant species were considered indicators of vegetation units. This approach stressed species' abundance regardless of their presence in other vegetation units; and while quantitative, it was intuitive, and not statistical (Chytrý et al. 2002a). A statistical approach, on the other hand, accounts for the relationship between the number of unique vegetation samples (relevés) in vegetation units and in the total data set (Barkman 1989). This promising approach is associated with the concepts of diagnostic species (Whittaker 1962, Westhoff and van der Maarel 1973, Jennings et al. 2008, 2009) and the concept of fidelity (Chytrý et al. 2002a). Some approaches to the concept of diagnostic species are stricter (e.g., European), constrained to character species (i.e., species occurring in a single vegetation unit for which it is characteristic)

and differential species (i.e., species occurring in a few vegetation units which they discriminate) (Mueller-Dombois and Ellenberg 1974, Bruelheide 2000, Chytrý et al. 2002a). Alternatively, some approaches apply a looser interpretation, including also constant species (i.e., the most frequent) and dominant species (i.e., species with high cover) (Westhoff and van der Maarel 1973, Brohman and Bryant 2005, Winthers et al. 2005, FGDC 2008, Jennings et al. 2008, 2009).

Fidelity is a measure of species concentration, (*sensu* Chytrý et al. 2002a) in vegetation units. It is based on species' frequencies observed within a vegetation unit compared with expected frequencies if the species' distributions were random, i.e., taking also out-of-unit species occurrence into consideration (within a total data set) (Barkman 1989, Chytrý et al. 2002a). Fidelity is thus a measure of species-unit association. In contrast, constancy as a usual and more common measure of species frequency in vegetation units in the past, does not take into consideration species outside its primary vegetation unit. It has been suggested that the use of fidelity relative to diagnostic species may increase the general validity of vegetation types in large phytosociological data bases representing broad taxonomic units such as orders or classes, and also in smaller data sets representing a geographically small but diverse area (with large ecological variation) (Chytrý et al. 2002a).

There are two major tasks in vegetation classification: (1) to distinguish meaningful (i.e., interpretable) groups of species within the original dataset and thereby reducing the number of relevés to a fewer number of more or less similar and meaningful **vegetation units**; and (2) to identify **diagnostic species** within these vegetation units.

We used the fidelity concept to analyse vegetation in the study area representing the

Rocky Mountains, northern Utah. Diagnostic species were not identified intuitively based on subjective constancy and dominance thresholds. Rather they were identified based on fidelity calculations, a more objective and statistical approach. We expect to reveal a descriptive power of diagnostic species and introduce meaningful vegetation units and therefore, a better understanding of vegetation patterns, particularly the distribution of species assemblages in the Rocky Mountains of northern Utah. Our specific objectives are to: (1) perform an objective classification based on specification of feasible vegetation units; and (2) identify species diagnostic of particular vegetation units and potentially useful for recognition of these vegetation units in the study area.

Methods

Study area

Franklin Basin and the T.W. Daniel Experimental Forest make up the study area (ca 16,000 ha, and ca 1,000 m of vertical extent) (Fig. 2.2). Franklin Basin (FB, 15,000 ha) is a montane-subalpine area situated between the Bear River Range and the Wasatch Range in the central Rocky Mountains on the Utah and Idaho border. The T.W. Daniel Experimental Forest (TWDEF, ca 1000 ha) is situated on a high ridge plateau of the Wasatch Range (10 km to the southeast of FB).

According to Bailey (1998) and McNab et al. (2007), the study area occurs within M331 Southern Rocky Mountains Steppe-Open Woodland-Coniferous Forest-Alpine Meadow Province, “D” Overthrust Mountain Section, “n” Northern Wasatch Range, and “o” Bear River Front Range Subsections. The mean total annual precipitation ranges from 720 to 1250 mm and mean annual air temperature ranges from 2.4 to 5.7 °C for

Temple Fork, Tony Grove Lake, Franklin Basin, and Utah State University (USU) Doc Daniel weather stations (<http://www.wcc.nrcs.usda.gov/snow/>).

The terrain is mountainous, rocky and steep with occasional flat to gently sloping high ridge-plateaus and benches. The elevation ranges from 2050-3060 m across the two study sites. The highest area of the Bear River Range was glaciated during the Pleistocene as manifested by glacial geomorphologic features like moraines, U-shaped valleys, erratics, and irregular glacial deposits (Atwood 1909, Young 1939, Degraff 1976). The study area is mostly built from calcareous sedimentary rocks (limestone, dolomite) with interlayered quartzite, and from Tertiary sediments consisting of grit, conglomerate, and siltstone of the Wasatch Formation at the TWDEF site. The soils are formed in residuum, colluvium, alluvium, glacial till and outwash, and occur on diverse landforms such as cliffs, talus slope, moraines, karst valleys, mountain slopes, landslides, plains, valleys, depressions, ravines, and wetlands (Schoeneberger et al. 2002).

Over half of the study area is occupied by forest ecosystems including Engelmann spruce (*Picea engelmannii*), subalpine fir (*Abies lasiocarpa*), Douglas-fir (*Pseudotsuga menziesii*), aspen (*Populus tremuloides*), and woodland ecosystems including mountain mahogany (*Cercocarpus ledifolius*) and Rocky Mountain juniper (*Juniperus scopulorum*). Substantial changes in fire regimes, often in combination with cutting and grazing, have led to dramatic changes in the structure and the age-class distribution of forest stands. In many places, 100- to 140-year-old stands are now predominant (Long 1994). Forests in the study area are thus characterized by mid- and late-seral stages where forest understory is usually well developed (Pfister and Arno 1980). Non-forested ecosystems include willow-riparian (*Salix spp.*) and wetlands, low shrublands (*Artemisia*

spp.), tall-forb meadows, and sparse vegetation on talus and rock outcrops, which may represent stable or temporary communities. Despite human impacts in last 120 years, the study area is considered relatively natural in terms of plant species composition (Bird 1964).

Data collection

We collected vegetation data (relevés) on 157 sample plots across the study area in the summers (late May-August) of 2006 and 2007. Sampling followed the Existing Vegetation Classification and Mapping Technical Guide (Brohman and Bryant 2005), and Terrestrial Ecological Unit Inventory Technical Guide (Winthers et al. 2005). After field reconnaissance, we sampled vegetation across a broad range of the physical environment in order to capture as much environmental variation as possible. In an effort to minimize the major influence of historical factors such as fires and logging and stress the impact of the physical environment on vegetation, we focused on mature, late-successional, and relatively stable plant communities. In the case of forest vegetation, this condition was characterized by advance regeneration of potential climax tree species (Pfister and Arno 1980, Pojar et al. 1987). We sampled stands reasonably uniform in physiognomy, floristic composition and environment (Jennings et al. 2009). We avoided ecotones, i.e., habitats in transition, where important environmental factors merge, as well as degraded or atypical stands. Because 6 sample plots did not fulfill these conditions entirely, they were excluded from the original scheme of 163 sample plots described in earlier chapters. A stratified (based on vegetation physiognomy) fixed (subjective selection) sampling design was used with sample plot size of 1000 m² for

forest and 100 m² for non-forested ecosystems and three replicates were considered the minimum for defining a preliminary vegetation unit (Podani 2000, Brohman and Bryant 2005, Jennings et al. 2009). The plots were usually circular, but the shape was adjusted according to the character of habitat, e.g., linear for riparian vegetation. Each sample plot was described by complete enumeration of vascular plant species abundances (canopy cover percentage for forest and ground cover percentage for forest and non-forested understory), strata (trees, shrubs, and understory), and by basic habitat (site) physiographic characteristics (e.g., elevation, slope, aspect, landform, parent material).

Data analysis

Vegetation analysis followed the Existing Vegetation Classification and Mapping Technical Guide (Brohman and Bryant 2005), the National Vegetation Classification Standard (FGDC 2008), and Standards for Associations and Alliances of the U.S. National Vegetation Classification (Jennings et al. 2009). We ran an analysis of the 157 relevés representing 327 species (Appendix B).

There were two steps in the analysis: (1) partitioning of the dataset into meaningful vegetation units/clusters; and (2) characterization of vegetation units based on diagnostic species.

To partition the community data we used two common approaches, represented by agglomerative and divisive methods. The agglomerative approach involved hierarchical clustering with Ward's and Flexible beta linkage methods combined with Euclidean and Sørensen distance measures (Bray and Curtis 1957, Ward 1963). Percentage cover was standardized by both logarithmic and less drastic square root transformations. The

divisive approach was represented by Modified TWINSpan, which prevents unsubstantiated division of homogeneous clusters (Roleček et al. 2009). Here, we applied three measures of within-cluster heterogeneity: Whittaker's beta (Whittaker 1960); total inertia (Greenacre 2000); and chord distance (Orlóci 1967). We deleted rare species with just one occurrence in the entire dataset before partitioning the dataset.

Using both the agglomerative and divisive approaches, nine alternative partitioning methods were assessed by OptimClass. It is a method for identifying the optimal partitioning of a set of relevés (Tichý et al. 2009). In OptimClass, for each partitioning method, the total number of faithful species was calculated by Fisher's exact test as a measure of species-to-unit fidelity for presence/absence data (Chytrý et al. 2002a). The best partitioning method was chosen from all solutions based on the total number of faithful species and interpretability of vegetation units. This best solution was thus a compromise between statistical assessment represented by Fisher's exact test and ecological feasibility represented by meaningful, i.e., interpretable vegetation units (Chytrý et al. 2002a). We removed small-member clusters, i.e., vegetation units represented just by one and two relevés within the best partitioning solution because small units created at high hierarchical levels are considered outliers (Tichý et al. 2009, McCune and Grace 2002, Jennings et al. 2009); they could be either undersampled or rare in the study area. Finally, we repeated the partitioning for the best method, i.e., we achieved the appropriate number of vegetation units/clusters based on interpretability, size of vegetation units and the number of faithful species.

We used fidelity calculation to determine faithful species in the second step of the analysis with the expectation that this method should improve the general validity of

vegetation types. There are many measures of fidelity such as u value (Bruehlheide 2000), chi-square statistic, G statistic, Fisher's exact test and Indicator Value Index (Dufrêne and Legendre 1997); we used **phi coefficient** of association (Sokal and Rohlf 1995). In common in other fidelity measures, it is based on comparison of observed/expected frequencies, i.e., probability of species' non-random occurrence (Chytrý et al. 2002a). The main advantage of this coefficient is that it is, unlike other fidelity measures, independent of the size of the dataset (the number of relevés in the dataset); therefore, it is useful to compare species fidelities among differently sized datasets and vegetation units (Chytrý et al. 2002a). The phi coefficient value ranges from -1 to +1 (-100 to +100 %) with positive value indicating the species-unit co-occurrence is more often than expected by chance only. If the species is completely faithful to the unit the value is 1 (100 %) (Chytrý et al. 2002a).

To minimize possible effects of unequal-sized vegetation units, we calculated the phi coefficient with presence/absence data after standardization of the units' size (number of relevés) (Tichý and Chytrý 2006). Species presence/absence provides a better fidelity measure than cover abundance because it is less affected by high fluctuations of species cover and observer bias (Chytrý et al. 2002a). Statistical significance of the phi coefficient was assessed by Fisher's exact test with $p < 0.05$.

There is no fixed threshold for the phi coefficient. We set a threshold of $\text{phi} > 40\%$ for faithful species. A higher threshold would be too restrictive producing a low number of faithful species. Conversely, a lower threshold would produce an unnecessarily large number of 'faithful' species with limited diagnostic power in this relatively small dataset. Diagnostic species were defined based on the following thresholds: faithful species, $\text{phi} >$

40 %; constant species, constancy (frequency) > 60 % (Mueller-Dombois and Ellenberg 1974); and dominant species, cover > 5 % This relatively low cover value was chosen in light of sparse communities such as woodland and high elevation open-canopy conifer forests that would be otherwise underestimated compared to thick forest communities. Character species were all faithful species (with $\phi > 40\%$) occurring in a single vegetation unit only and thus highly faithful to it. Differential species were all faithful species occurring in a few vegetation units; in fact, these species discriminate these units.

As analysis outputs we created: (1) an **advanced combined synoptic table**; and (2) an **analysis of synoptic table** where diagnostic species are explicitly listed. In the advanced combined synoptic table, species were sorted by the ϕ coefficient threshold value ($\phi > 40\%$). This fidelity information was combined and compared with constancy (frequency) value. Both the advanced combined synoptic table and analysis of synoptic table are available as an automatic procedure in the JUICE program (Tichý 2002).

The relationship among vegetation units and “goodness” of each unit was based on number of faithful species and the average positive fidelity calculated as a simple mean of non-negative fidelity values. We also calculated species richness as a simple number of species within a vegetation unit and an indicator of local (alpha) species diversity (e.g., Whittaker 1972, McCune and Grace 2002, Colwell et al. 2004). But this diversity measure should be considered with caution because it is highly dependent on the number of relevés (Gotelli and Colwell 2001, Colwell et al. 2004). We considered good, i.e., well defined, units as those with three and more faithful species and high average positive fidelity (> 30 %). Weak, i.e., poorly defined, units were those with few or no faithful

species and low average positive fidelity ($< 30\%$) (Chytrý et al. 2002a).

Based on these outputs, we proposed vegetation types as species assemblages at the floristic level of vegetation alliances and associations (Grossman et al. 1998, FGDC 2008, Jennings et al. 2008, 2009). Alliances were suggested as compilations of diagnostic species which are differential and constant (frequent), and with particular preference for tree species (FGDC 2008, Jennings et al. 2009). Associations were suggested as compilations of character species. In the case of a weak (poorly defined) vegetation unit, because no faithful species were present, the most constant (frequent) and dominant species were considered as diagnostic.

JUICE software version 7. 0. 48. (Tichý 2002) and PC-ORD 5 (McCune and Mefford 2006) were used in the partitioning analysis. Diagnostic species within vegetation units were analyzed using JUICE.

Results

Partitioning the dataset

The more relevés in a vegetation unit the better that unit may be defined. Especially in the case of our small dataset we expected greater descriptive power of diagnostic species inside larger vegetation units than in small units (less than three relevés) where there was a greater possibility of faithful species presence solely by chance. Therefore, when appropriate, we opted for larger vegetation units. Using OptimClass, we chose the partitioning based on Ward/Euclidean/square root transformation as the best of the nine alternative solutions. It retained both a high total number of faithful species (395) and relatively large units (32). In contrast, the Ward/Sørensen solution provided the highest

number of faithful species (403) but also detailed clustering representing 48 vegetation units, some of which were relatively small and potentially weak (Fig. 4.1). The flexible beta and Modified TWINSpan solutions did not produce as many faithful species as Ward/Euclidean clustering (e.g., Brown 2006).

The repeated division after the removal of small-member clusters, i.e., vegetation units represented just by one and two relevés within the best partitioning solution, resulted in the appropriate partitioning into 34 vegetation units. These units were readily interpretable and had meaningful size related to the size of the entire data set. Additional division yielded units that were too small; they became less interpretable despite the higher number of faithful species.

Characterization of vegetation units

The advanced combined synoptic table displayed differences within and between the vegetation units. Species were sorted by decreasing fidelity at the phi coefficient positive threshold value ($\phi > 40$) into diagonally arranged blocks (Tichý 2002). Fidelity values were complemented by constancy/frequency values (Appendix C). It is obvious that consistency between fidelity and constancy was not perfect. Rather, species with high fidelity did not inevitably have high constancy and species completely constant were not necessarily faithful. This was true for the majority of tree species and, for example, some undergrowth species such as *Osmorhiza chilensis*, *Pedicularis racemosa* and *Berberis repens*, important indicator species in habitat typing.

The diagonal blocks, unique for each vegetation unit, consisted of character species. Additionally, more or less regular blocks or single faithful species lower in the synoptic

table occurring in a few vegetation units, represented differential species. This sorting helped to distinguish character and differential species and together with the analysis of synoptic table facilitated characterization of the vegetation units by diagnostic species (Appendix D). Potential alliances and associations were identified for all units with short description of the habitat. Table 4.1 includes forest (conifer and broad-leaved) and woodland (conifer and broad-leaved) vegetation units; Table 4.2 includes non-forested/rangeland (shrubland, dwarf-shrubland, herbaceous and sparse vegetation) vegetation units (Lund 2006, Grossman et al. 1998, FGDC 2008). Habitat characterization of the vegetation units was based on simple physiographic factors and elevation stratification relative to the elevation range within the study area: low elevation < ca 2350 m; high elevation 2350-2650 m; subalpine 2650-2950 m; alpine > 2950 m.

Some species were present in the majority of vegetation units, e.g., *Symphoricarpos oreophilus*, *Thalictrum fendlerii*, *Osmorhiza chilensis*, *Paxistima myrsinites*. In contrast to these generalists, other species appear to be habitat specialists, as they were restricted to as few as two units. e.g., *Zigadenus elegans*, *Cercocarpus ledifolius*, *Juniperus scopulorum*. Interestingly, several tree species appear in the majority of forest vegetation units and showed a high degree of coexistence, e.g., *Picea engelmannii* with *Abies lasiocarpa* or *Populus tremuloides* with conifers (Appendix C, D).

Of the 34 vegetation units, 23 were good (well defined) having three and more faithful species and the average positive fidelity greater than 30. Eleven units were weak (poorly defined) with less than three faithful species and the average positive fidelity less than 30. The average species richness of weak groups was 62.3 whereas the richness of good groups was 51.8. Aspen units appeared to be somewhat weak (the average species

richness 66.3) except the unit on wet habitat (vegetation unit 20, species richness 43) (Appendix C, Table 4.1, 4.2).

Discussion

Interpretation of classification

The good and weak vegetation units were associated with interesting patterns which may facilitate their explanation and interpretation. For example, based on habitat description (Table 4.1, 4.2), the good units were associated with extreme environments such as wet, rocky, or strongly calcareous sites. In contrast, the weak units were associated with moderate environments such as well drained, non-skeletal sites on moderately deep soils. The distribution of those species which we interpret to be site specialists appeared to be driven by environmental factors such as soil moisture and pH, whereas occurrence of site generalists may be explained by spatial/temporal factors, such as dispersal processes and patch dynamics (Pandit et al. 2009). Additionally, vegetation units in intermediate environmental conditions tended to have greater species richness represented by mostly habitat generalists and fewer specialists (Hájek et al. 2007). For example, units 1, 2, 20, 32, 33, 34 (Table 4.1, 4.2) with wet conditions and units 10, 11, 12, 27, 28 with dry calcareous conditions had the highest number of faithful species as potential specialists and relatively low species richness whereas the units 3, 6, 16, 17, 18, 19, 21, 22, 23 with intermediate environmental conditions had few or no faithful species but high species richness mostly consisting of generalists.

In the study area, habitats with intermediate environmental conditions (e.g., slightly undulating or flat terrain, moderate-deep soils, little or no rocks) tend to be accessible

and, therefore, potentially subject to greater human-caused disturbances such as logging and domestic livestock use. In general, disturbed sites probably have fewer site specialists and many common species (Gaublomme et al. 2008). This characterization is consistent with our data; accessible, presumably more highly disturbed sites do in fact support a richness of site generalists and a low number of specialists that may reflect site history, e.g., grazing in past 120 years (Appendix C, Table 4.1, 4.2) (Hájek et al. 2007, Pärtel et al. 2001).

The 11 weak units may reflect the challenge associated with sampling special habitats. In some cases it was difficult to avoid ecotones, for example, riparian vs. valley bottom transition (unit 3), and delineate a real mosaic of narrow corridors affected by high ground water table. Another example is a dry habitat of cliffs with inter-layered benches (unit 31), resulting in a sample representing a mosaic of rocks and deeper soil. Additionally, in the case of high-elevation open-canopy spruce-fir forest, typical forest species were described together with species typical of non-forest vegetation units. Because we analyzed both forest and non-forested ecosystems together, “goodness” of these high-elevation forest units could be affected by difficulties connected to separation of forest from non-forested vegetation units (e.g., units 3, 8, 9).

Finally, dataset size may over-emphasis or strengthens the ‘goodness’ of vegetation units; some units appear to be good in our dataset but their status might change with spatial expansion of the sampling, likely resulting in an increasing number of species and potential expansion of environmental conditions (Chytrý et al. 2002a). Listed faithful species should be interpreted with caution and within the context of this study, but we assume they may represent regional character or differential species (Willner 2001,

Chytrý et al. 2002a). The diagnostic species were interpreted for a relatively narrow geographical area, yet covered a broad environmental range. Therefore, we suggest this local geographical context may be close to regional (e.g., Chytrý et al. 2002b).

Potential value of classification

The most extensive vegetation classifications of forests in the central Rocky Mountains were based on habitat and community typing (e.g., Daubenmire 1952, Pfister and Arno 1980, Mauk and Henderson 1984, Mueggler 1988). Habitat and community type analyses did not include woodland and riparian plant communities in contrast to our classification which includes a much broader range of vegetation types, including non-forested communities.

In these earlier classifications, indicator species were both highly constant and dominant within a particular vegetation unit (habitat or community type) and their selection was subjective and largely quantitative. Moreover, looking at the distribution of these indicators more closely reveals that they do not occur only inside one habitat/community type (e.g., *Berberis repens*, *Osmorhiza chilensis* and *Pedicularis racemosa* habitat types of *Pseudotsuga menziesii*, *Picea engelmannii* and *Abies lasiocarpa* series) but also across habitat series (e.g., consider *Picea engelmannii* and *Abies lasiocarpa* occurrence throughout these series). Consequently, these habitat/community type indicator species may not be useful in discriminating between environmental conditions. Our analysis revealed faithful species different from indicator species of habitat/community types. We assume that faithful (character or differential) species are better vegetation unit descriptors than common and abundant species

(generalists) because they are more reflective of the underlying environment (Chapter 5).

In the case of the weak units (i.e., conifer units 3, 6, 9; aspen units 16, 17, 18, 19; and shrubland units 21, 22, 23, 31) additional research is needed to decide if sampling was appropriate. For example, classification of environmentally intermediate aspen communities is difficult using the fidelity and diagnostic species approach. Because of generally rich understory cover, and presence of many generalists in contrast to few species with high fidelity, ecotones between aspen units are both abundant and ambiguous. This vast floristic variability of aspen units is consistent with exclusion of aspen communities from vegetation geo-climatic zonation because of their environmental amplitude (Chapter 3) and their not always clear successional status (Mueggler 1988). We suspect a more detailed environmental description and sampling of aspen habitat (e.g., pH of soils or nutrient characteristics) might lead to better vegetation discrimination of aspen units.

Our vegetation classification revealed that major tree species such as Engelmann spruce (*Picea engelmannii*), subalpine fir (*Abies lasiocarpa*), Douglas-fir (*Pseudotsuga menziesii*), limber pine (*Pinus flexilis*) and aspen (*Populus tremuloides*) appear in a high degree of coexistence with other tree species (Appendix C, D). This finding is consistent with vegetation geo-climatic zonation (Chapter 3).

Summary and conclusions

Using the concept of diagnostic species and fidelity, we created a vegetation classification based on vegetation sampling of the floristically and environmentally complex study area in the Rocky Mountains of northern Utah. Based on cluster analysis

of a community data set, we identified thirty-four vegetation units. For each species, fidelity and constancy was calculated. Then, diagnostic species were determined for each vegetation unit at the floristic level of alliances and associations.

We compared our approach with habitat and community type classification in the central Rocky Mountains. We suggest that: (1) strict separation of tree species is unsubstantiated from a classification standpoint; (2) diagnostic species are more strongly associated with their underlying physical environment than indicator species *sensu* habitat type; and (3) our vegetation classification results from complex analysis of existing vegetation in broad range of ecosystems (forest, woodland, riparian, non-forested).

Any taxonomic classification based strictly on vegetation is somewhat limited in ecological interpretations. Ultimately, both vegetation and the physical environment should be included as a part of a comprehensive ecosystem classification.

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Table 4.1. Forest and woodland vegetation units, proposed alliance and association compilations and habitat description.

Unit	Alliance	Association	Label	Habitat
1	<i>Abies lasiocarpa</i>	<i>Galium boreale</i>	Abla-Pien/Gabo	Low elevation riparian
	<i>Picea engelmannii</i>	<i>Potentilla gracilis</i>		
	<i>Angelica arguta</i>	<i>Epilobium ciliatum</i>		
	<i>Lonicera involucrata</i>	<i>Potentilla fruticosa</i>		
	<i>Salix boothii</i>	<i>Salix bebbiana</i>		
2		<i>Zigadenus elegans</i>	Abla-Pien/Ziel	High elevation wetlands
		<i>Aconitum columbianum</i>		
		<i>Salix wolfii</i>		
		<i>Pedicularis groenlandica</i>		
		<i>Senecio triangularis</i>		
3	<i>Abies lasiocarpa</i>	<i>Thalictrum fendleri</i>	Abla-Pien/Thfe	High elevation conifer valley bottoms
	<i>Picea engelmannii</i>	<i>Pedicularis racemosa</i>		
	<i>Arnica cordifolia</i>	<i>Rudbeckia occidentalis</i>		
		<i>Ligusticum filicinum</i>		
		<i>Osmorhiza chilensis</i>		
4		<i>Lonicera utahensis</i>	Abla-Pien/Lout	Subalpine shady slopes
		<i>Ranunculus jovis</i>		
		<i>Arnica latifolia</i>		
		<i>Pyrola secunda</i>		
		<i>Aster occidentalis</i>		

Unit	Alliance	Association	Label	Habitat
5	<i>Abies lasiocarpa</i> <i>Picea engelmannii</i> <i>Populus tremuloides</i>	<i>Ligusticum porteri</i> <i>Poa bolanderi</i> <i>Hieracium albiflorum</i> <i>Osmorhiza chilensis</i> <i>Ligusticum filicinum</i>	Abla-Pien-Potr/Lipo	Subalpine undulating plateaus
6		<i>Osmorhiza chilensis</i> <i>Rudbeckia occidentalis</i> <i>Thalictrum fendleri</i> <i>Pseudotsuga menziesii</i>	Abla-Pien-Potr/Osch	Undulating glacial moraines
7	<i>Abies lasiocarpa</i> <i>Picea engelmannii</i> <i>Pseudotsuga menziesii</i> <i>Carex geyeri</i>	<i>Rubus parviflorus</i> <i>Goodyera oblongifolia</i> <i>Sorbus scopulina</i> <i>Chimaphila umbellata</i> <i>Pyrola secunda</i>	Abla-Pien-Psme/Rupa	High elevation shady skeletal slopes
8		<i>Rubus idaeus</i> <i>Juncus parryi</i> <i>Poa cusickii</i>	Abla-Pien-Psme/Ruid	Quartzite talus Sparse vegetation
9	<i>Abies lasiocarpa</i> <i>Picea engelmannii</i> <i>Pinus flexilis</i> <i>Pseudotsuga menziesii</i>	<i>Anemone multifida</i> <i>Thlaspi montanum</i> <i>Pedicularis racemosa</i> <i>Aster engelmannii</i> <i>Penstemon leonardii</i>	Abla-Pien-Pifl/Anmu	Subalpine rocky calcareous slopes/flats

Unit	Alliance	Association	Label	Habitat
10	<i>Juniperus scopulorum</i>	<i>Lomatium grayi</i>	Jusc/Logr	Very dry skeletal conifer woodland slopes
	<i>Artemisia tridentata</i>	<i>Calochortus nuttallii</i>		
	<i>Elymus spicatus</i>	<i>Artemisia arbuscula</i>		
	<i>Balsamorhiza sagittata</i>	<i>Zigadenus paniculatus</i>		
		<i>Ceanothus velutinus</i>		
11	<i>Cercocarpus ledifolius</i>	<i>Mertensia oblongifolia</i>	Cele/Meob	Dry calcareous slopes
	<i>Artemisia tridentata</i>	<i>Crepis acuminata</i>		
	<i>Pseudotsuga menziesii</i>	<i>Viola purpurea</i>		
	<i>Pinus flexilis</i>	<i>Chrysothamnus viscidiflorus</i>		
	<i>Elymus spicatus</i>	<i>Ceanothus velutinus</i>		
12		<i>Petradoria pumila</i>	Cele/Pepu	Extremely dry calcareous rocks and cliffs
		<i>Solidago nana</i>		
		<i>Aster ascendens</i>		
		<i>Comandra umbellata</i>		
13	<i>Pseudotsuga menziesii</i>	<i>Acer grandidentatum</i>	Psme-Abla/Acgr	Low elevation shady slopes
	<i>Abies lasiocarpa</i>	<i>Arnica cordifolia</i>		
	<i>Symphoricarpos oreophilus</i>	<i>Smilacina racemosa</i>		
		<i>Viola adunca</i>		
14	<i>Pseudotsuga menziesii</i>	<i>Linanthastrum nuttallii</i>	Psme-Abla-Pifl/Linu	Subalpine sunny calcareous slopes
	<i>Pinus flexilis</i>	<i>Aster glaucodes</i>		
	<i>Abies lasiocarpa</i>	<i>Acer glabrum</i>		
	<i>Juniperus communis</i>	<i>Berberis repens</i>		

Unit	Alliance	Association	Label	Habitat
15	<i>Pseudotsuga menziesii</i> <i>Pinus flexilis</i> <i>Symphoricarpos oreophilus</i>	<i>Astragalus tenellus</i> <i>Bromus anomalus</i> <i>Shepherdia canadensis</i> <i>Lomatium graveolens</i> <i>Leucopoa kingii</i>	Psme-Pifl/Aste	High elevation sunny calcareous slopes
16	<i>Populus tremuloides</i> <i>Abies lasiocarpa</i> <i>Picea engelmannii</i>	<i>Poa leptocoma</i> <i>Lupinus argenteus</i> <i>Artemisia ludoviciana</i>	Potr-Abla-Pien/Pole	Moderate quartzite slopes
17	<i>Populus tremuloides</i> <i>Abies lasiocarpa</i> <i>Lathyrus pauciflorus</i>	<i>Valeriana occidentalis</i> <i>Tragopogon dubius*</i> <i>Rudbeckia occidentalis</i>	Potr-Abla/Vaoc	Toe slopes, glacial moraines, moderate slopes
18		<i>Symphoricarpos oreophilus</i> <i>Thalictrum fendleri</i> <i>Senecio serra</i>	Potr-Abla/Syor	Undulating glacial moraines
19		<i>Scrophularia lanceolata</i> <i>Valeriana occidentalis</i> <i>Symphoricarpos oreophilus</i>	Potr-Abla/Scla	Moderate alkaline/calcareous mid-slopes
20	<i>Populus tremuloides</i> <i>Veratrum californicum</i> <i>Carex pachystachya</i> <i>Elymus cinereus</i>	<i>Cynoglossum officinale</i> <i>Dactylis glomerata*</i> <i>Allium bisceptrum</i> <i>Ranunculus orthorhynchus</i> <i>Bromus ciliatus*</i>	Potr/Cyof	Moist and wet valley bottoms

* Introduced species

Table 4.2. Non-forested vegetation units, proposed alliance and association compilations and habitat description.

Unit	Alliance	Association	Label	Habitat
21	<i>Artemisia spiciformis</i> <i>Symphoricarpos oreophilus</i>	<i>Chrysothamnus viscidiflorus</i> <i>Elymus cinereus</i> <i>Senecio serra</i> <i>Agastache urticifolia</i>	Arsp/Chvi	Low elevation valley bottoms/slopes
22		<i>Poa arnowiae</i> <i>Eriogonum heracleoides</i> <i>Penstemon cyananthus</i> <i>Eriogonum umbellatum</i>	Arsp/Poar	High elevation valley bottoms/slopes
23		<i>Agastache urticifolia</i> <i>Rudbeckia occidentalis</i> <i>Geranium viscosissimum</i> <i>Potentilla glandulosa</i>	Arsp/Agur	High elevation valley bottoms
24		<i>Wyethia amplexicaulis</i> <i>Poa bulbosa</i> * <i>Antennaria parvifolia</i> <i>Poa secunda</i> <i>Artemisia arbuscula</i>	Arsp/Wyam	Low elevation undulating moraines
25	<i>Rudbeckia occidentalis</i> <i>Ligusticum filicinum</i> <i>Delphinium occidentale</i>	<i>Ranunculus adoneus</i> <i>Carex multcostata</i> <i>Polygonum douglasii</i> <i>Stipa nelsonii</i> <i>Bromus carinatus</i>	Ruoc/Raad	Subalpine depressions/colluvial outwashes

Unit	Alliance	Association	Label	Habitat
26	<i>Agastache urticifolia</i>	<i>Aster integrifolius</i>	Agur/Asin	Subalpine flats/slopes
	<i>Osmorhiza occidentalis</i>	<i>Heuchera parviflora</i>		
	<i>Geranium viscosissimum</i>	<i>Elymus lanceolatus</i>		
	<i>Helianthella uniflora</i>			
27	<i>Linum kingii</i>	<i>Clematis occidentalis</i>	Liki/Cloc	Subalpine sunny slopes
	<i>Linum lewisii</i>	<i>Linanthastrum nuttallii</i>		
	<i>Penstemon compactus</i>	<i>Aster glaucodes</i>		
	<i>Lomatium graveolens</i>	<i>Berberis repens</i>		
28		<i>Hymenoxys acaulis</i>	Liki/Hyac	Alpine talus and rocks
		<i>Phlox hoodii</i>		Shallow soils
		<i>Synthyris pinnatifida</i>		
		<i>Phlox pulvinata</i>		
		<i>Anemone multifida</i>		
29	<i>Ivesia gordonii</i>	<i>Juncus parryi</i>	Ivgo/Jupa	Quartzite talus and rocks
	<i>Leucopoa kingii</i>	<i>Apocynum androsaemifolium</i>		Sparse vegetation
	<i>Eriogonum umbellatum</i>	<i>Epilobium canum</i>		
	<i>Penstemon leonardii</i>	<i>Solidago multiradiata</i>		
	<i>Potentilla glandulosa</i>	<i>Arenaria congesta</i>		
30		<i>Monardella odoratissima</i>	Ivgo/Mood	Subalpine talus and rocks
		<i>Eriogonum caespitosum</i>		Shallow soils
		<i>Penstemon humilis</i>		
		<i>Artemisia dracunculus</i>		
		<i>Heuchera parvifolia</i>		
31	<i>Artemisia tridentata</i>	<i>Comandra umbellata</i>	Artr/Coum	Low elevation dry rocky slopes
	<i>Symphoricarpos oreophilus</i>	<i>Balsamorhiza sagittata</i>		
		<i>Elymus spicatus</i>		

Unit	Alliance	Association	Label	Habitat
32	<i>Salix boothii</i>	<i>Salix wolfii</i>	Sabo/Sawo	High elevation wetlands
	<i>Salix drummondiana</i>	<i>Polygonum bistortoides</i>		
	<i>Carex rostrata</i>	<i>Pedicularis groenlandica</i>		
	<i>Carex nebrascensis</i>	<i>Juncus ensifolius</i>		
	<i>Juncus balticus</i>	<i>Saxifraga odontoloma</i>		
33		<i>Betula occidentalis</i>	Sabo/Beoc	Low elevation riparian
		<i>Salix lutea</i>		
		<i>Salix lasiandra</i>		
		<i>Salix exigua</i>		
		<i>Cornus sericea</i>		
34		<i>Ranunculus macounii</i>	Sabo/Rama	High elevation riparian
		<i>Equisetum arvense</i>		
		<i>Arnica chamissonis</i>		
		<i>Geum macrophyllum</i>		
		<i>Veratrum californicum</i>		

* Introduced species

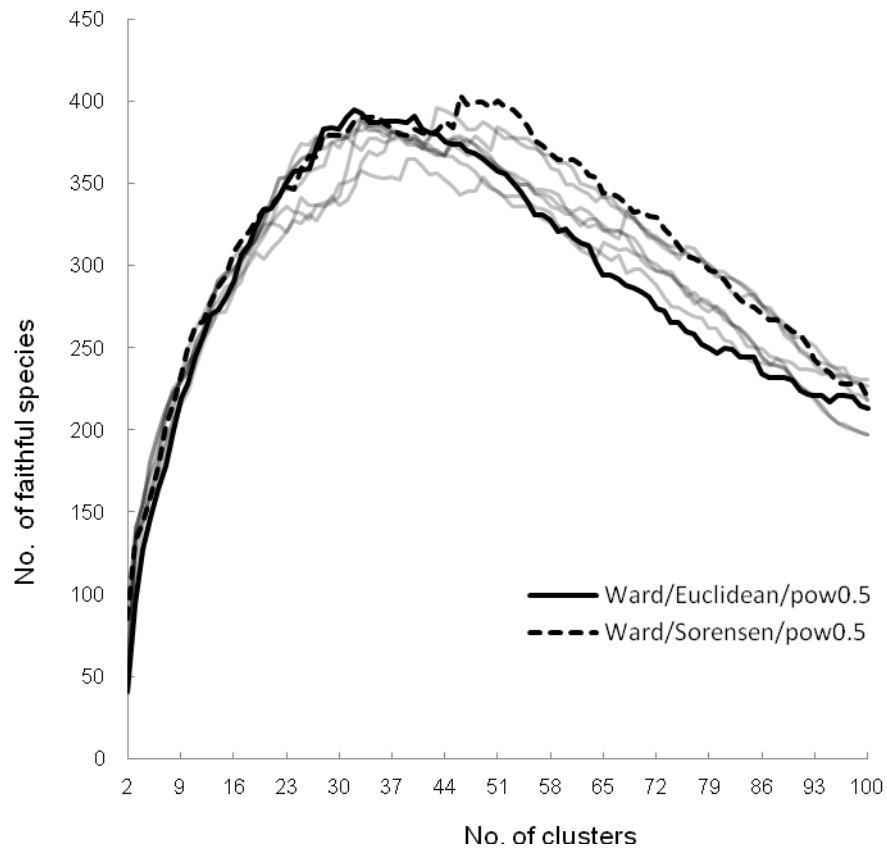


Figure 4.1. Results of the OptimClass method, Fisher's exact test with $p < 0.01$. Each of nine curves represents one solution based on partitioning method, distance measure and species cover transformation for number of clusters/vegetation units from 2 to 100 (on the horizontal axis). The vertical axis represents the total number of faithful species in all clusters for the given partitioning. Both bold and dashed curves are two best solutions in terms of the highest number of faithful species. The dashed curve represents statistically the best (403 faithful species in 48 clusters) but worst interpretable solution. The bold curve represents the most feasible solution with the highest number of faithful species (395) in 32 relatively large and interpretable clusters before final partitioning.

CHAPTER 5
COMPREHENSIVE ECOSYSTEM CLASSIFICATION IN THE ROCKY
MOUNTAINS, NORTHERN UTAH⁴

Abstract

Fundamentals of the direct gradient analysis, hierarchical organization of terrestrial ecosystems together with the biogeoclimatic classification approach used in British Columbia were used to develop a comprehensive ecosystem classification in a mountainous study area in the northern Utah.

This classification was derived from sampling of both forest (spruce-fir, Douglas-fir, aspen, juniper and mahogany woodland) and non-forested (willow-riparian, shrublands, tall-forb meadows) ecosystems. One-hundred-sixty-three sample plots were described by physiographic features and soil properties such as nutrient pools and dynamics.

Principal component analysis revealed dominant environmental gradients affecting vegetation in the subalpine vegetation geo-climatic zone. We discriminated sample plots using cluster analysis, CART and RandomForests classification for each gradient, thus constructing an ecological grid and specifying site classes as the basis for the site classification. We coupled this site classification with existing vegetation classification, and allocated plant communities into an environmental space represented by the site grid. Within a broad climatic framework represented by the subalpine vegetation geo-climatic zone, this overlay resulted in a comprehensive ecosystem classification, allowing us to discriminate ecosystems.

⁴ Coauthored by Antonin Kusbach, James N. Long, and Helga Van Miegroet

This classification enables a direct assessment of ecosystems' environmental properties and vegetation potential; and may allow us to infer site disturbance history and successional status of vegetation.

Introduction

“Show me your classification system, and I will tell you how far you are in elaborating the problem” - W. Kubiena

Early land classifications in the U.S., e.g., the habitat type concept (Daubenmire 1952), potential natural vegetation (Küchler 1969), and ecoregion classification (Bailey 1998), have been based on vegetation and macroclimate (Major 1951). Whereas habitat types rely completely on vegetation (Pfister 1976), other approaches for classifying terrestrial ecosystems, such as ecoregions, potential natural vegetation groups or biogeoclimatic zones (Krajina 1965), use combinations of vegetation and macroclimate based on the concept that broad macroclimate is reflected by similar vegetation over vast areas. However, vegetation is not driven only by macroclimate.

Following disturbances, ecosystems display a mosaic of plant communities in different successional stages. Together with animals, microbes and physical environment (hereafter referred to as a site), vegetation is a component of ecosystems irrespective of their size (Fosberg 1967, Pojar et al. 1987, Jennings et al. 2008). Thus, vegetation reflects the physical environment, disturbance history and biotic interactions (e.g., mycorrhizae and allelopathy) and is dynamic. In contrast, the physical environment represents a relatively stable framework. Therefore, **comprehensive ecosystem classification** should include both vegetation and physical environment components.

Our general goal is to better understand the relationship between vegetation and physical environment in the study area representing the Rocky Mountains of northern Utah. To create a comprehensive ecosystem classification, we set out to: (1) construct a **site classification**, which organizes ecosystems, independent of vegetation strictly based on environmental properties; (2) couple this site classification with a **vegetation classification** which organizes plants according to their fidelity to a particular location (Chapter 4) to create a comprehensive ecosystem classification within the context of a broad climatic zonation or **zonal (climatic) classification** (Meidinger and Pojar 1991, Chapter 3).

Methods

Study area

Franklin Basin and the T.W. Daniel Experimental Forest make up the study area (ca 16,000 ha, and ca 1,000 m of vertical extent) (Fig. 2.2). Franklin Basin (FB, 15,000 ha) is a montane-subalpine area situated between the Bear River Range and the Wasatch Range in the central Rocky Mountains on the Utah and Idaho border. The T.W. Daniel Experimental Forest (TWDEF, ca 1000 ha) is situated on a high ridge plateau of the Wasatch Range (10 km to the southeast of FB).

According to Bailey (1998) and McNab et al. (2007), the study area occurs within M331 Southern Rocky Mountains Steppe-Open Woodland-Coniferous Forest-Alpine Meadow Province, “D” Overthrust Mountain Section, “n” Northern Wasatch Range, and “o” Bear River Front Range Subsections. The mean annual precipitation ranges from about 720 to 1250 mm and mean annual air temperature ranges from 2.4 to 5.7 °C for

Temple Fork, Tony Grove Lake, Franklin Basin, and Utah State University (USU) Doc Daniel weather stations (<http://www.wcc.nrcs.usda.gov/snow/>).

The terrain is mountainous, rocky and steep with occasional flat to gently sloping high ridge-plateaus and benches. The elevation ranges from 1590-3060 m across the three study sites. The highest area of the Bear River Range was glaciated during the Pleistocene as manifested by glacial geomorphologic features like moraines, U-shaped valleys, erratics, and irregular glacial deposits (Atwood 1909, Young 1939, Degraff 1976). The study area is mostly built from calcareous sedimentary rocks (limestone, dolomite) with interlayered quartzite, and from Tertiary sediments consisting of grit, conglomerate, and siltstone of Wasatch Formation at the TWDEF site. The soils are formed in residuum, colluvium, alluvium, glacial till and outwash, and occur on diverse landforms such as cliffs, talus slope, moraines, karst valleys, mountain slopes, landslides, plains, valleys, depressions, ravines, and wetlands (Schoeneberger et al. 2002).

Over half of the study area is occupied by forest ecosystems including Engelmann spruce (*Picea engelmannii*), subalpine fir (*Abies lasiocarpa*), Douglas-fir (*Pseudotsuga menziesii*), aspen (*Populus tremuloides*), and woodland ecosystems including mountain mahogany (*Cercocarpus ledifolius*) and Rocky Mountain juniper (*Juniperus scopulorum*). Substantial changes in fire regimes, often in combination with cutting and grazing, have led to dramatic changes in the structure and the age-class distribution of forest stands. In many places, 100- to 140-year-old stands are now predominant (Long 1994). Forests in the study area are thus characterized by mid- and late-seral stages where forest understory is usually well developed (Pfister and Arno 1980). Non-forested ecosystems include willow-riparian habitats (*Salix spp.*) and wetlands, low shrublands

(*Artemisia spp.*), tall-forb meadows, and sparse vegetation on talus and rock outcrops, which may represent stable or temporary communities. Despite human impacts in last 120 years, the study area is considered as relatively natural in terms of plant species composition (Bird 1964).

Data collection

Sampling was intensive enough to capture as much ecosystem variation as possible focusing on all major existing plant communities (Brohman and Bryant 2005) occupying all major landforms (Schoeneberger et al. 2002), but avoiding ecotones and recently disturbed (burned, logged, damaged by insects) areas. For a site classification, we selected 136 sample plots within the subalpine zone (Chapter 3) from the entire dataset of 163 plots established in the summers of 2006 and 2007 in the study area (Chapter 2). A stratified (based on vegetation physiognomy) fixed (subjective selection) sampling design was used with circular plot size of 1000 m² for forest and 100 m² for non-forested ecosystems (Brohman and Bryant 2005). Three replicate plots were considered the minimum for defining a vegetation unit. We anticipated 50-60 units based on a preliminary reconnaissance of the study area.

We described each sample plot by environmental variables such as relatively static or constant attributes i.e., physiographic variables (slope aspect, slope gradient, topographic position and slope shape (Lotspeich 1980); dynamic attributes such as O and A horizon thickness, humus form, pH, nutrient pools, attributes describing relatively slow processes; and attributes such as nutrient supply rates describing relatively fast processes (Table 2.1). One soil pit was dug in each plot to the unweathered parent material and described

using the National Cooperative Soil Survey protocols (Soil Survey Staff 1999, 2006, Schoeneberger et al. 2002). Humus form was identified following Green et al. (1993).

More detailed description of the various site characterization methods was provided in Chapter 2. Briefly, one composite soil sample from 0-30 cm was collected from a pedon face in each plot, air dried and sieved (< 2 mm), and the fine fraction analyzed for texture classes (sandy, loamy, clayey) using the feel-method (Thien 1979). Samples were then analyzed for pH (1:1 soil in water, Corning pH analyzer) and total C and N (LECO CN analyzer, Leco Corp., St. Joseph, MI). Exchangeable cations using a mechanical vacuum extractor (Holmgren et al. 1977), followed by extractant analysis on inductively-coupled plasma spectrophotometer (ICP) (Iris Advantage, Thermo Electron, Madison, WI); extractable P [the Olsen P method (Olsen et al. 1954), sodium bicarbonate extraction, Thermo Electron Spectronic 20 Genesys spectrophotometer]; and mineralizable N [7-day anaerobic incubation and extraction (Keeney and Bremner 1966), NH_4 analysis (Lachat Quickchem 8000 Flow Injection Analyzer)] were determined as a static-absolute nutrient availability index (SNAI).

To determine a dynamic-relative nutrient availability index (DNAI) (Qian and Schoenau 2002), plant root simulators (PRSTM-probes; Western Ag Innovations, Inc., Saskatoon, Canada), a combination of anion and cation exchange membranes, were buried vertically into the mineral soil at each site for six weeks (during September and November). PRSTM-probes were cleaned and sent to Western Ag Innovations for extraction and chemical analysis including Ca, Mg, K, S, Fe, Mn, Zn, Cu, Pb, Al, NH_4 cations, and NO_3 and PO_4 anions (Table 2.1).

Data analysis

The dataset represented 136 sample plots and 42 environmental variables. We performed several analytical steps in order to discriminate sites that distinguish plant communities and reveal vegetation-site relationships. These included: (1) direct gradient analysis represented by ordination of environmental data; (2) cluster analysis of the important environmental variables; (3) discriminant analysis represented by Classification and Regression Trees (CART) and RandomForests of these environmental variables; (4) regressions among soil properties; and (5) overlay site classification with the previous vegetation classification (Chapter 4).

We used Principal components analysis (PCA) ordination (Pearson 1901) to determine the relative importance of the environmental factors to each principal component (PC) and interpret PCs as environmental gradients associated with the sample plots. Orthogonal rotations and correlation type of a cross-products matrix were used to get independent, mutually uncorrelated PCs (Lattin et al. 2003). Significance of PCs was tested by a Monte Carlo randomization test (based on proportion-based p -values for each PC). In order to document the relationship of the variables with the PCs and interpret PCs, we calculated correlation coefficients (loadings) with each ordination axis, and the linear (parametric Pearson's r) and rank (nonparametric Kendall's τ) relationships between the ordination scores and the observed variables. Our use of r and τ is suggested to be, even in relatively small datasets, more conservative than p -values for the null hypothesis of no relationship between ordination scores and variables (McCune and Grace 2002). We set the threshold for r and $\tau > 0.35$. For variables conversion and transformation see Chapter 2.

To group the sample plots according to the important environmental factors obtained in the PCA, we ran a hierarchical cluster analysis for the environmental gradients; we used Ward's linkage method with compatible Euclidean distance matrix. We transformed the variables with $|\text{skewness}| > 1$ to be close to multivariate normality and standardized the data by adjustment to standard deviate (z-scores). We checked the dataset for outliers given cutoff of 2.0 standard deviations from the grand mean (McCune and Mefford 2006). Cluster dendrograms were scaled by a distance objective function (Wishart 1969) and resulting height rescaled for information remaining. The decision as to how many clusters to keep was based on a compromise between minimizing the number of clusters and maximizing of information retained i.e., on cluster stability (McCune and Grace 2002, Lattin et al. 2003). The best meaningful solution was verified by pseudo F function (Calinski and Harabasz 1974).

The cluster analysis grouped the plots by important environmental factors. Despite the grouping, we still did not know in what way the clusters differed. In order to detect cluster differences, we performed CART classification (Breiman et al. 1984) using “tree” function, and RandomForests (Breiman 2001) classification using “RandomForest” function in the R software. The CART classification was validated by 10-fold cross-validation and a classification tree was pruned to avoid over-fitting; this means we pruned the tree at a point where its additional growth did not bring any improvement. RandomForests classification verified the CART results. These two classifications highlighted: (1) the most important environmental factors associated with sites clustering along each gradient; and (2) possible misclassifications of sites. In addition, the CART classification provided factor's split thresholds that were used for labeling of relative **site**

classes. Based on this site classification, we constructed a **site grid** similar to an edatopic grid (Pogrebnyak 1930, Rysin 1982) for the subalpine vegetation geo-climatic zone.

We performed simple regressions of important nutrient concentrations and supply rates (N, P, K, Ca, Mg, Fe) with important soil properties suggested by the CART and RandomForests classification.

We overlaid the site classification represented by site classes in the site grid and the vegetation classification (Chapter 4) represented by plant associations within the subalpine zone (Chapter 3) to detect the relationship between environmental properties and existing vegetation. Each plant association was represented by minimum of three sample plots.

We used R software, ver. 2. 7. 2., (<http://www.r-project.org/>), and PC-ORD 5 (McCune and Mefford 2006) in the analysis.

Results

Environmental ordination

Principal component analysis (PCA) on the environmental data revealed four ecologically meaningful principal components (PC 1-4) (Table 5.1) that were interpreted as a soil nutrition, topography-moisture, microbial activity and soil development based on loadings (Table 5.2), and synthesized as soil fertility (PC 1, 3, 4), and topography-moisture (PC 2) gradients (Chapter 2). Based on PCs loadings, we identified eighteen important nutrient factors associated with a soil fertility gradient e.g., soil pH, CaCO₃ content, and nutrient SNAs and DNAs; and seven important topography factors representing a topography-moisture gradient such as slope position, slope gradient,

terrain shape, soil depth, coarse rock fragment content, water table and mottles (Table 5.2). PCs were interpreted and synthesized in the same way as in the Chapter 2. Loadings were somewhat different from those reported in Table 2.3 because in current analysis only data set for the subalpine vegetation geo-climatic zone was used.

Cluster analysis

The important environmental indicators for soil fertility and topography-moisture gradient were used in cluster analysis to identify environmentally similar sites and assemble them into homogeneous clusters. This means that members of a cluster have similar soil fertility or topography-moisture properties but differ significantly from members of other clusters in those environmental parameters. Three-cluster and two-cluster solutions were statistically best for the topography-moisture and the fertility gradient, respectively, as indicated by amount information remaining and Pseudo F function. However, for meaningful ecological interpretation, the two-cluster solution of the fertility gradient was too general; therefore, we considered a similar three-cluster solution with broader interpretability also for the fertility gradient (Fig. 5.1a, b).

Discriminant analysis

Three-cluster solutions for both gradients were analyzed by CART and RandomForests. For the soil fertility gradient, CART performed a pruned classification tree with three terminal nodes based on splits of Al concentration and CaCO_3 content (Fig. 5.2a). Because CaCO_3 content represents soil alkalinity and Al concentration soil acidity, the terminal node 1 ($\text{Al}_s \geq 2.985$ mg/kg soil) represents acidic sites, the terminal node 3 ($\text{CaCO}_3 \geq 0.35$ %) represents calcareous sites, and the terminal node 2 ($\text{CaCO}_3 <$

0.35 %; $Al_s < 2.985$ mg/kg soil) represents neutral or alkaline sites. The misclassification error rate was 7.4 % in the CART. RandomForests identified Al concentration and $CaCO_3$ content as environmental variables most strongly associated with the three-cluster solution of the fertility gradient (the first two variables in Fig. 5.2b). RandomForests was run with the number of trees (ntree) = 1000 and four variables used in each split (mtry = 4). Misclassification error rate was 6.6 %. CART and RandomForests results were consistent. Based on soil pH (the fourth variable in Fig. 5.2b) in combination with Al and $CaCO_3$ as additional parameters, the fertility gradient was expressed by three classes: (1) acidic; (2) neutral-alkaline; and (3) calcareous (Table 5.3a).

For the topography-moisture gradient, CART performed a pruned classification tree with three terminal nodes based on splits of soil depth and water table (Fig. 5.3a). The terminal node 1 (soil depth < 72.5 cm) represents sites with shallow soil, the terminal node 2 (soil depth > 72.5 cm; water table class > 3.5) represents sites with deep soils and no water table, and the terminal node 3 (soil depth > 72.5 cm; water table class < 3.5) represents sites with deep soils and water table present. CART calculated the misclassification error rate of 4.4 %. RandomForests calculated soil depth and water table as the variables most strongly associated with the three-cluster solution of the of the topography-moisture gradient (Fig. 5.3b). RandomForests was run with the number of trees (ntree) = 1000 and two variables used in each split (mtry = 2). Misclassification error rate was 1.5 %. CART and RandomForests results were consistent.

Three terminal nodes in this classification discriminated three topography-moisture classes, which were interpreted as: (1) convex/dry; (2) linear-concave/mesic; and (3)

concave/wet (Table 5.3b). However, three-cluster classification for the topography-moisture gradient was too general and not sufficient to characterize the broad environmental variability of the gradient e.g., in the wet class, height of a water table is a fundamental driver of vegetation distribution. Therefore, each of those three clusters representing three preliminary topography-moisture classes were clustered and analyzed separately by CART and RandomForests to get more detailed structuring for the topography-moisture gradient.

In a repeated clustering (using the same Ward/Euclidean method), we assessed three-cluster solution as the best for the convex/dry and the concave/wet class, and two-cluster solution for the liner-concave/mesic class as the best as indicated by amount information remaining and Pseudo F function (Fig. 5.4a, b, c). In all instances, CART and RandomForests yielded consistent results. For the convex/dry class, CART performed a pruned classification tree with three terminal nodes based on coarse rock fragment content (RF) and slope gradient. The terminal node 1 ($RF > 72.5 \%$; $slope < 51.5 \%$) represents rocky, skeletal sites on backslopes, the terminal node 2 ($RF > 72.5 \%$; $slope > 51.5 \%$) represents steep rocky sites and cliffs, and the terminal node 3 ($RF < 72.5 \%$) represents skeletal crests, shoulders and backslopes (Fig. 5.5a). CART calculated the misclassification error rate of 10.6 %. RandomForests also identified slope gradient and rock fragment content as the variables most strongly associated with the three-cluster solution of the convex/dry class (Fig. 5.5a). RandomForests was run with the number of trees ($ntree = 1000$) and two variables used in each split ($mtry = 2$). Misclassification error rate was 6.1 %.

For the concave/wet class, CART performed a pruned classification tree with three terminal nodes based on slope position (topos) and water table. The terminal node 1 (topos > 3.5; water table > 60.5 cm) represents moist flats, the terminal node 2 (topos > 3.5; water table < 60.5 cm) represents wet flats and toeslopes, and the terminal node 3 (topos < 3.5) represents very wet concave footslopes (Fig. 5.5b). CART calculated the misclassification error rate of 6.3 %. RandomForests (ntree = 500, mtry = 2, error rate 6.3 %) also identified slope position and water table as most strongly associated with the two-cluster solution of the concave/wet class (Fig. 5.5b).

For the linear-concave/mesic class, CART performed a pruned classification tree with two terminal nodes based on variable mottles. The terminal node 1 (mottles > 3.5 meaning mottles are present) represents fresh, mostly concave toeslopes and depressions, and the terminal node 2 (mottles < 3.5 meaning no mottles) represents linear slopes (Fig. 5.5c). CART calculated the misclassification error rate of 2.1 %. RandomForests (ntree = 1000, mtry = 4, error rate 2.1 %) also identified mottles as most strongly associated with the two-cluster solution in this class (Fig. 5.5c).

As a result, the topography-moisture gradient was finally classified into eight classes: (1) extremely dry; (2) very dry; (3) dry; (4) slightly dry; (5) fresh; (6) moist; (7) wet; and (8) very wet (Table 5.3b).

Regressions of soil acidity and nutrient availability

Across the sampling plots, there was a significant negative association between soil pH and essential nutrient supply rates (N, P, K, Fe), i.e., at high pH, there were fewer nutrients available in a soil solution (Fig. 5.6). There was a significant positive

correlation of Ca and Mg supply rates with soil pH, i.e., at high pH, there were enough Ca and Mg available in soil solution.

Regression results indicated lower plant nutrient supply of important macronutrients such as N, P, and K at high pH. This suggests that highly alkaline or calcareous soils are less fertile regardless of high supply of secondary macronutrients such as Ca and Mg.

Overlay of the site and vegetation classification

In a previous vegetation classification scheme, plant species were organized into plant alliances and associations (Table 5.4) (Chapter 4). Because each sample plot was associated with a specific site designation (based on topography-moisture and soil fertility) the plant associations could be organized into the site grid for the subalpine vegetation geo-climatic zone. This site-vegetation grid was constructed separately for forest (spruce-fir, Douglas-fir, mountain mahogany and aspen) and non-forested (sagebrush, tall-forb meadow and willow) ecosystems (Fig. 5.7). In this overlay, each plant association was characterized by **site quality**, a specific location in the grid identified by a particular combination of the soil fertility and topography-moisture within the environmental space. Once connected with site properties, these plant associations then represent an ecosystem, i.e., a relatively stable physical environment upon which the specific plant community grows (e.g., a riparian forest) (Major 1951, Pojar et al. 1987). Practically, we may recognize two or more different plant associations on a site with the same stable physical environment (these associations share a common space in the site grid). From this we may infer contrasting vegetation successional trajectories within the same relatively stable environmental context (D. Roberts 2009, personal communication).

Based on site quality, plant communities can be compared each other. While some associations are clearly detached from the others (e.g., 11, 12 mountain mahogany woodland in Table 5.4, Fig. 5.7b), others share the environmental space partially or entirely (e.g., 16-20 aspen in Table 5.4, Fig. 5.7d). This means that some communities (e.g., mountain mahogany on dry, calcareous, and rocky sites) occur in unique environmental settings, whereas others demonstrate a broad ecological amplitude (e.g., aspen).

Discussion

Site classes

A two-step consecutive differentiation within the topography-moisture gradient revealed eight meaningful topographical-moisture classes, while a simpler division did not explain ecosystem distribution along the gradient adequately. Local topography is a fundamental predictor of physical soil properties (especially soil depth, rock fragment content and slope gradient) that affect soil water status, including drainage (Schoeneberger et al. 2002).

Aridity of the classes 1 through 3 is reflected by high rock fragment content, high slope position, steep slope gradient and shallow soils. Soils within these classes remain dry for a long period of the year especially in summer. Mesic conditions of the classes 4 through 5 are expressed by greater soil depth and by the presence of mottles, which indicate water stagnation in a soil profile for some time during the year. These soils may remain moist for a substantial period of the year especially after spring runoff. Wetness of the classes 6 through 8 is expressed by depth to the water table fluctuating in a soil

profile throughout a year. With a perched water table, soils of these classes remain wet for the entire year.

The wet classes (6 through 8) may be characterized by surplus of soil water throughout the year while the dry classes (1 through 3) experience substantial water deficits. The mesic classes (4 through 5) could be considered more or less balanced in terms of soil water supply depending on climatic inputs (amount of precipitation as a snowpack in winter wet season).

Soil pH emerged as the strongest indicator of the fertility gradient. It is strongly correlated with the plant nutrient availability (Fig. 5.6) (e.g., Lindsay 1976, Kotar 1988, Brady and Weil 1999). Relatively acid soils in class 1 are associated with presence of quartzite or its weathered remnants (rich in silica) especially in glacial deposits (till). Such soils with high amount of Al and H (acidic) cations, and little or no CaCO_3 are subject to greater leaching of important nutrients such as base cations (K, Ca, Mg) especially in a humid mountainous climate. This results in low plant nutrient availability and low soil fertility.

Because of the generally alkaline status of soils derived from alkaline limestone and dolomite rocks, soils in class 2 have high buffering capacity, i.e., high ability to resist pH change. Soil fertility is high as important base nutrients are retained on the soil exchange complex (high base saturation) and other macronutrients (N, P, and S) are relatively available (Brady and Weil 1999, Montana State University 2005).

In dry, calcareous environment of soils in class 3, all important macronutrients (primary macronutrients N, P, K) are less available for plants (Lindsay 1976, Brady and Weil 1999). The supply rates of NH_4 , P, K and Fe are low at high pH (Fig. 5.6); K is

fixed in minerals (mainly clays) in dry conditions; similarly Ca is precipitated as CaCO_3 ; and important micronutrients (metals) are fixed in minerals or bound tightly to the soil (Brady and Weil 1999, Montana State University 2005). As a consequence, calcareous sites can experience important nutrient deficiencies and appear to be less productive due to aridity, i.e., soil water is the strongest driver of plant nutrient availability (Lindsay 1976, Brady and Weil 1999, Montana State University 2005). Note that all cells for this calcareous class and classes 4-8 of the topography-moisture gradient are empty indicating that there are no moist or wet calcareous sites (Fig. 5.7).

Allocation of plant communities

The association of plant communities (associations) and environmental factors represented by the site-vegetation grids (Fig. 5.7) provides considerable insight into the influence of environmental gradients on the distribution of vegetation, likely successional patterns (ecosystem dynamics), and even particular misclassifications of plant communities that might have happened in vegetation classification (Chapter 4).

A great advantage of the site classification is that plant communities are mutually comparable inside the site grid. Sites with the same quality have the same vegetation potential (Cajander 1926, Bakuzis 1969), that is, are capable of supporting the same climax vegetation. The presence of different existing plant communities indicates that they likely experienced different succession or different disturbance history (e.g., Pojar et al. 1987, McCune and Grace 2002) (consistent with the concept of alternate states in rangeland e.g., NRCS Ecological Site Description, Ripplinger 2010). Thus, representation of plant communities within the site grid allows assessment of

environmental-disturbance relationships and ecosystem dynamics (Pojar et al. 1987, Meidinger and Pojar 1991, D. Roberts 2009, personal communication).

In the site grid, some plant communities seem to be detached from the others while others share the environmental space partially or entirely. For example, the 12-Cele/Pepu mountain mahogany is the only community occurring on extremely dry calcareous sites (Fig. 5.7b). This unique environmental setting is thus a good predictor of mountain mahogany climax vegetation. In contrast, the 5-Abla-Pien/Lipo and 6-Abla-Pien/Osch spruce-fir communities (Table 5.4) share a similar space in the site grid (Fig. 5.7a). Similarly, the Douglas-fir 14-Psme-Abla-Pifl/Linu and 15-Psme-Pifl/Aste penetrate (Fig. 5.7c). Spruce-fir communities 5 and 6 were classified differently because of different age, yet were considered as one ecosystem (mild, undulating slope forest) growing in a similar environment (Table 5.4). In the case of Douglas-fir communities, different classes in an apparently similar environment may be explained either by different elevations of Douglas-fir communities which cannot be displayed in the site grid (the site grid is constructed for one vegetation geo-climatic zone only, Chapter 3) or by succession, i.e., on the basis of different age of the Douglas-fir stands. Here again, these communities were considered as one ecosystem (subalpine/high elevation sunny-slope forest) (Table 5.4).

Aspen communities cover a large space inside the grid reflecting its exceptional ecological plasticity and amplitude (e.g., Mueggler 1988, Klinka et al. 1999, Chapter 3) and high genetic variability (Mock et al. 2008). Aspen does not occur on very wet sites (class 8) or extremely dry sites (class 1). It is most prevalent in the center of the site grid (in intermediate environmental condition – topo-moisture class 3, 4 and fertility class 2)

where four associations share the similar space (16-19 in Table 5.4, Fig. 5.7d). There is essentially low environmental difference among those communities that is consistent with their somewhat weak vegetation classification (Chapter 4). On the other hand, despite some similarity of relatively poor (16-Potr-Abla-Pien/Pole, blue color in Fig. 5.7d) and alkaline/calcareous (19-Potr-Abla/Scla; yellow) aspen with others, these two communities together with wet aspen (20-Potr/Cyof, orange) indicate that reasonable environmental difference among the extremes exists within aspen communities. These environmental extremes are more obvious from continuous ordination space based on PCA (Fig. 5.8) rather than from categorical classes of a site grid (Fig. 5.7d). The wet aspen community (20-Potr/Cyof) is the only true separate group (no common space with any of others) with clearly different environment that might be considered potential stable riparian broad-leaved forest ecosystem (Table 5.4). Conifer encroachment into aspen communities (16-19-Potr-abla associations) is indicative of successional transition. Based on the site grid, conifer encroachment does not seem to be associated with specific environmental site properties. Absence of conifers in moist and wet aspen community (20-Potr/Cyof) seems to indicate that this community may be more resistant to conifer encroachment than the others.

The display of plant communities within the site grid can also reveal potential mistakes in vegetation classification. For example, the 7-Abla-Pien-Psme/Rupa spruce-fir association occurs in two clearly detached places in the grid. Environmental difference of two site classes between these two places (from slightly dry to extremely dry) (Fig. 5.7a) is so large that they cannot be considered the same plant association. There is an obvious

misclassification of the spruce-fir association (7); this association should be reclassified and divided into two separate communities.

For non-forested plant communities, there is a clear environmental difference between two major sagebrush communities: *Artemisia spiciformis* (Shultz 2009) (21, 22, 23) and *Artemisia tridentata* sub. *vaseyana* (31) (Fig. 5.7e). While the first occurs on more mesic sites, the second occurs on dry and alkaline-to-calcareous conditions. *Artemisia spiciformis* associations penetrate each other and also share a similar space with spruce-fir and aspen forest communities. This indicates either potential successional transition of the sagebrush communities towards conifer/aspen forest or local climatic differences such as cold air drainage in valley bottoms between relatively stable sagebrush communities and surrounding forest ecosystem (Table 5.4). This cold air drainage, where 21 and 22 - *Artemisia spiciformis* occur, may maintain sagebrush stability by preventing of forest establishment. The 31 - *Artemisia tridentata* association shares a similar environmental space with juniper and mountain mahogany associations suggesting transitional possibilities of sagebrush steppe towards these woodland types (Fig. 5.7b, e).

Potential value of a comprehensive classification

The comprehensive ecosystem classification resulting from the joining of the site and vegetation classifications demonstrated a considerable influence of the physical environment (climate and site) on vegetation distribution. The allocation of plant communities inside the site grid appears to be a powerful tool for ecosystem environmental assessment. Because vegetation distribution is influenced not only by

environmental factors (site properties), but also by disturbance history, knowledge of vegetation environmental demands can help us to infer vegetation dynamics and succession.

The comprehensive ecosystem classification as demonstrated for the subalpine vegetation geo-climatic zone (Chapter 3) may be used as a means of communication in ecosystem research, outreach and education, as well as a tool for practical interpretations and applications in ecosystem management (Pojar et al. 1987, Kotar 1988, Sharik 2010). For example, based on a particular climate change scenario and knowledge of current site quality of ecosystems we can infer potential change in site properties (e.g., soil chemistry microclimate), from which we can predict possible ecosystem changes and transition. Additionally, knowledge about the site quality may contribute to ecosystem planning, conservation or utilization.

Summary and conclusions

Our site classification was based on the most important environmental factors and gradients within the subalpine vegetation geo-climatic zone. These gradients were topography-moisture and fertility. We specified important site classes associated with each gradient and constructed a site grid. We overlaid the existing vegetation classification onto the site classification, resulting in allocation of plant associations into the site grid. This comprehensive classification integrated zonal classification with both vegetation and environmental classification and enabled environmental specification of existing vegetation. Each plant community was defined by a particular site quality and

compared with other communities within the framework of the site grid (an environmental space of the subalpine vegetation geo-climatic zone).

This specification and comparison is important in understanding the relationship of vegetation with physical environment in the study area. Based on vegetation comparison, we were able to identify distinct ecosystems and infer possible ecosystem dynamics i.e., we could ascertain site disturbance history and successional status of vegetation.

Compared to the Natural Resources Conservation Service (NRCS) Ecological Site Description (ESD), which is soil-based and works largely for low-elevation rangelands, our classification assesses a broader range of important ecosystem components; it reflects relationships between climate, vegetation and physical site factors across a range of both forest and non-forested ecosystems. Our classification included all potential ecosystems in the study area; we assessed conifer and broad-leaved woodland, conifer and broad-leaved forest as well as non-forested ecosystems such as riparian wetland shrubland, dwarf-shrubland sagebrush steppe, tall-forb subalpine meadows, and sparse vegetation of highest mountain peaks.

The comprehensive ecosystem classification should serve as a general means of communication and therefore may be used as a valuable tool not only in ecosystem research but also in practical ecosystem management.

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Table 5.1. PCA summary with interpretation and synthesis of principal components. Significant principal components are indicated by *p* values, ecologically meaningful-interpretable PCs are in bold. NA-not applicable.

	PC1	PC2	PC3	PC4	PC5	PC6
Eigenvalue	8.34	7.41	4.99	2.54	2.21	1.59
% of Variance	19.85	17.66	11.89	6.04	5.26	3.78
Cumulative % of Var.	19.85	37.49	49.38	55.41	60.67	64.44
<i>p</i> - value	0.0002	0.0002	0.0002	0.0002	0.0002	0.99
Interpretation	Soil nutrition	Topo-moisture	Microb. activity	Soil development	NA	NA
Synthesis	Soil fertility	Topo-moisture	Soil fertility		NA	NA

Table 5.2. PCA loadings. Significant Pearson's (r), and Kendall's (τ) coefficients are in bold; both significant loadings express a significant variable for the particular PC (shaded). Variables are defined in Table 2.1.

PC Variable	PC1		PC2		PC3		PC4
	r	tau	r	tau	r	tau	r
topos	-0.15	-0.14	-0.81	-0.65	0.11	0.00	-0.01
sl	0.14	0.16	0.64	0.49	0.16	0.09	-0.01
av	-0.31	-0.23	-0.05	-0.05	0.37	0.25	0.00
shape	-0.16	-0.11	-0.72	-0.63	-0.08	-0.07	0.16
Ohor	0.02	0.01	-0.37	-0.25	0.76	0.43	0.06
Ahor	0.11	0.11	-0.27	-0.21	-0.62	-0.47	0.10
hum	0.06	0.03	0.29	0.21	-0.72	-0.45	-0.09
sdepth	-0.32	-0.20	-0.70	-0.55	-0.10	-0.08	0.18
RF	0.07	0.06	0.70	0.56	0.17	0.11	-0.24
parmat	0.43	0.28	-0.56	-0.40	-0.25	-0.19	0.14
wtable	-0.10	-0.07	0.68	0.45	-0.35	-0.23	0.19
mottles	-0.04	-0.03	0.74	0.52	-0.32	-0.20	0.12
cvalue	-0.74	-0.60	0.24	0.08	0.09	0.12	0.03
text	0.41	0.31	-0.48	-0.28	0.17	0.07	0.05
pH	0.85	0.66	0.23	0.16	-0.06	-0.04	0.05
CaCO ₃	0.73	0.59	0.42	0.38	0.11	0.10	0.02
Nmin_d	0.09	0.06	0.17	0.12	-0.65	-0.48	-0.34
Nox	0.60	0.43	-0.41	-0.23	-0.15	-0.25	-0.07
NO ₃ _d	0.10	0.07	0.17	0.12	-0.68	-0.52	-0.31
NH ₄ _d	-0.24	-0.14	-0.01	0.03	0.43	0.31	-0.15
Cox	0.76	0.62	-0.23	-0.04	0.10	-0.02	-0.03
C/N	0.31	0.17	0.29	0.20	0.47	0.38	0.07
Ca_d	0.48	0.32	-0.26	-0.16	-0.36	-0.24	-0.04
Mg_d	0.59	0.38	-0.02	-0.02	0.17	0.08	-0.13
K_d	-0.62	-0.43	0.07	-0.08	-0.41	-0.24	0.14
P_d	-0.38	-0.28	-0.28	-0.25	-0.49	-0.34	0.11
Fe_d	-0.33	-0.24	-0.60	-0.36	-0.05	-0.10	-0.31
Mn_d	-0.48	-0.37	-0.35	-0.20	0.18	0.07	-0.27
Zn_d	-0.07	-0.04	-0.36	-0.18	-0.44	-0.31	-0.29
S_d	-0.29	-0.23	-0.43	-0.21	0.16	0.06	-0.34
Al_d	-0.07	-0.04	0.19	0.12	-0.21	-0.13	-0.14
Ca_s	0.80	0.62	-0.35	-0.13	-0.03	-0.08	0.06

Mg_s	0.83	0.65	-0.17	-0.10	0.27	0.09	-0.01
K_s	0.07	0.05	-0.35	-0.28	-0.44	-0.31	0.25
NH4_s	0.07	0.01	-0.56	-0.35	0.10	0.02	-0.05
Nmin_s	0.66	0.48	-0.49	-0.22	0.00	-0.12	-0.09
P_s	-0.33	-0.22	-0.33	-0.29	-0.45	-0.35	0.11
Al_s	-0.73	-0.59	0.11	0.04	0.30	0.24	-0.03
Fe_s	-0.35	-0.28	-0.06	-0.05	0.44	0.35	-0.02
S_s	-0.02	0.04	-0.55	-0.35	-0.05	-0.03	-0.02
Mn_s	-0.65	-0.49	-0.34	-0.30	-0.12	-0.08	0.11
Zn_s	-0.55	-0.41	-0.02	-0.03	0.08	0.07	-0.03

Table 5.3. Site classification and site class specification: a) topography-moisture gradient; b) fertility gradient. Variables are defined in Table 2.1.

a

1 Acidic	2 Neutral-alkaline	3 Calcareous
pH 5.8 (5.7, 5.9) ^a	pH 6.5 (6.4, 6.6) ^a	pH 7.6 (7.5, 7.7) ^a
CaCO ₃ = 0 %	CaCO ₃ < 0.35 %	CaCO ₃ ≥ 0.35 %
Al ≥ 2.985 mg/kg soil	Al < 2.985 mg/kg soil	Al = 0 mg/kg soil

^a Mean (95 % confident limits), F = 217.2, p < 0.0001, N = 136

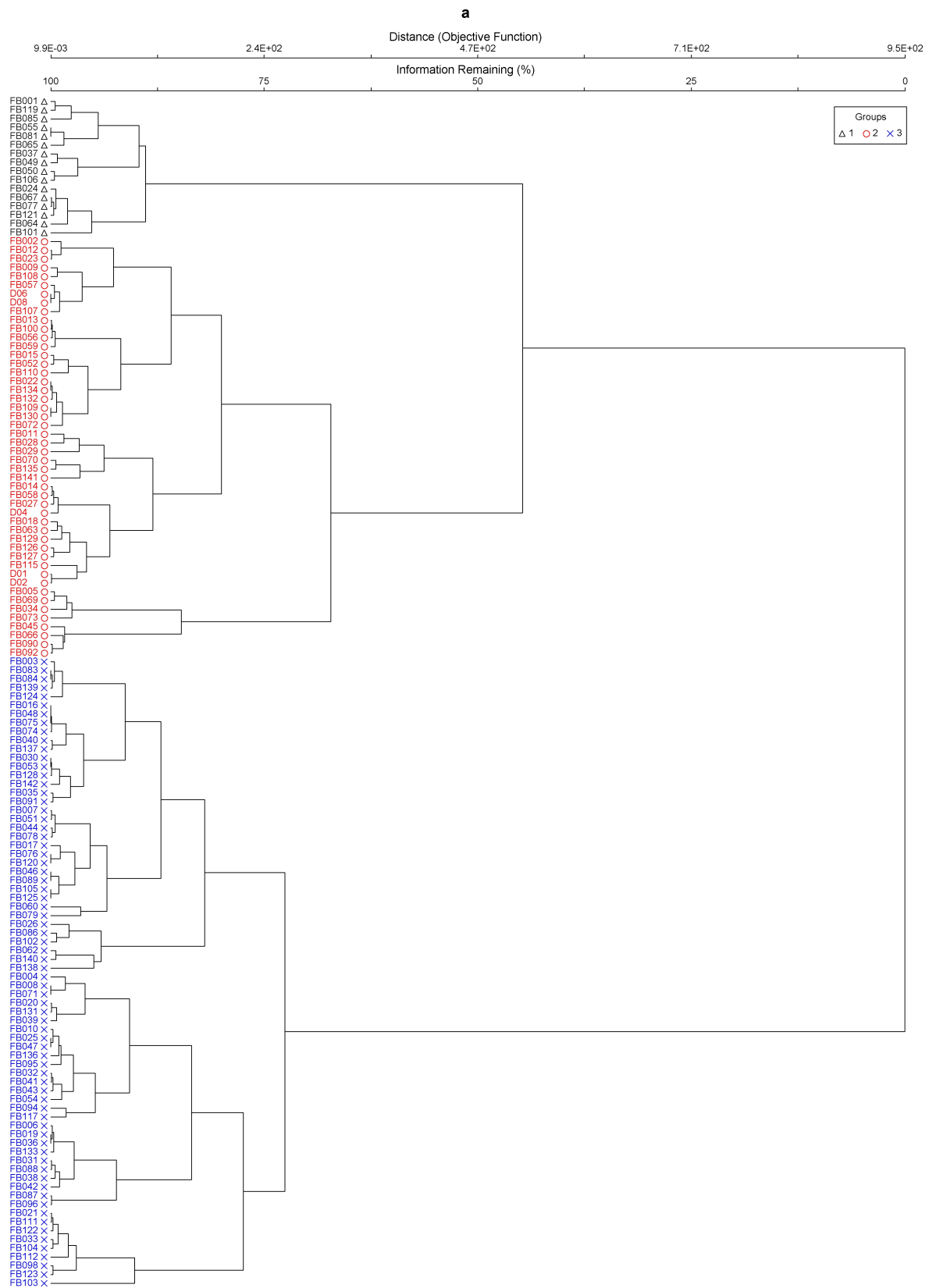
b

First run	1 Convex/dry			3 Linear-concave/mesic			2 Concave/wet	
Preliminary	Shallow soils < 72.5 cm			Deep soils ≥ 72.5 cm			Deep soils ≥ 72.5 cm	
classes	NO water table			NO water table			Water table present	
	Summits, shoulders, backslopes			Mottles may present > 30 cm			Mottles ≤ 30 cm	
				Flats, back-, foot-, toeslopes			Flats, toeslopes	
Repeated run	1 Extr. dry	2 Very dry	3 Dry	4 Slightly dry	5 Fresh	6 Moist	7 Wet	8 Very wet
	RF > 72.5 %	RF > 72.5 %	RF ≤ 72.5 %	NO mottles	Mottles present	Water table > 60.5 cm	Water table < 60.5 cm	Water table < 30 cm
	sl > 51.5 %	sl ≤ 51.5 %				Mottles ≤ 60 cm	Mottles ≤ 35 cm	
	sdepth < 40 cm	sdepth < 50 cm	sdepth < 70 cm	Linear	Concave	Linear	Linear -concave	Concave
	Steep slopes Cliffs	Backslopes	Crests Shoulders Backslopes	Backslopes Toeslopes Footslopes	Toeslopes Depressions	Flats	Flats Toeslopes	Footslopes

Table 5.4. Plant alliances and associations, and ecosystems.

No	Alliance	Association	Label	Ecosystem
1	<i>Abies lasiocarpa</i> <i>Picea engelmannii</i>	<i>Galium boreale</i>	Abla-Pien/Gabo	Riparian forest
2		<i>Zigadenus elegans</i>	Abla-Pien/Ziel	Wetland forest
3		<i>Thalictrum fendleri</i>	Abla-Pien/Thfe	High elevation, valley bottom open-canopy forest
4		<i>Lonicera utahensis</i>	Abla-Pien/Lout	Subalpine shady-slope forest
5	<i>Abies lasiocarpa</i> <i>Picea engelmannii</i> <i>Populus tremuloides</i>	<i>Ligusticum porteri</i>	Abla-Pien-Potr/Lipo	Mild, undulating slope forest
6		<i>Osmorhiza chilensis</i>	Abla-Pien-Potr/Osch	
7	<i>Abies lasiocarpa</i> <i>Picea engelmannii</i> <i>Pseudotsuga menziesii</i>	<i>Rubus parviflorus</i>	Abla-Pien- Psme/Rupa	High elevation skeletal shady-slope forest
8		<i>Rubus idaeus</i>	Abla-Pien- Psme/Ruid	Quartzite talus forest
10	<i>Juniperus scopulorum</i>	<i>Lomatium grayi</i>	Jusc/Logr	Very dry skeletal slope conifer woodland
11	<i>Cercocarpus ledifolius</i>	<i>Mertensia oblongifolia</i>	Cele/Meob	Dry slope broad-leaved woodland
12		<i>Petradoria pumila</i>	Cele/Pepu	Extremely dry, rocks/cliff broad-leaved woodland
13	<i>Pseudotsuga menziesii</i> <i>Abies lasiocarpa</i>	<i>Acer grandidentatum</i>	Psme-Abla/Acgr	Low elevation shady-slope forest

No	Alliance	Association	Label	Ecosystem
14	<i>Pseudotsuga menziesii</i> <i>Pinus flexilis</i> <i>Abies lasiocarpa</i>	<i>Linanthastrum nuttallii</i>	Psme-Abla-Pifl/Linu	Subalpine/high elevation sunny-slope forest (ribbons)
15	<i>Pseudotsuga menziesii</i> <i>Pinus flexilis</i>	<i>Astragalus tenellus</i>	Psme-Pifl/Aste	
16	<i>Populus tremuloides</i> <i>Abies lasiocarpa</i> <i>Picea engelmannii</i>	<i>Poa leptocoma</i>	Potr-Abla-Pien/Pole	
17	<i>Populus tremuloides</i> <i>Abies lasiocarpa</i>	<i>Valeriana occidentalis</i>	Potr-Abla/Vaoc	Mild, undulating slope forest
18		<i>Symphoricarpos oreophilus</i>	Potr-Abla/Syor	
19		<i>Scrophularia lanceolata</i>	Potr-Abla/Scla	
20	<i>Populus tremuloides</i>	<i>Cynoglossum officinale</i>	Potr/Cyof	Riparian broad-leaved forest
21	<i>Artemisia spiciformis</i>	<i>Chrysothamnus viscidiflorus</i>	Arsp/Chvi	Low elevation valley bottom/slope dwarf-shrubland (sagebrush steppe)
22		<i>Poa arnowiae</i>	Arsp/Poar	High elevation valley bottom/slope dwarf-shrubland (sagebrush steppe)
23		<i>Agastache urticifolia</i>	Arsp/Agur	
25	<i>Rudbeckia occidentalis</i>	<i>Ranunculus adoneus</i>	Ruoc/Raad	Subalpine depression/colluvial outwash meadow
26	<i>Agastache urticifolia</i>	<i>Aster integrifolius</i>	Agur/Asin	Subalpine flat/mild slope meadow
28	<i>Linum kingii</i>	<i>Hymenoxys acaulis</i>	Liki/Hyac	Alpine talus and rock sparse vegetation meadow
29	<i>Ivesia gordonii</i>	<i>Juncus parryi</i>	Ivgo/Jupa	Quartzite talus and rock sparse vegetation meadow
30		<i>Monardella odoratissima</i>	Ivgo/Mood	Subalpine talus and rock sparse vegetation meadow
31	<i>Artemisia tridentata</i>	<i>Comandra umbellata</i>	Artr/Coum	Low elevation dry rocky slope dwarf-shrubland (sagebrush steppe)
32	<i>Salix boothii</i>	<i>Salix wolfii</i>	Sabo/Sawo	High elevation wetland shrubland
33		<i>Betula occidentalis</i>	Sabo/Beoc	Low elevation riparian shrubland
34		<i>Ranunculus macounii</i>	Sabo/Rama	High elevation riparian shrubland



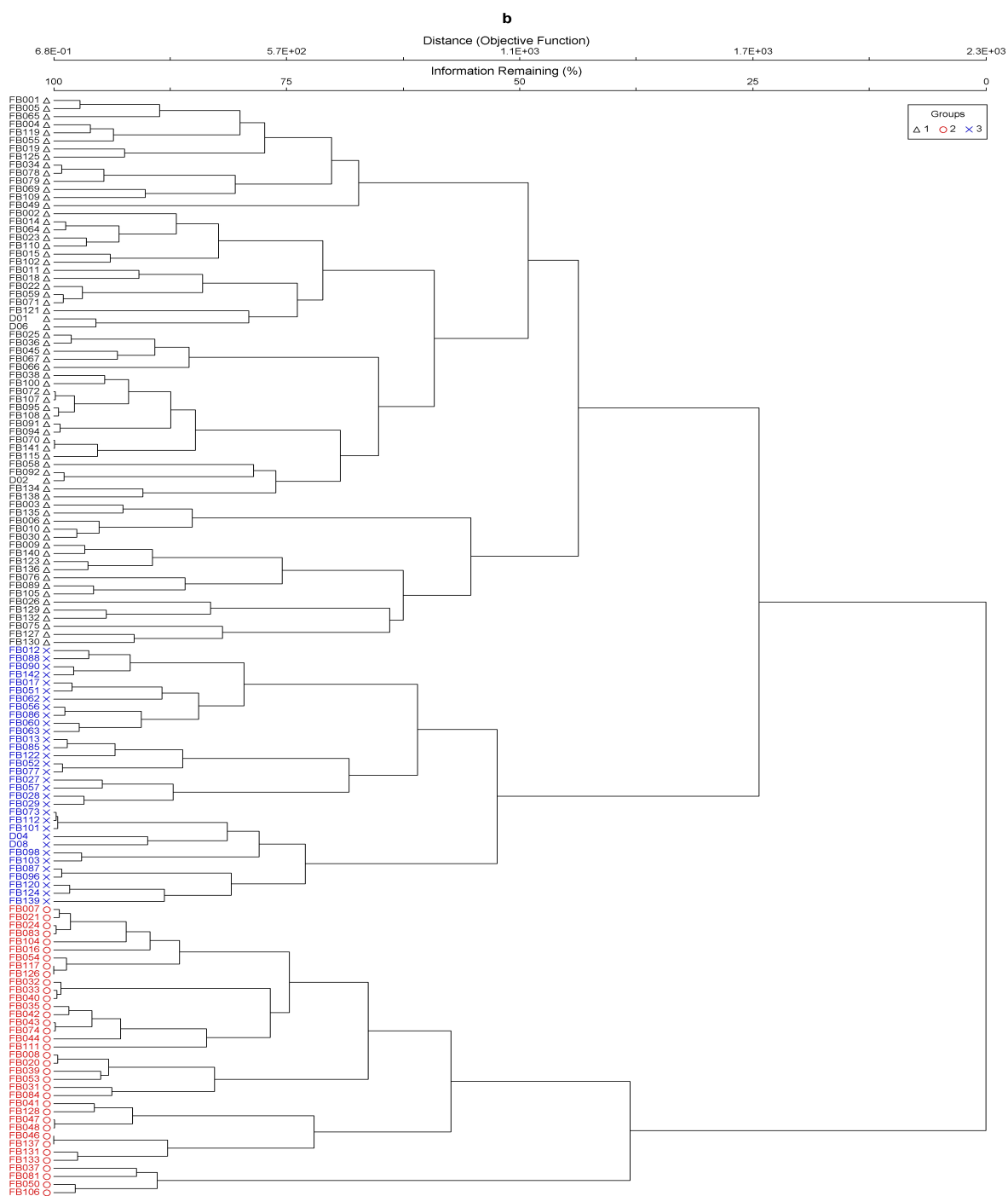


Figure 5.1. Cluster analysis of environmental gradients. Three cluster solutions for: a) topography-moisture gradient; b) fertility gradient. See text for explanation.

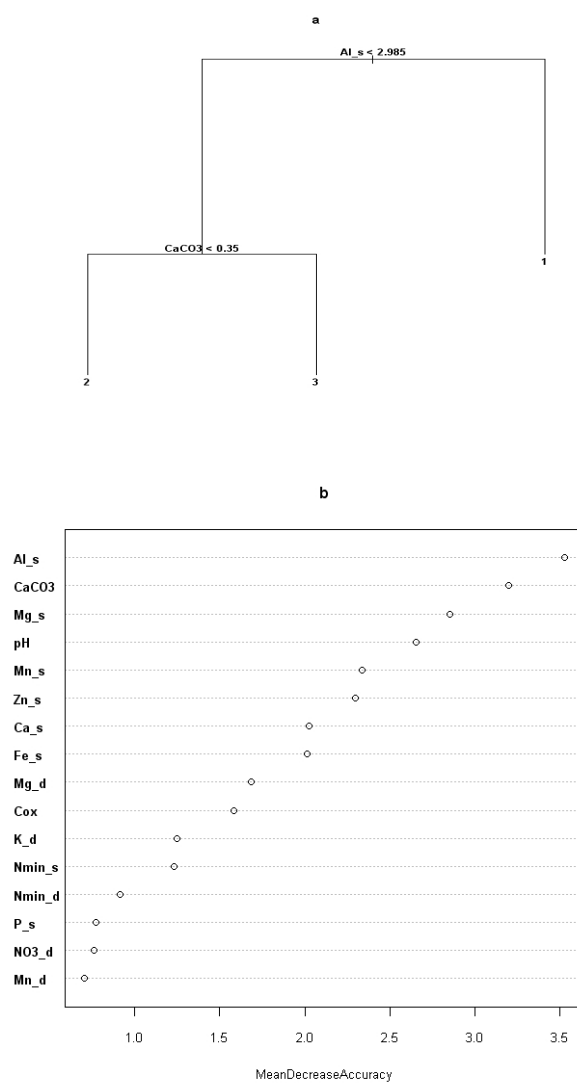


Figure 5.2. CART and RandomForests classification of the fertility gradient: a) a pruned classification tree with thresholds of important variables and three terminal nodes, each represents one fertility class; b) variable importance in RandomForests analysis.

Variables are defined in Table 2.1. See text for explanation.

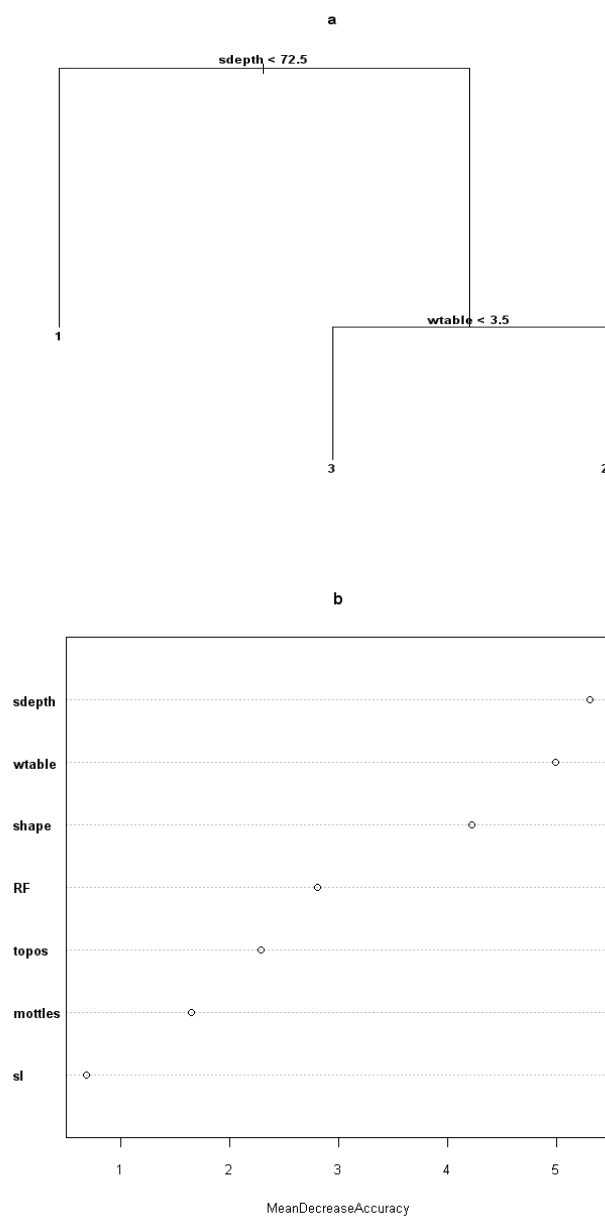
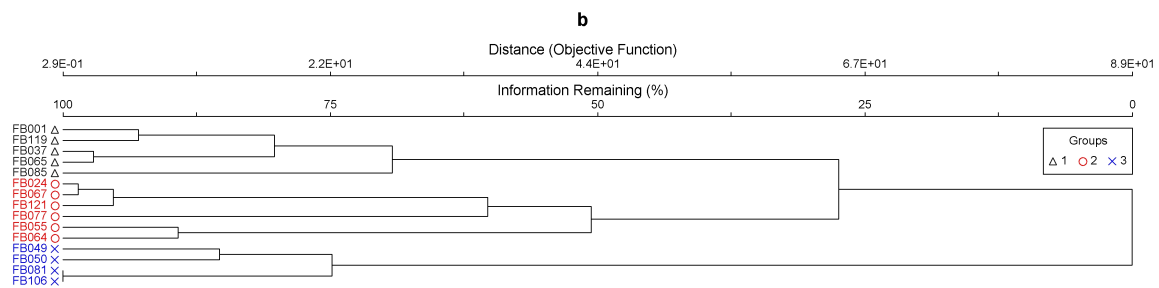
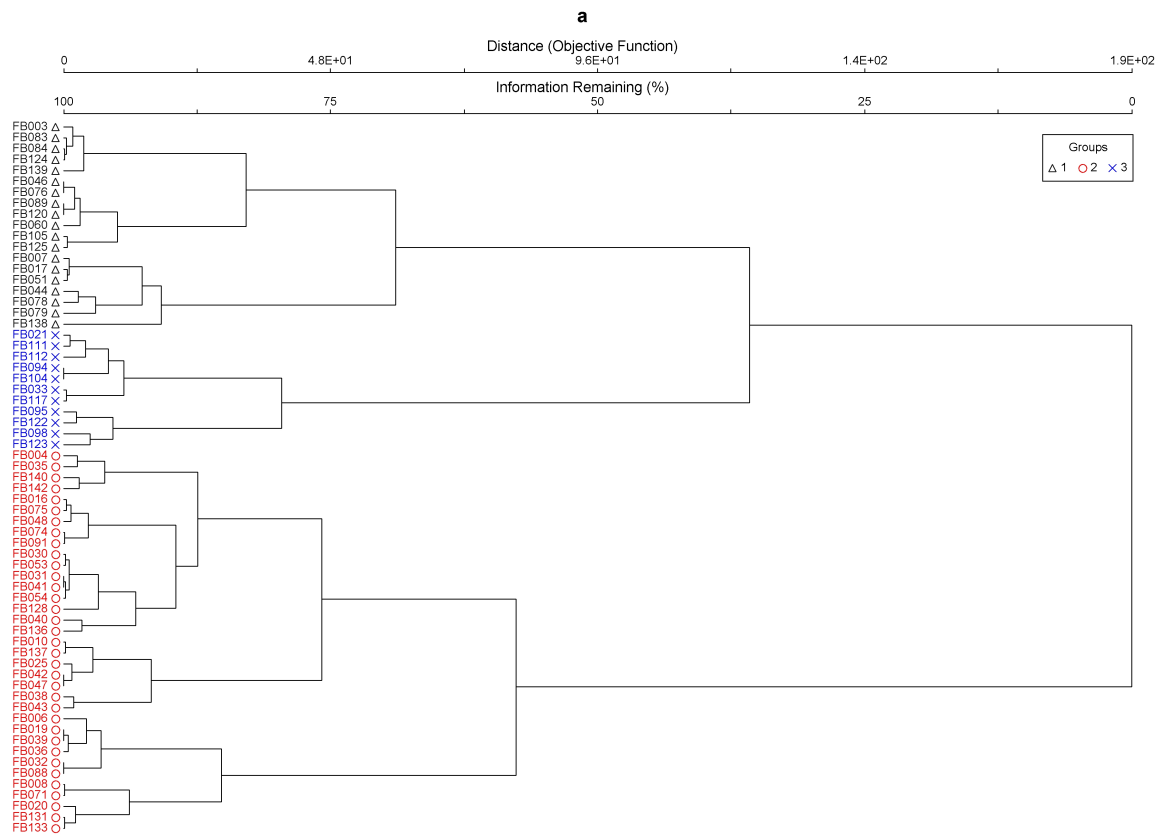


Figure 5.3. CART and RandomForests classification of the topography-moisture gradient: a) a pruned classification tree with thresholds of important variables and three terminal nodes, each represents one topography-moisture class; b) variable importance in RandomForests analysis. Variables are defined in Table 2.1. See text for explanation.



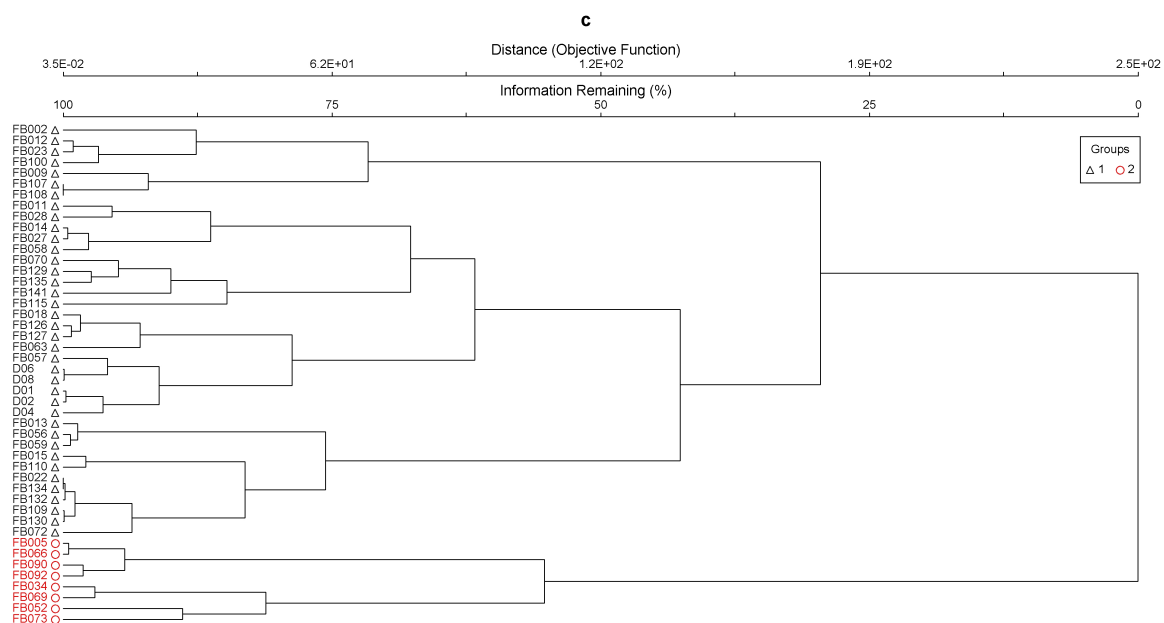
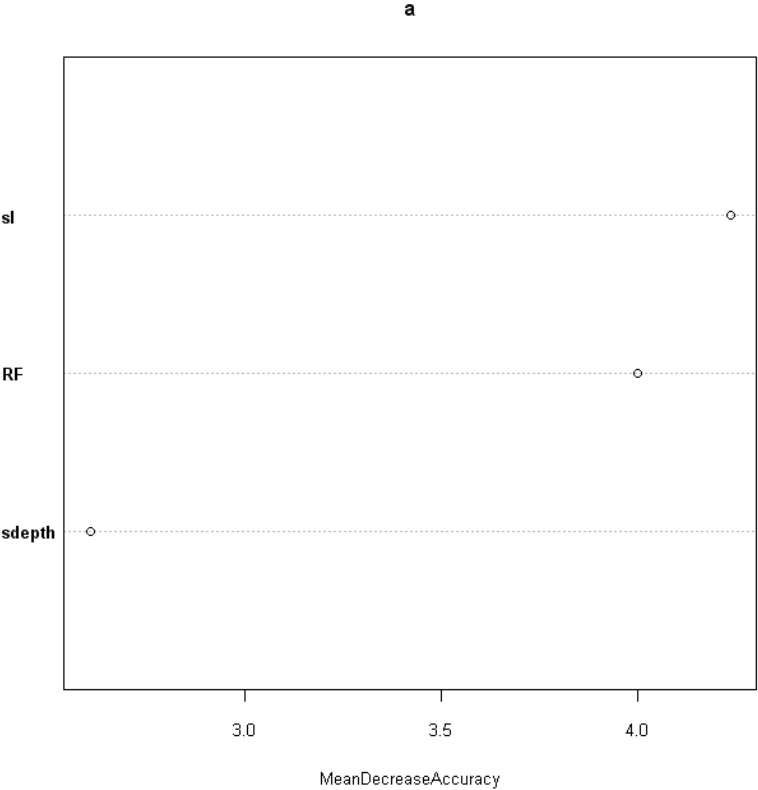
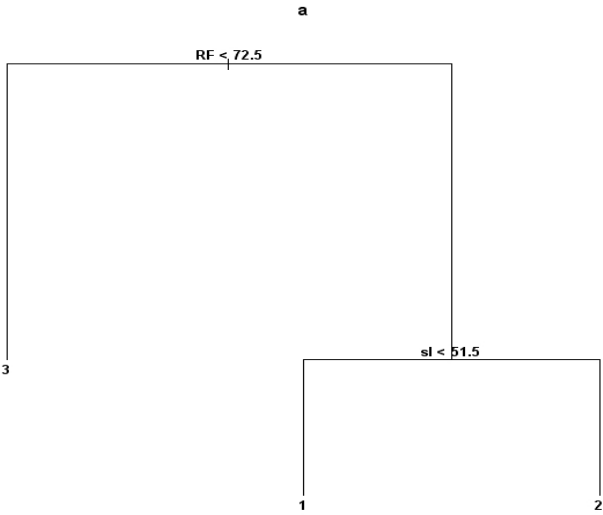
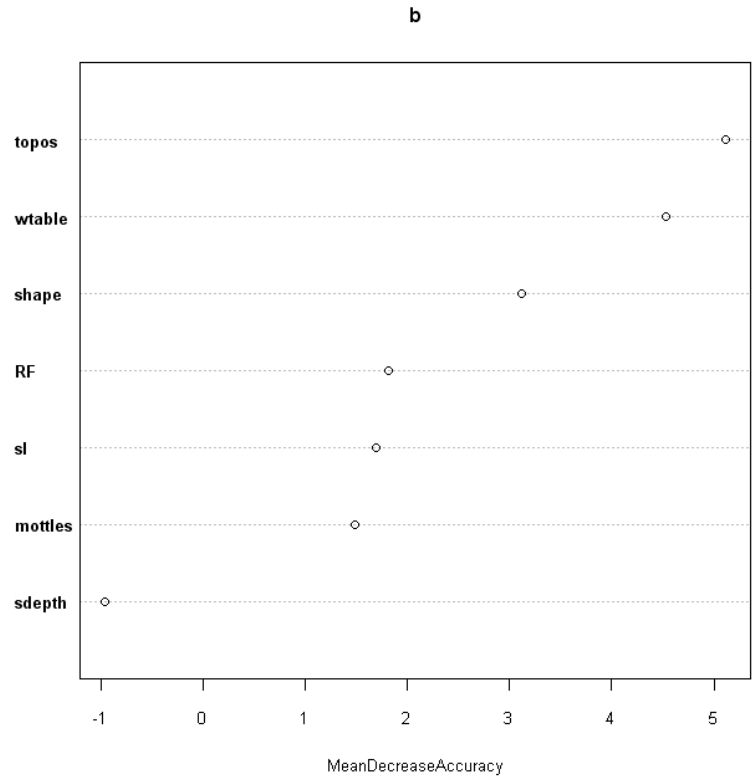
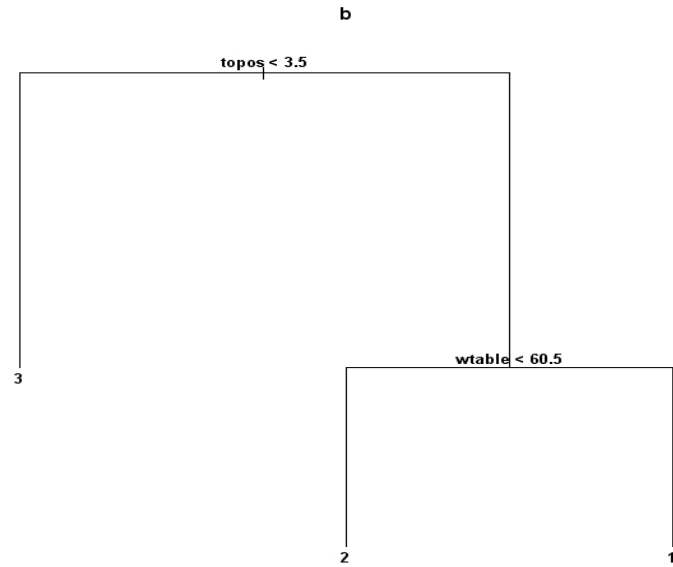


Figure 5.4. Cluster analysis of three topography-moisture classes: a) convex-dry; b) concave-wet; and c) linear-concave-mesic.





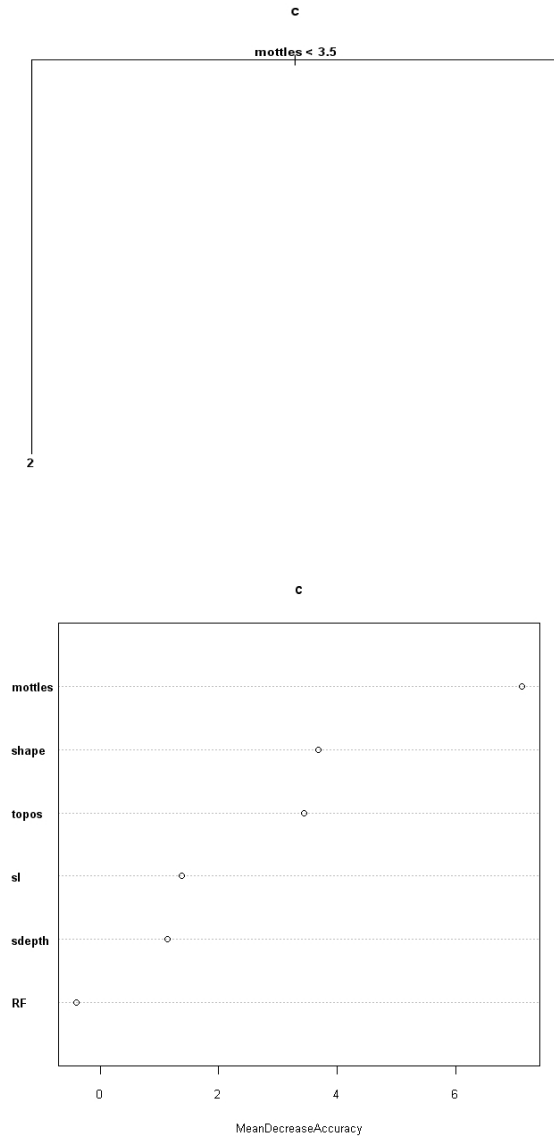


Figure 5.5. CART and RandomForests classification of the topography-moisture classes: a) convex-dry; b) concave-wet; and c) linear-concave-mesic. Pruned classification trees with thresholds of important variables and terminal nodes for final topography-moisture classes. Variable importance in RandomForests analysis. Variables are defined in Table 2.1. See text for explanation.

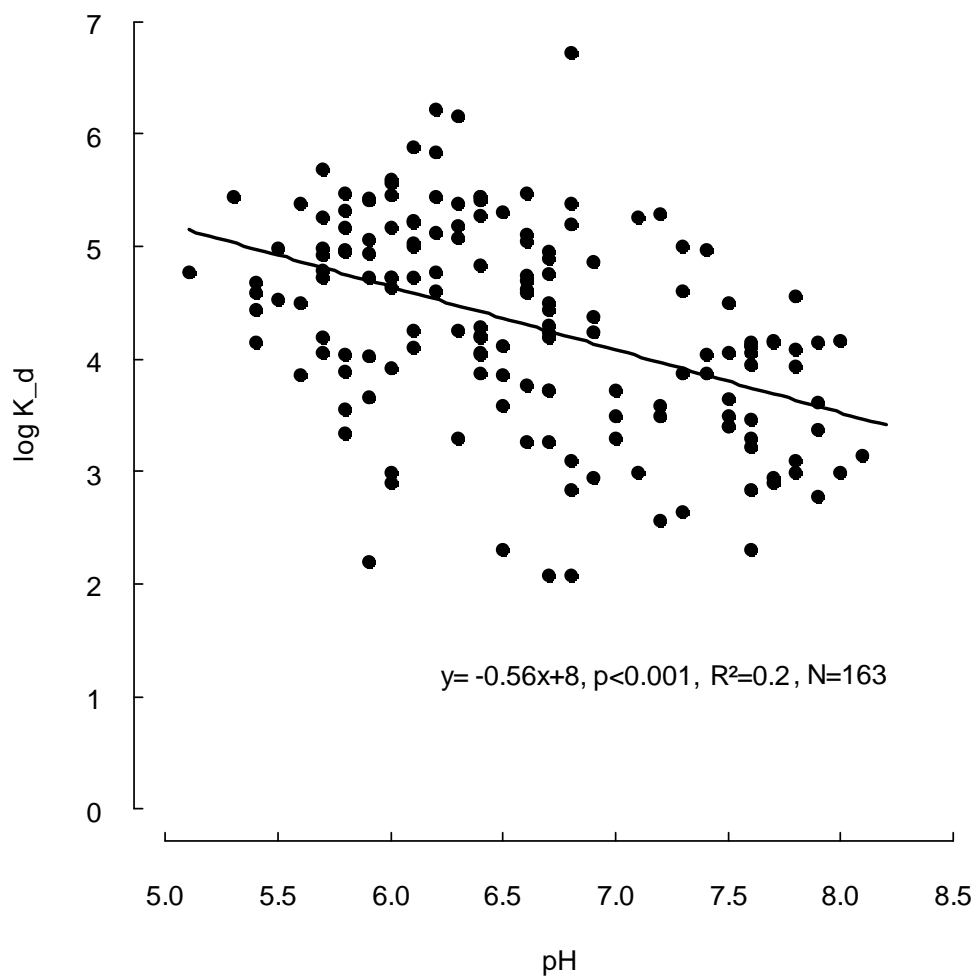
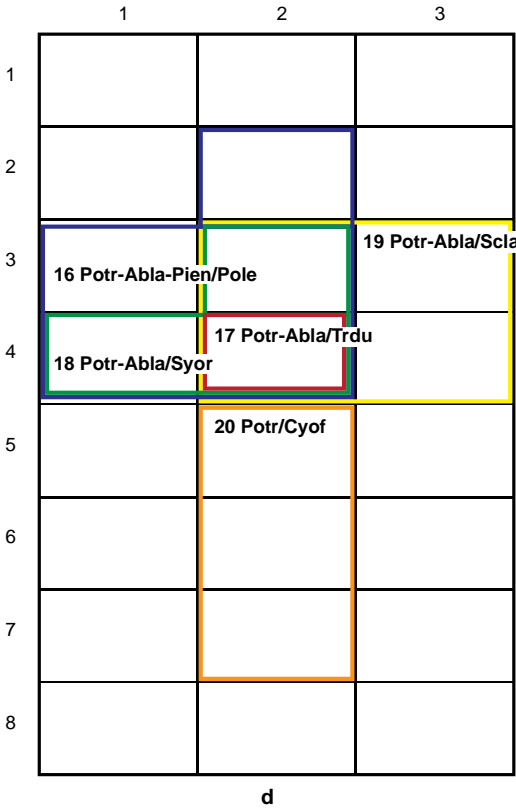
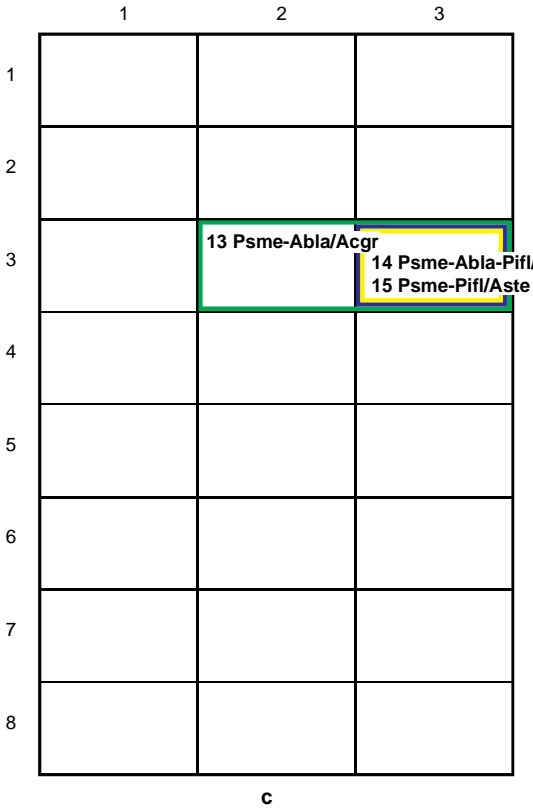
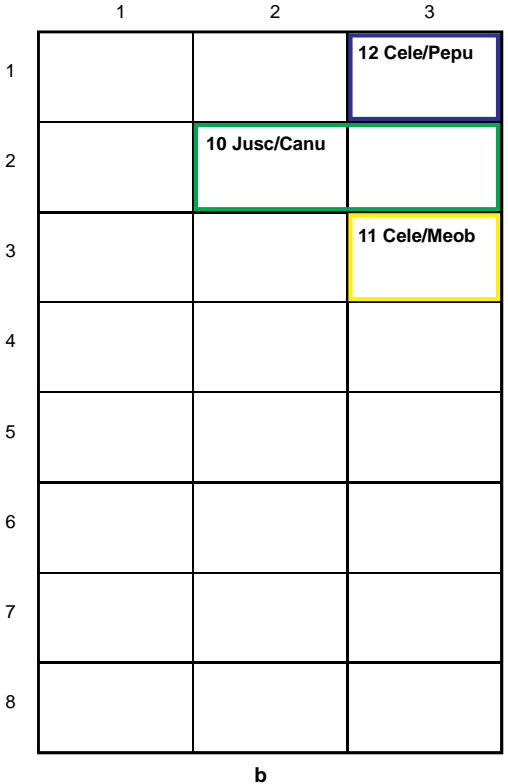
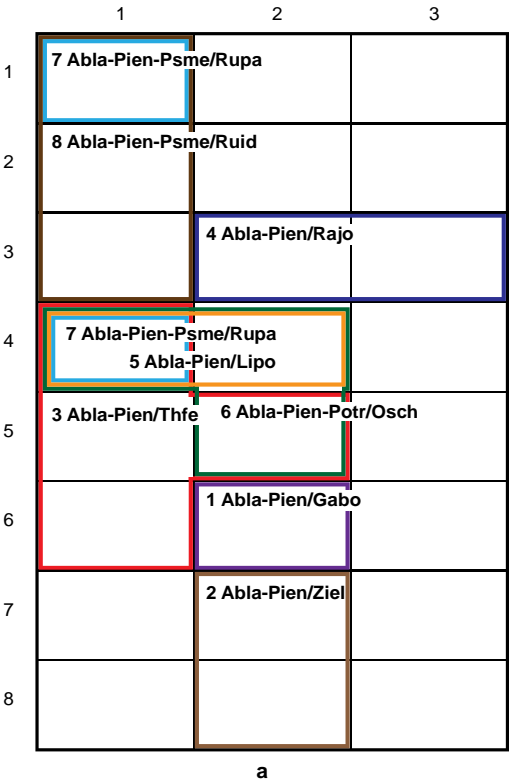


Figure 5.6. Regression of soil pH with K supply rate. Less K is available in a soil solution in high soil pH. The same significant pattern was found for the other essential nutrient supply rates (N, P, Fe).



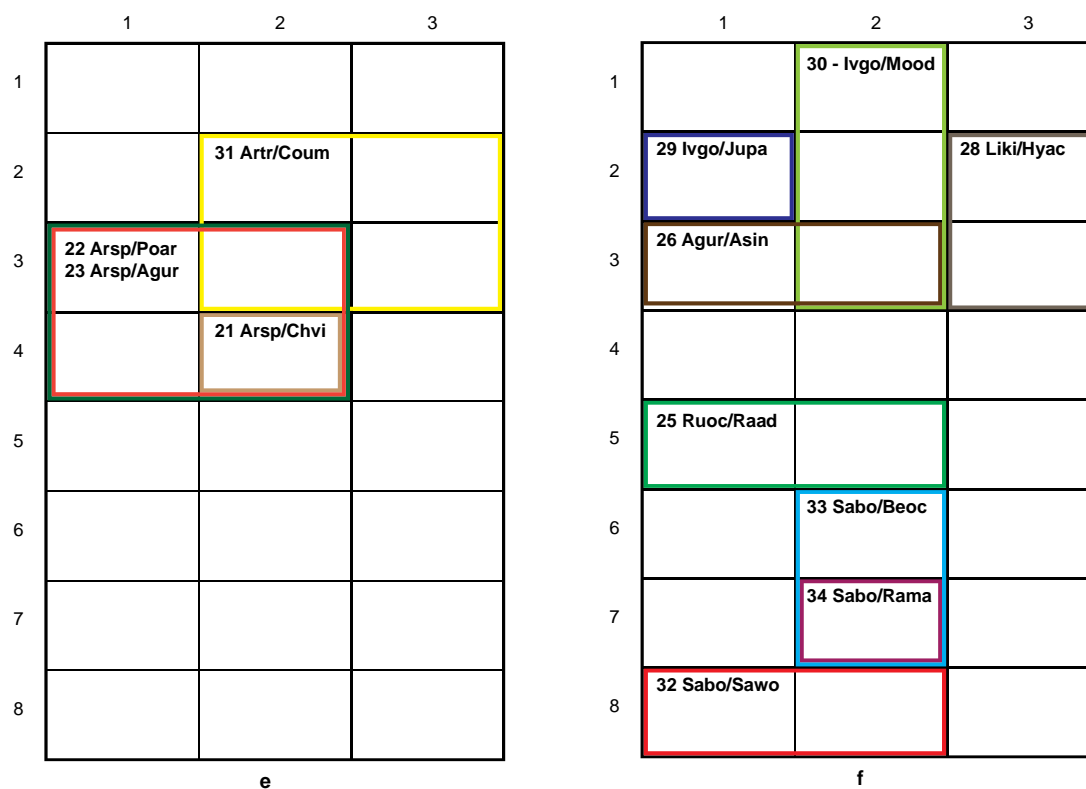


Figure 5.7. Site-vegetation ecological grids showing plant associations and their relation to topography-moisture and fertility. Site classes (1-8 for topography-moisture, 1-3 for fertility) are defined in Table 5.3. Plant associations are defined in Table 5.4.: a) spruce-fir; b) woodland; c) Douglas-fir; d) aspen; e) sagebrush; and f) tall-forb meadows.

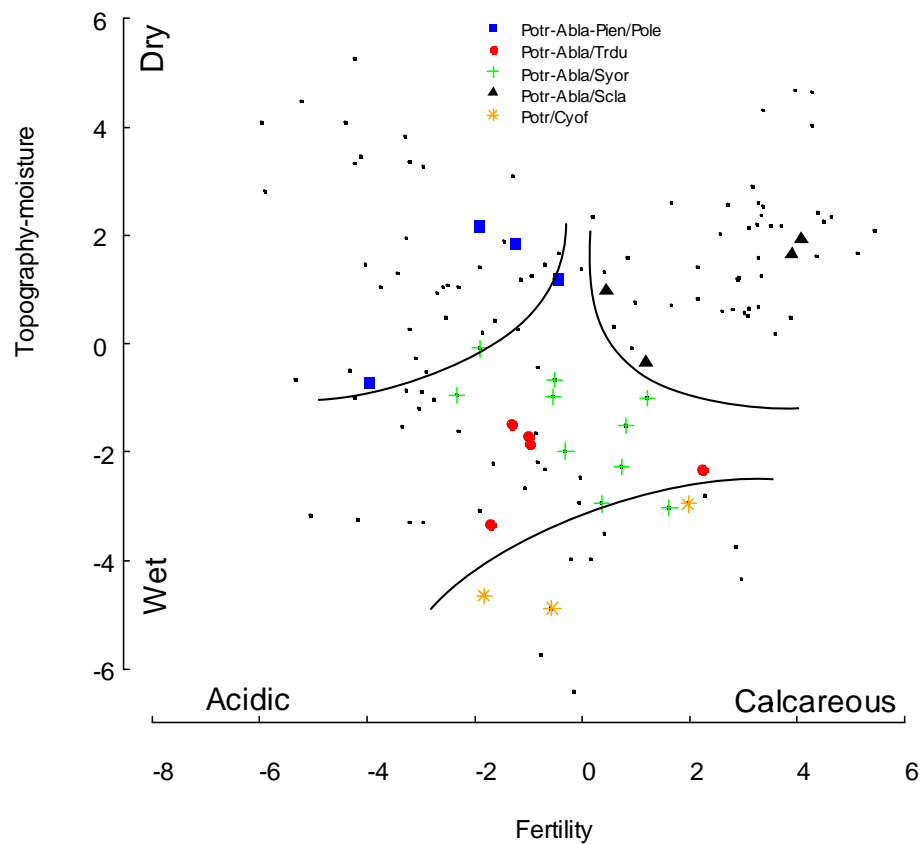


Figure 5.8. Environmental difference among aspen communities projected into a continuous ordination space represented by fertility and topography-moisture axes. Note that the 3x8 dimensional site grid – the site classes (Fig. 5.7) were replaced by continuous ordination scores on both axes. Environmentally discrete communities (poor, wet and alkaline-calcareous) are separated by curves. Environmentally intermediate communities meet in the center. Aspen communities are defined in Table 5.4.

CHAPTER 6

SUMMARY AND CONCLUSIONS

Ecosystem organization

The proposed terrestrial ecosystem classification was developed based on a conceptual framework of ecosystem hierarchical organization. This framework is consistent with the general belief that physical environment (site), represented by climate, soil moisture and soil nutrients, considerably contributes to vegetation distribution. We confirmed the environmental heterogeneity of the study area, derived important environmental gradients influencing ecosystem patterns within that area, and synthesized these gradients into ecologically meaningful levels of ecosystem organization: (1) macroclimate or regional climate; (2) mesoclimate or local climate; and (3) soil fertility.

This organizational structure is consistent with the lower elements of the ECOMAP/TEUI standard and was used as a framework for additional structuring of the organization levels e.g., site classification in building a comprehensive ecosystem classification.

Zonal classification

In order to better understand broad vegetation patterns in a Rocky Mountain landscape, we examined the relationships between vegetation and environmental variables for zonal sites (*sensu* Krajina and Bailey) as represented by sites with mature vegetation, moderate topographic and intermediate soil characteristics. We assessed and classified the response of the complex vegetation to those important environmental factors operating at the highest level of our ecosystem organization - regional climate.

We defined two vegetation geo-climatic zones as areas with the same floristic structure in climatic climax. These zones were: montane with juniper/Douglas-fir; and subalpine with Engelmann spruce/subalpine fir as climatic climax species. We characterized these zones based on climate and landform geomorphology/soils. Regional climate was represented by elevation, precipitation, air and soil temperatures; and geomorphology by zonal soil types. Aspen was excluded from the zonation due its great ecological amplitude. Even stable or climax aspen was not considered as zonal (climatic climax) vegetation.

We argue the vegetation geo-climatic zonation is a conceptual improvement on earlier approaches to vegetation zonation in the region.

Vegetation classification

To answer the question “What are vegetation patterns and important species assemblages in the study area?” we performed a vegetation classification based on the concept of diagnostic species and fidelity. We identified thirty-four vegetation units; for each species, fidelity and constancy was calculated and diagnostic species were identified at the floristic level of plant alliances and associations. These species are useful for recognition of the vegetation units in the field.

Diagnostic species were compared with indicator species of extensive habitat type classification in the central Rocky Mountains. We assumed that diagnostic species are more indicative of the underlying environment than indicator species *sensu* habitat type because they reflect a closer relation of species with the physical environment.

Additionally, our vegetation classification describing existing vegetation across a broad range of ecosystems (forest, woodland, riparian, non-forested) reflects vegetation

dynamics and may be included as a part of a comprehensive ecosystem classification.

Site classification

In order to better understand the relationship of vegetation with physical environment in the study area, we performed a site classification based on the most important environmental factors and gradients within the subalpine vegetation geo-climatic zone. These gradients were topography-moisture and fertility. We specified important site classes associated with each gradient and constructed a 3x8 dimensional site grid.

We overlaid the existing vegetation classification onto the site classification resulting in allocation of plant associations into the site grid within the subalpine vegetation geo-climatic framework. Based on this allocation, we identify ecosystems as plant associations connected to relatively stable physical environment that is defined by the site grid. Each ecosystem was thus defined by a particular site quality and compared with other ecosystems within the framework of the site grid.

This comprehensive ecosystem classification integrates three independent classifications: zonal (climatic); vegetation; and site classification.

Reflection of the ecosystem classification

The independent classifications synthesized into the comprehensive ecosystem classification provided insight into fundamental ecological questions. We conclude: (1) vegetation zonation is climate based; (2) there are two firm vegetation geo-climatic zones in the study area; (3) thirty-four important plant communities exist in the study area; (4) in addition to climate, there are topography-moisture and soil fertility environmental factors associated with the distribution of vegetation.

There was a remarkable pattern that emerged across the independent classifications; namely that it is very difficult to designate aspen as a consistent ecosystem environmentally. In Chapter 3, aspen could not be distinguished as a separate vegetation geo-climatic zone relative to montane and subalpine zone. In Chapter 4, aspen emerged “fuzzy” floristically as a vegetation unit except in the case of moist and wet communities. In Chapter 5, aspen communities overlap pretty much environmentally, again, except for moist and wet communities, illustrating the broad environmental amplitudes (moisture and fertility) of aspen units. Based on that aspen pattern, we could not find any environmental factor able to clearly separate seral aspen communities from persistent communities in this study. There are probably other factors inherent to aspen that are responsible for its climax or persistent status that are little affected by physical environment.

By integrating three classifications, the comprehensive classification system provides insight into processes and outcomes associated with, for example, disturbance history and changes of ecosystems under climate change. Knowing a plant community's topographical-moisture context and fertility demands, and site vegetation potential, we can assess site disturbance history and vegetation successional status. Additionally, based on knowledge of site properties, we can better estimate potential changes of these properties under different climate change scenarios.

These examples reflect the potential value of a comprehensive classification system. A tool for communication remains its overarching function. The system can be used as an efficient tool not only in ecosystem studies and research but also in applied ecosystem interpretation, planning and management.

APPENDICES

APPENDIX A

ZONAL SOILS PROPERTIES AND CLASSIFICATION

Variables are defined in Table 2.1. Soil description and classification follow Schoeneberger et al. (2002) and Soil Survey Staff (2006).

Plot: FB013; elevation: 2515 m; parent material: glacial till; vegetation: spruce-fir; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (μg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
O	0-5	hemimor													
A	5-9	10YR 5/2	10YR 4/2	scl	5.4		15	9.9	108.0	1018.0	109.0	14.6	1.4	0.0	71
E	9-28	10YR 6/3	10YR 4/4		5.4		10								
Bt1	28-60	10YR 6/4		sc	5.6		50								
Bt2	60-120	10YR 6/4			5.7	60	60								
C	120-130	10YR 5/3					60+								
Eutric Haplocryalfs															

Plot: FB014; elevation: 2395 m; parent material: glacial till; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A	0-30	7.5YR 3/2	7.5YR 2.5/2	scl	6.1		10	74.3	616.0	2848.0	161.0	2.9	3.7	0.3	66
E	30-55	10YR 6/3	5YR 4/4		6.4		20								
Bt1	55-120	5YR 5/4	5YR 6/4	sc			75								
Bt2	120-130				75										
Typic	Argicryolls														

Plot: FB015; elevation: 2400 m; parent material: glacial till; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A1	0-15	10YR 5/3	10YR 3/3	scl	6.1		+	59.1	1470.0	2597.0	156.0	3.7	3.4	0.3	100
A2	15-30	10YR 5/3	10YR 3/4				+								
EA	30-60	10YR 6/4	10YR 5/4	cl			10								
Bt	60-80			sc		70	15								
BCt	80-100						75+								
Eutric	Haplocryalfs														

Plot: FB018; elevation: 2155 m; parent material: glacial till; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	($\mu\text{g}/10\text{ cm}^2/6\text{weeks}$)	(%)	(%)	(%)
O	0-6	resimor													
A	6-15	10YR 4/2-3					20					3.2			
AE	15-25	10YR 4/3	10YR 3/3	scl	6.1		30	25.8	346.0	2708.0	104.0		2.7	0.2	100
E1	25-55	10YR 5/3	10YR 4/3	sl-ls			50								
E2	55-85														
Bt	85-100	10YR 5/3	10YR 5/4	sc			50+								
Mollic Palecryalfs															

Plot: FB022; elevation: 2225 m; parent material: glacial till; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	($\mu\text{g}/10\text{ cm}^2/6\text{weeks}$)	(%)	(%)	(%)
O	0-11	humimor													
A1	11-14	10YR 4/2	10YR 3/2				10					1.4			
A2	14-16	10YR 4/3	10YR 3-4/2					31.7	572.0	3116.0	263.0		2.5	0.2	84
AE	16-35	7.5YR 4/3	10YR 4/3	scl	6.5		10								
E	35-50	10YR 6/3	10YR 4/3				15								
Bt1	50-80	10YR 5/3	10YR 4/3	sc		50	20								
Bt2	80-100														
Typic Haplocryalfs															

Plot: FB052; elevation: 2780 m; parent material: glacial till; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(µg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-5	lignomor													
A	5-20	10YR 4-5/3			5.7		10	19.6	133.6	1297.8	327.2	5.3	2.0	0.1	83
EA	20-40	10YR 5/4		scl			20								
E	40-60	10YR 6/3					50								
Bt1	60-75	10YR 6/4				75	60								
Bt2	75-95	10YR 6/6		sc			70								
C	95-120						70+								
Typic	Haplocryalfs														

Plot: FB057; elevation: 2540 m; parent material: glacial till; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(µg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-8	hemimor													
A	8-35	10YR 5/3	10YR 3/2	scl	5.7			53.6	354.0	2184.0	107.0	3.7	3.3	0.2	100
EB	35-70	10YR 6/3	10YR 5/3				+								
Bt	70-100	10YR 5/3-4	10YR 6/4				+								
C	100-120			sc			50+								
Typic	Argicryolls														

Plot: FB059; elevation: 2210 m; parent material: colluvium/till; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A1	0-15						10								
A2	15-45	10YR 4/2	10YR 3/3	cl	6.1		20	73.7	482.0	3329.0	207.0	2.3	5.2	0.4	98
EB	45-60	10YR 6/3	10YR 5/3				50								
Bt	60-110	10YR 6/4	10YR 6/4				50								
BCt	110-120			c			50+								
Pachic	Palecryolls														

Plot: FB063; elevation: 2745 m; parent material: colluvium/quartzite; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A	0-35	10YR 5/3-4	10YR 4-5/4	scl	5.7		10	30.8	284.0	892.0	88.0		0.5	0.1	85
BE	35-55	7.5YR 5/6	10YR 5/3				20								
Bt	55-75			sc			50								
BCt	75-100						50+								
Eutric	Haplocryalfs														

Plot: FB069; elevation: 2390 m; parent material: colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (μg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A	0-22	10YR 4/3	10YR 3/3	scl	6.3		10	119.5	241.0	3591.0	699.0	0.5	4.7	0.4	100
AE	22-55	10YR 4/4	10YR 3/4												
E	55-70														
Btg	70-100			sc		70									
Typic	Palecryolls														

Plot: FB073; elevation: 2510 m; parent material: glacial till/colluvium; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-10	resimor													
A	10-20	10YR 5/2-3	7.5YR 4/3	sl	6.0			29.9	275.0	1155.0	252.0	8.0	1.3	0.1	100
E	20-32			_____			10								
Bt1	32-50														
Bt2	50-80			scl		50									
BCt	80-100						10+								
Eutric Haplocryalfs															

Plot: FB107; elevation: 2040 m; parent material: colluvium/alluvium; vegetation: aspen; zone: montane

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A1	0-20	10YR 4/3	10YR 2/1					51.7	220.0	2902.0	290.0	1.5	3.3	0.3	83
A2	20-50	10YR 5/3	10YR 2/2	cl	6.6										
AE	50-83	10YR 4/3-4	10YR 3/3	_____			+								
Bt1	83-110	7.5YR 4/3	7.5YR 3/3												
Bt2	110-130	7.5YR 5/4	7.5YR 4/4	c											
Pachic Palecryolls															

Plot: FB109; elevation: 2440 m; parent material: colluvium/outwash; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A	0-40	10YR 3/2	10YR 2/2	cl	6.7		+	104.2	324.0	6005.0	785.0	1.0	6.6	0.4	96
AB	40-70	10YR 5/3	10YR 4/3												
Bt	70-120	10YR 5/4	10YR 4/4	c			40								
BCt	120-130														

Pachic Palecryolls

Plot: FB110; elevation: 2425 m; parent material: colluvium/outwash; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A	0-38	10YR 3/3	10YR 2/2	cl	6.9		+	58.4	366.0	4487.0	375.0	4.7	4.4	0.3	84
BA	38-62	10YR 5/3	10YR 4/3												
Bt1	62-100	10YR 5/4	10YR 4/4	c			+								
Bt2	100-120														

Typic Palecryolls

Plot: FB115; elevation: 2255 m; parent material: colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A	0-30	10YR 3/3	10YR 2/2	cl	6.8		+	91.8	394.0	3459.0	376.0	2.0	4.3	0.3	80
AE	30-45	10YR 5/2	10YR 3/2												
E	45-70	5YR 6/3	5YR 5/2	c			10								
Bt	70-120	5YR 5/4	5YR 4/4												

Pachic Palecryolls

Plot: FB116; elevation: 2480 m; parent material: colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A	0-20	10YR 5/3	10YR 3/2-3	scl	6.9		10	89.1		5109.0	452.0	2.9	7.8	0.4	82
EA	20-35	7.5YR 6/3	7.5YR 4/3-4				30								
Bt1	35-50	7.5YR 6/4	7.5YR 5/6	sc	6.5		50								
Bt2	50-75	5YR 5/6	5YR 4/6												

Mollic Haplocryalfs

Plot: FB127; elevation: 2350 m; parent material: colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A1	0-21	10YR 3/2	10YR 2/2	scl	6.0		5	111.4	252.0	5484.0	546.0	2.8	8.8	0.6	88
A2	21-55	10YR 3/2	10YR 3/3				5								
E	55-70	10YR 6/3	10YR 4/4	c			60								
Bt	70-100	5YR 5/4	5YR 5/6												

Pachic Palecryolls

Plot: FB129; elevation: 2245 m; parent material: colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A1	0-20	10YR 3/3	10YR 3/3	scl	6.2		10	65.9	220.0	3080.0	209.0	3.4	3.8	0.3	74
A2	20-55	10YR 4/3	10YR 3/3												
EBt	55-70	7.5YR 5/4	7.5YR 4/6	c			30								
Bt	70-100	7.5YR 6/4	7.5YR 5/6												

Pachic Palecryolls

Plot: FB132; elevation: 2465 m; parent material: glacial till; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A1	0-20	10YR 4/2-3	10YR 3/3		6.3		+	117.7	325.0	2790.0	196.0	1.0	4.3	0.3	74
A2	20-35	10YR 4/3	10YR 3/3				10								
AE	35-60		10YR 3/3	scl			20								
EBt	60-100		10YR 4/4				35								
Bt	100-120		7.5YR 4/6	sicl			35								
Pachic Palecryolls															

Plot: FB134; elevation: 2300 m; parent material: glacial till/colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
A	0-10	10YR 4/2	10YR 3/3	scl	6.2		10	35.5	295.0	6088.0	575.0	0.4	10.2	0.6	86
AE	10-42	10YR 4/4	10YR 3/3	sc			10								
Bt1	42-75		10YR 3/6				15								
Bt2	75-100		10YR 4/4	c			20								
BCt	100-120		10YR 4/6				20								
Pachic Argicryolls															

Plot: FB136; elevation: 2750 m; parent material: quartzite; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(µg/10 cm ² /6weeks)	(%)	(%)	(%)
A	0-20	10YR 5/2	10YR 3/2	scl	5.9		30	45.6		1906.0	128.0	14.2	3.9	0.2	75
Bw	20-50	10YR 6/3	10YR 4/4				60								
Typic	Haplocryolls														

Plot: FB140; elevation: 2640 m; parent material: till/quartzite; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A1	0-15	10YR 5/3	10YR 3/3		6.0		15	28.7		1178.0	118.0	8.6	1.7	0.1	90
A2	15-37	10YR 5/4	10YR 3/3	sl			30								
Bw	37-56	10YR 6/4	10YR 5/4				50								
BC	56-70	10YR 7/4	10YR 6/4				75+								

Typic Haplocryolls

Plot: FB141; elevation: 2600 m; parent material: colluvium; vegetation: aspen; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
A1	0-10	10YR 4/2	10YR 3/2		6.6		10	33.1	120.0	3197.0	158.0	1.2	2.9	0.2	91
A2	10-38	10YR 4/3	10YR 3/3				15								
AE	38-70	10YR 4/4		sc			20								
Bt	70-90	10YR 5/4					20								
C	90-100						30								

Pachic Palecryolls

Plot: B3; elevation: 2285 m; parent material: colluvium; vegetation: Douglas-fir; zone: montane

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-6	humimor													
A	6-16	10YR 4/3	10YR 3/3	sc1	6.6		+	58.6	722.0	4603.0	315.0	0.6	3.5	0.2	100
AE	16-45	10YR 4/4	10YR 3/3				+								
Bt1	45-68	10YR 5/4	10YR 4/4		6.4		10								
Bt2	68-90	10YR 6/3	10YR 5/6	sc			30								
BCt	90-120														

Typic Argixerolls

Plot: D1; elevation: 2660 m; parent material: Wasatch Formation; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-7	humimor													
A	7-12	10YR 4/3	10YR 3/3	scl	6.1		20	22.1	131.0	1542.0	63.0	3.6	1.3	0.0	100
E	15-65	10YR 6/4	7.5YR 4/6				25								
EBt	65-95	7.5YR 6/6	7.5YR 5/8				35								
BCt	95-100						70								
Typic	Haplocryalfs														

Plot: D2; elevation: 2685 m; parent material: Wasatch Formation; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-7	humimor													
A	7-10	7.5YR 5/3	7.5YR 2.5/3	scl	6.0		20	49.6	235.0	1957.0	84.0	1.8	2.8	0.1	100
E	10-75	7.5YR 5/4	7.5YR 3/4	s			30								
EBt	75-100	7.5YR 6/6	7.5YR 4/6	scl			30								
BCt	100-110	5YR 6/6	5YR 4/6	sc			50								
Typic	Haplocryalfs														

Plot: D4; elevation: 2635 m; parent material: Wasatch Formation; vegetation: spruce-fir; zone: subalpine

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-6	lignomor													
A	6-11	10YR 5/4	10YR 3/3		6.6		20	23.9	265.0	1766.0	95.0	4.6	2.9	0.1	100
E	11-55	7.5YR 5/3	7.5YR 3/4	scl			50								
EBt	55-70	7.5YR 6/4	7.5YR 4/4				40								
BCt	70-120	5YR 5/6	5YR 4/6	cl			50								
Typic	Haplocryalfs														

Plot: D6; elevation: 2580 m; parent material: Wasatch Formation; vegetation: spruce-fir; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (μg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
O	0-5	hemimor													
A	5-8	10YR 4/3	10YR 3/3	scl	6.2		15	20.9	227.0	2136.0	86.0	10.4	1.5	0.1	100
E	8-28	10YR 6/3	10YR 4/4				20								
EBt	28-55	10YR 7/3	10YR 5/4												
Bt	55-105	7.5YR 6/6	7.5YR 4/6	cl			30								
BCt	105-120	5YR6/6	5YR4/6				50+								
Typic	Haplocryalfs														

Plot: D8; elevation: 2630 m; parent material: Wasatch Formation; vegetation: spruce-fir; zone: subalpine

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (μg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
O	0-5	humimor													
A	5-9	7.5YR 5/4	7.5YR 3/3	sc	5.7		10	22.7	233.0	2542.0	273.0	0.8	1.5	0.1	100
E	9-28	7.5YR 6/4	10YR 3/4				20								
Bt1	28-60	2.5YR 5/6	2.5YR 4/8				10								
Bt2	60-110	2.5YR 5/8	2.5YR 4/8	c			10								
C	110-120			s			+								
Typic	Haplocryalfs														

Plot: R1; elevation: 1920 m; parent material: outwash/colluvium; vegetation: juniper; zone: montane

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (μg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
O	0-1	rhizomull													
A	1-40	10YR 4/2	10YR 2/2	sic	6.6			75.5	732.0	3982.0	637.0	0.8	2.7	0.2	100
AE	40-75	10YR 4-5/3	10YR 3/3												
Bt	75-120	7.5YR 6/6	7.5YR 4/6	c			5								
Pachic	Argixerolls														

Plot: SD1; elevation: 1590 m; parent material: outwash/colluvium; vegetation: juniper; zone: montane

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS	
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)	
A	0-20	10YR 4/2	10YR 3/2	sc	7.0			36.3	517.0	4016.0	628.0	1.2	2.3	0.1	100	
AE	20-40	10YR 5/3	10YR 4/3													
Bt	40-60	10YR 5/4	10YR 4/4	c												
Btk	60-90															
Calcic	Argixerolls															

Plot: T1; elevation: 1810 m; parent material: outwash/till; vegetation: juniper; zone: montane

Horizon	Depth (cm)	Color dry	Color moist	Text	pH	Mottles (cm)	RF (%)	Nmin_s (mg/kg soil)	K_s (mg/kg soil)	Ca_s (mg/kg soil)	Mg_s (mg/kg soil)	Mn_d (µg/10 cm ² /6weeks)	C _{ox} (%)	N _{ox} (%)	BS (%)
O	+	rhizomull													
A	0-35	10YR 4/2	10YR 3/2	sc	6.1		25	38.9	495.0	4688.0	398.0	0.6	3.9	0.3	100
AE	35-50	10YR 5/3	10YR 4/3												
Bt	50-60	10YR 6/4	10YR 4/6	c			50+								
Typic	Argixerolls														

Plot: T2; elevation: 1820 m; parent material: colluvium; vegetation: Douglas-fir; zone: montane

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(μg/10 cm ² /6weeks)	(%)	(%)	(%)
O	0-5	hemimor													
A	5-25	10YR 3/2	10YR 2/2	sc	7.2		+	47.2	1335.0	4080.0	401.0	0.4	3.7	0.3	99
AE	25-50	10YR 4/3	10YR 3/3				10								
EBt	50-70	10YR 6/3	10YR 4/6				10								
Bt	70-100	10YR 5/3	10YR 4/4	c			20								
BCt	100-120	10YR 6/6	10YR 5/6			30									
Typic	Argixerolls														

Plot: T3; elevation: 1830 m; parent material: colluvium; vegetation: Douglas-fir; zone: montane

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	($\mu\text{g}/10\text{ cm}^2/6\text{weeks}$)	(%)	(%)	(%)
O	0-6	hemimor													
A	6-16	10YR 3/2	10YR 2/2	cl	6.4		10	57.1	415.0	3942.0	306.0	0.4	3.0	0.2	100
AE	16-55	10YR 4/3	10YR 3/3				20								
E	55-100	10YR 5/4	10YR 4/4				20								
Bt	100-120	10YR 5/3	10YR 4/3-4	c			20								
Typic		Argixerolls													

Plot: T4; elevation: 1800 m; parent material: outwash/till; vegetation: juniper; zone: montane

Horizon	Depth	Color dry	Color moist	Text	pH	Mottles	RF	Nmin_s	K_s	Ca_s	Mg_s	Mn_d	C _{ox}	N _{ox}	BS
	(cm)					(cm)	(%)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	(mg/kg soil)	($\mu\text{g}/10\text{ cm}^2/6\text{weeks}$)	(%)	(%)	(%)
A1	0-20	10YR 4/3	10YR 3/3	sic	6.1		40	35.1	803.0	3141.0	338.0	1.0	2.3	0.2	100
A2	20-38	10YR 5/3	10YR 4/3				50								
AE	38-55	7.5YR 5/4	7.5YR 4/6				50								
Bt	55-80			c			50+								
Typic		Argixerolls													

APPENDIX B
PLANT SPECIES LIST

List of 324 vascular plant species recognized in the study area following The Plants database (USDA NRCS 2006) and delimited for Utah (Shultz et al. 2006). >> means accepted name.

Genus/Species	Common name	Authority	Symbol
<i>Abies lasiocarpa</i>	subalpine fir	(Hook.) Nutt.	ABLA
<i>Acer glabrum</i>	Rocky Mountain maple	Torr.	ACGL
<i>Acer grandidentatum</i>	bigtooth maple	Nutt.	ACGR3
<i>Aconitum columbianum</i>	Columbian monkshood	Nutt.	ACCO4
<i>Actaea rubra</i>	red baneberry	(Ait.) Willd.	ACRU2
<i>Agastache urticifolia</i>	nettleleaf giant hyssop	(Benth.) Kuntze	AGUR
<i>Agoseris glauca</i>	pale agoseris	(Pursh) Raf.	AGGL
<i>Agropyron cristatum</i>	crested wheatgrass	(L.) Gaertn.	AGCR
<i>Achillea millefolium</i>	common yarrow	L.	ACMI2
<i>Allium bisceptrum</i>	twincest onion	S. Wats.	ALBI2
<i>Alnus incana</i>	gray alder	(L.) Moench	ALIN2
<i>Amelanchier alnifolia</i>	Saskatoon serviceberry	(Nutt.) Nutt. ex M. Roemer	AMAL2
<i>Anemone multifida</i>	>>Pulsatilla patens ssp. multifida	(Pritz.) Zamel, non Poir.	ANMU8
<i>Anemone quinquefolia</i>	>>Anemone multifida var. stylosa	L. var. stylosa (A. Nels.) Dutton & Keener	ANQUS
<i>Angelica arguta</i>	Lyall's angelica	Nutt.	ANAR3
<i>Antennaria luzuloides</i>	rush pussytoes	Torr. & Gray	ANLU2
<i>Antennaria microphylla</i>	littleleaf pussytoes	Rydb.	ANMI3
<i>Antennaria parvifolia</i>	small-leaf pussytoes	Nutt.	ANPA4
<i>Apocynum androsaemifolium</i>	spreading dogbane	L.	APAN2
<i>Aquilegia caerulea</i>	Colorado blue columbine	James	AQCA2
<i>Arabis lyallii</i>	Lyall's rockcress	S. Wats.	ARLY
<i>Arabis nuttallii</i>	Nuttall's rockcress	B.L. Robins.	ARNU
<i>Arenaria congesta</i>	ballhead sandwort	Nutt.	ARCO5
<i>Arnica amplexicaulis</i>	clasping arnica	Nutt. var. <i>piperi</i> H. St. John & Warren	ARAM2
<i>Arnica cordifolia</i>	heartleaf arnica	Hook.	ARCO9
<i>Arnica chamissonis</i>	Chamisso arnica	Less.	ARCH3
<i>Arnica latifolia</i>	broadleaf arnica	Bong.	ARLA8
<i>Arnica longifolia</i>	spearleaf arnica	D.C. Eat.	ARLO6
<i>Arnica mollis</i>	hairy arnica	Hook.	ARMO4
<i>Artemisia arbuscula</i>	little sagebrush	Nutt.	ARAR8
<i>Artemisia dracunculus</i>	tarragon	L.	ARDR4
<i>Artemisia ludoviciana</i>	white sagebrush	Nutt.	ARLU

Genus/Species	Common name	Authority	Symbol
Artemisia spiciformis	>>Artemisia tridentata ssp. spiciformis	Osterhout	ARSP8
Artemisia tridentata	big sagebrush	Nutt.	ARTR2
Aster ascendens	>>Symphyotrichum ascendens	Lindl.	ASAS5
Aster engelmannii	>>Eucephalus engelmannii	(D.C. Eat.) Gray	ASEN2
Aster foliaceus	>>Symphyotrichum foliaceum var. apricum	Lindl. ex DC. var. apricus Gray	ASFOA
Aster glaucodes	>>Eurybia glauca	Blake	ASGL3
Aster hesperius	>>Symphyotrichum lanceolatum ssp. Hesperium, var. hesperium	Gray	ASHE
Aster integrifolius	>>Eurybia integrifolia	Nutt.	ASIN3
Aster occidentalis	>>Symphyotrichum spathulatum var. spathulatum	(Nutt.) Torr. & Gray	ASOC
Aster pauciflorus	alkali marsh aster	(Nutt.) A. Löve & D. Löve	ALPA14
Aster perelegans	>>Eucephalus elegans	A. Nels. & J.F. Macbr.	ASPE3
Astragalus geyeri	Geyer's milkvetch	A. Gray	ASGE
Astragalus tenellus	looseflower milkvetch	Pursh	ASTE5
Balsamorhiza sagittata	arrowleaf balsamroot	(Pursh) Nutt.	BASA3
Balsamorhiza macrophylla	cutleaf balsamroot	Nutt.	BAMA4
Berberis repens	creeping barberry	Lindl.	BERE
Betula glandulosa	resin birch	Michx.	B EGL
Betula occidentalis	water birch	Hook.	BEOC2
Boykinia jamesii	>>Telesonix heucheriformis	(Torr.) Engl. var. heucheriformis (Rydb.) Engl.	BOJAH
Brickellia californica	California brickellbush	(Torr. & Gray) Gray	BRCA3
Bromus anomalus	nodding brome	Rupr. ex Fourn.	BRAN
Bromus carinatus	>>Bromus marginatus	Hook. & Arn. var. linearis Shear	BRCAL2
Bromus ciliatus	fringed brome	L.	BRCI2
Bromus inermis	smooth brome	Leyss.	BRIN2
Bromus tectorum	cheatgrass	L.	BRTE
Calamagrostis rubescens	pinegrass	Buckley	CARU
Calochortus nuttallii	sego lily	Torr. & Gray	CANU3
Carex atrata	>>Carex heteroneura var. epapillosa	auct. p.p. non L.	CAAT13
Carex geyeri	Geyer's sedge	Boott	CAGE2
Carex haydeniana	cloud sedge	Olney	CAHA6
Carex hoodii	Hood's sedge	Boott	CAHO5
Carex microptera	smallwing sedge	Mackenzie	CAMI7

Genus/Species	Common name	Authority	Symbol
<i>Carex multcostata</i>	manyrib sedge	Mackenzie	CAMU6
<i>Carex nebrascensis</i>	Nebraska sedge	Dewey	CANE2
<i>Carex occidentalis</i>	western sedge	Bailey	CAOC2
<i>Carex pachystachya</i>	chamisso sedge	Cham. ex Steud.	CAPA14
<i>Carex phaeocephala</i>	dunhead sedge	Piper	CAPH2
<i>Carex raynoldsii</i>	Raynolds' sedge	Dewey	CARA6
<i>Carex rossii</i>	Ross' sedge	Boott	CARO5
<i>Carex rostrata</i>	>>Carex utriculata	Stokes var. utriculata (Boott) Bailey	CAROU
<i>Carex xerantica</i>	whitescale sedge	L.H. Bailey	CAXE
<i>Castilleja applegatei</i>	wavyleaf Indian paintbrush	Fern.	CAAP4
<i>Castilleja chromosa</i>	>>Castilleja applegatei ssp. martinii	A. Nels.	CACH7
<i>Castilleja linariifolia</i>	Wyoming Indian paintbrush	Benth.	CALI4
<i>Castilleja miniata</i>	giant red Indian paintbrush	Dougl. ex Hook.	CAMI12
<i>Castilleja occidentalis</i>	western Indian paintbrush	Torr.	CAOC4
<i>Castilleja rhexifolia</i>	splitleaf Indian paintbrush	Rydb.	CARH4
<i>Ceanothus velutinus</i>	snowbrush ceanothus	Dougl. ex Hook.	CEVE
<i>Cercocarpus ledifolius</i>	curl-leaf mountain mahogany	Nutt.	CELE3
<i>Chimaphila umbellata</i>	pipsissewa	(L.) W. Bartram	CHUM
<i>Cirsium calcareum</i>	Cainville thistle	(M.E. Jones) Woot. & Standl.	CICA10
<i>Cirsium eatonii</i>	Eaton's thistle	(Gray) B.L. Robins.	CIEA
<i>Cirsium hookerianum</i>	white thistle	Nutt.	CIHO
<i>Cirsium undulatum</i>	wavyleaf thistle	(Nutt.) Spreng.	CIUN
<i>Claytonia lanceolata</i>	lanceleaf springbeauty	Pall. ex Pursh	CLLA2
<i>Clematis occidentalis</i>	western blue virginsbower	(Hornem.) DC.	CLOC2
<i>Collinsia parviflora</i>	maiden blue eyed Mary	Lindl.	COPA3
<i>Collomia linearis</i>	tiny trumpet	Nutt.	COLI2
<i>Comandra umbellata</i>	bastard toadflax	(L.) Nutt.	COUM
<i>Corallorhiza striata</i>	hooded coralroot	Lindl.	COST
<i>Corallorhiza trifida</i>	yellow coralroot	Chatelain	COTR3
<i>Cornus sericea</i>	redosier dogwood	L.	COSE16
<i>Crepis acuminata</i>	tapertip hawksbeard	Nutt.	CRAC2
<i>Cymopterus hendersonii</i>	>>Pteryxia hendersonii	(Coult. & Rose) Cronq.	CYHE3

Genus/Species	Common name	Authority	Symbol
<i>Cynoglossum officinale</i>	gypsyflower	L.	CYOF
<i>Cystopteris fragilis</i>	brittle bladderfern	(L.) Bernh.	CYFR2
<i>Dactylis glomerata</i>	orchardgrass	L.	DAGL
<i>Delphinium nuttallianum</i>	twolobe larkspur	Pritz. ex Walp.	DENU2
<i>Delphinium occidentale</i>	>>Delphinium occidentale	(S. Wats.) S. Wats. ssp. cucullatum (A. Nels.) Ewan	DEOCC
<i>Descurainia californica</i>	Sierra tansymustard	(Gray) O.E. Schulz	DECA6
<i>Descurainia pinnata</i>	western tansymustard	(Walt.) Britt.	DEPI
<i>Descurainia richardsonii</i>	>>Descurainia incana ssp. incisa	O.E. Schulz ssp. incisa (Engelm.) Detling	DERII
<i>Deschampsia cespitosa</i>	>>Deschampsia caespitosa	(L.) Beauv. [orthographic variant]	DECE
<i>Deschampsia elongata</i>	slender hairgrass	(Hook.) Munro	DEEL
<i>Disporum trachycarpum</i>	roughfruit fairybells	(S. Wats.) Benth. & Hook. f.	DITR2
<i>Draba crassifolia</i>	snowbed draba	Graham	DRCR2
<i>Draba maguirei</i>	Maguire's draba	C.L. Hitchc.	DRMA2
<i>Draba nemorosa</i>	woodland draba	L.	DRNE
<i>Elymus canadensis</i>	Canada wildrye	L.	ELCA4
<i>Elymus cinereus</i>	>>Leymus cinereus	Scribn. & Merr.	ELCI2
<i>Elymus elymoides</i>	squirreltail	(Raf.) Swezey	ELEL5
<i>Elymus glaucus</i>	blue wildrye	Buckl.	ELGL
<i>Elymus lanceolatus</i>	thickspike wheatgrass	(Scribn. & J.G. Sm.) Gould ssp. <i>lanceolatus</i>	ELLAL
<i>Elymus scribneri</i>	spreading wheatgrass	(Vasey) M.E. Jones	ELSC4
<i>Elymus spicatus</i>	bluebunch wheatgrass	(Pursh) Gould	ELSP3
<i>Elymus trachycaulus</i>	slender wheatgrass	(Link) Gould ex Shinnars	ELTR7
<i>Epilobium angustifolium</i>	>>Chamerion angustifolium ssp. circumvagum	L. ssp. circumvagum Mosquin	EPANC
<i>Epilobium brachycarpum</i>	tall annual willowherb	K. Presl	EPBR3
<i>Epilobium canum</i>	hummingbird trumpet	(Greene) P.H. Raven	EPCA3
<i>Epilobium ciliatum</i>	fringed willowherb	Raf.	EPCI
<i>Equisetum arvense</i>	field horsetail	L.	EQAR
<i>Erigeron eatonii</i>	Eaton's fleabane	Gray	EREA
<i>Erigeron speciosus</i>	aspen fleabane	(Lindl.) DC.	ERSP4
<i>Eriogonum caespitosum</i>	matted buckwheat	Nutt.	ERCA8
<i>Eriogonum heracleoides</i>	parsnipflower buckwheat	Nutt.	ERHE2
<i>Eriogonum umbellatum</i>	sulphur-flower buckwheat	Torr.	ERUM

Genus/Species	Common name	Authority	Symbol
Erysimum asperum	>>Erysimum capitatum var. capitatum	(Nutt.) DC.	ERAS2
Erythronium grandiflorum	yellow avalanche-lily	Pursh	ERGR9
Festuca idahoensis	Idaho fescue	Elmer	FEID
Festuca subulata	bearded fescue	Trin.	FESU
Fragaria vesca	woodland strawberry	L.	FRVE
Fragaria virginiana	Virginia strawberry	Duchesne	FRVI
Frasera speciosa	elkweed	Dougl. ex Griseb.	FRSP
Galium aparine	stickywilly	L.	GAAP2
Galium bifolium	twingleaf bedstraw	S. Wats.	GABI
Galium boreale	northern bedstraw	L.	GABO2
Galium triflorum	fragrant bedstraw	Michx.	GATR3
Geranium richardsonii	Richardson's geranium	Fisch. & Trautv.	GERI
Geranium viscosissimum	sticky purple geranium	Fisch. & C.A. Mey. ex C.A. Mey.	GEVI2
Geum macrophyllum	largeleaf avens	Willd.	GEMA4
Gilia aggregata	>>Ipomopsis aggregata ssp. aggregata	(Pursh) Spreng.	GIAG
Glyceria striata	fowl managrass	(Lam.) A.S. Hitchc.	GLST
Goodyera oblongifolia	western rattlesnake plantain	Raf.	GOOB2
Habenaria dilatata	>>Platanthera dilatata var. dilatata	(Pursh) Hook.	HADI7
Habenaria unalascensis	>>Piperia unalascensis	(Spreng.) S. Wats.	HAUN
Hackelia micrantha	Jessica sticktight	(Eastw.) J.L. Gentry	HAMI
Hackelia patens	spotted stickseed	(Nutt.) I.M. Johnston	HAPA
Helianthella uniflora	oneflower helianthella	(Nutt.) Torr. & Gray	HEUN
Heracleum lanatum	>>Heracleum maximum	Michx.	HELA4
Heterotheca villosa	hairy false goldenaster	(Pursh) Shinnars	HEVI4
Heuchera parvifolia	littleleaf alumroot	Nutt. ex Torr. & Gray	HEPA11
Heuchera rubescens	pink alumroot	Torr.	HERU
Hieracium albiflorum	white hawkweed	Hook.	HIAL2
Hieracium cynoglossoides	houndstongue hawkweed	Arv.-Touv.	HICY
Holodiscus dumosus	rockspirea	(Nutt. ex Hook.) A. Heller	HODU
Hydrophyllum capitatum	ballhead waterleaf	Dougl. ex Benth.	HYCA4
Hymenoxys acaulis	>>Tetraneuris acaulis var. arizonica	(Pursh) Parker var. arizonica (Greene) Parker	HYACA2
Chaenactis alpina	>>Chaenactis douglasii var. alpina	(Gray) M.E. Jones	CHAL2

Genus/Species	Common name	Authority	Symbol
Chrysothamnus nauseosus	>>Ericameria nauseosa ssp. nauseosa var. speciosa	(Pallas ex Pursh) Britt. ssp. albicaulis (Nutt.) Hall & Clements	CHNAA3
Chrysothamnus viscidiflorus	yellow rabbitbrush	(Hook.) Nutt.	CHVI8
Ivesia gordonii	Gordon's ivesia	(Hook.) Torr. & Gray	IVGO
Juncus balticus	mountain rush	Willd. var. <i>balticus</i>	JUBAB2
Juncus ensifolius	swordleaf rush	Wikstr.	JUEN
Juncus mertensianus	Mertens' rush	Bong.	JUME3
Juncus parryi	Parry's rush	Engelm.	JUPA
Juniperus communis	common juniper	L.	JUCO6
Juniperus scopulorum	Rocky Mountain juniper	Sarg.	JUSC2
Koeleria macrantha	prairie Junegrass	(Ledeb.) J.A. Schultes	KOMA
Lathyrus lanszwertii	Nevada pea	Kellogg	LALA3
Lathyrus pauciflorus	fewflower pea	Fern.	LAPA5
Lesquerella multiceps	manyhead bladderpod	Maguire	LEMU2
Leucopoa kingii	spike fescue	(S. Wats.) W.A. Weber	LEKI2
Ligusticum filicinum	fernleaf licorice-root	S. Wats.	LIFI
Ligusticum porteri	Porter's licorice-root	Coult. & Rose	LIPO
Linanthes nuttallii	>>Linanthes nuttallii ssp. nuttallii	(Gray) Ewan	LINU4
Linum kingii	King's flax	S. Wats.	LIKI2
Linum lewisii	prairie flax	Pursh	LILE3
Lithophragma parviflorum	smallflower woodland-star	(Hook.) Nutt. ex Torr. & Gray	LIPA5
Lithospermum ruderales	western stone seed	Dougl. ex Lehm.	LIRU4
Lomatium graveolens	king desertparsley	(S. Wats.) Dorn & Hartman	LOGR6
Lomatium grayi	Gray's biscuitroot	(Coult. & Rose) Coult. & Rose	LOGR
Lonicera involucrata	twinberry honeysuckle	Banks ex Spreng.	LOIN5
Lonicera utahensis	Utah honeysuckle	S. Watson	LOUT2
Lupinus argenteus	silvery lupine	Pursh	LUAR3
Lupinus sericeus	silky lupine	Pursh	LUSE4
Machaeranthera commixta	Bigelow's tansyaster	(A. Gray) Greene var. <i>commixta</i> (Greene) B.L. Turner	MABIC
Melica bulbosa	oniongrass	Geyer ex Porter & Coult.	MEBU
Mertensia ciliata	tall fringed bluebells	(James ex Torr.) G. Don	MECI3
Mertensia lanceolata	prairie bluebells	(Pursh) DC. var. <i>coriacea</i> (A. Nelson) Higgins & S.L. Welsh	MELAC
Mertensia oblongifolia	oblongleaf bluebells	(Nutt.) G. Don	MEOB

Genus/Species	Common name	Authority	Symbol
Mimulus guttatus	seep monkeyflower	DC.	MIGU
Mitella pentandra	five-stamen miterwort	Hook.	MIPE
Monardella odoratissima	>>Monardella glauca	Benth. ssp. glauca (Greene) Epling	MOODG2
Nemophila breviflora	basin nemophila	Gray	NEBR
Orthocarpus luteus	yellow owl's-clover	Nutt.	ORLU2
Orthocarpus tolmiei	Tolmie's owl's-clover	Hook. & Arn.	ORTO
Osmorhiza chilensis	>>Osmorhiza berteroi, sweetcicely	DC.	OSBE
Osmorhiza occidentalis	western sweetroot	(Nutt. ex Torr. & Gray) Torr.	OSOC
Parnassia fimbriata	fringed grass of Parnassus	Koenig	PAFI3
Paxistima myrsinites	Mountain lover	(Pursh) Raf.	PAMY
Pedicularis groenlandica	elephant-head lousewort	Retz.	PEGR2
Pedicularis racemosa	sickle-top lousewort	Dougl. ex Benth.	PERA
Pellaea breweri	Brewer's cliffbrake	D.C. Eat.	PEBR4
Penstemon compactus	compact penstemon	(Keck) Crosswhite	PECO10
Penstemon cyananthus	Wasatch beardtongue	Hook.	PECY2
Penstemon humilis	low beardtongue	Nutt. ex Gray	PEHU
Penstemon leonardii	Leonard's beardtongue	Rydb.	PELE9
Penstemon whippleanus	Whipple's penstemon	Gray	PEWH
Petradoria pumila	grassy rockgoldenrod	(Nutt.) Greene	PEPU7
Petrophytum caespitosum	mat rockspirea	(Nutt.) Rydb.	PECA12
Phacelia hastata	silverleaf phacelia	Dougl. ex Lehm.	PHHA
Phacelia heterophylla	varileaf phacelia	Pursh	PHHE2
Phleum alpinum	alpine timothy	L.	PHAL2
Phleum pratense	timothy	L.	PHPR3
Phlox hoodii	spiny phlox	Richardson	PHHO
Phlox pulvinata	cushion phlox	(Wherry) Cronq.	PHPU5
Picea engelmannii	Engelmann spruce	Parry ex Engelm.	PIEN
Pinus contorta	lodgepole pine	Dougl. ex Loud.	PICO
Pinus flexilis	limber pine	James	PIFL2
Plantago eriopoda	redwool plantain	Torr.	PLER
Plantago lanceolata	narrowleaf plantain	L.	PLLA
Plantago tweedyi	Tweedy's plantain	Gray	PLTW

Genus/Species	Common name	Authority	Symbol
<i>Poa arnowiae</i>	Wasatch bluegrass	Soreng	POAR21
<i>Poa bolanderi</i>	Bolander's bluegrass	Vasey	POBO
<i>Poa bulbosa</i>	bulbous bluegrass	L.	POBU
<i>Poa cusickii</i>	Cusick's bluegrass	Vasey ssp. <i>epilis</i> (Scribn.) W.A. Weber	POCUE2
<i>Poa fendleriana</i>	muttongrass	(Steud.) Vasey	POFE
<i>Poa leptocoma</i>	marsh bluegrass	Trin.	POLE2
<i>Poa nervosa/wheeleri</i>	>>Poa wheeleri	(Hook.) Vasey var. <i>wheeleri</i> (Vasey) C.L. Hitchc.	POWH2
<i>Poa pratensis</i>	Kentucky bluegrass	L.	POPR
<i>Poa reflexa</i>	nodding bluegrass	Vasey & Scribn. ex Vasey	PORE
<i>Poa secunda</i>	Sandberg bluegrass	J. Presl	POSE
<i>Polemonium foliosissimum</i>	towering Jacob's-ladder	Gray	POFO
<i>Polygonum bistortoides</i>	American bistort	Pursh	POBI6
<i>Polygonum douglasii</i>	Douglas' knotweed	Greene	PODO4
<i>Polystichum lonchitis</i>	northern hollyfern	(L.) Roth	POLO4
<i>Populus tremuloides</i>	quaking aspen	Michx.	POTR5
<i>Potentilla diversifolia</i>	varileaf cinquefoil	Lehm.	PODI2
<i>Potentilla fruticosa</i>	>>Dasiphora floribunda	auct. non L.	POFR4
<i>Potentilla glandulosa</i>	sticky cinquefoil	Lindl.	POGL9
<i>Potentilla gracilis</i>	slender cinquefoil	Dougl. ex Hook.	POGR9
<i>Prunus virginiana</i>	chokecherry	L.	PRVI
<i>Pseudotsuga menziesii</i>	Douglas-fir	(Mirbel) Franco	PSME
<i>Pterospora andromedea</i>	woodland pinedrops	Nutt.	PTAN2
<i>Purshia tridentata</i>	antelope bitterbrush	(Pursh) DC.	PUTR2
<i>Pyrola asarifolia</i>	liverleaf wintergreen	Michx.	PYAS
<i>Pyrola secunda</i>	>>Orthilia secunda	L.	PYSE
<i>Ranunculus adoneus</i>	alpine buttercup	Gray	RAAD
<i>Ranunculus inamoenus</i>	graceful buttercup	Greene	RAIN
<i>Ranunculus jovis</i>	Utah buttercup	A. Nels.	RAJO
<i>Ranunculus macounii</i>	Macoun's buttercup	Britt.	RAMA2
<i>Ranunculus orthorhynchus</i>	straightbeak buttercup	Hook.	RAOR3
<i>Ribes hudsonianum</i>	northern black currant	Richardson	RIHU
<i>Ribes lacustre</i>	prickly currant	(Pers.) Poir.	RILA

Genus/Species	Common name	Authority	Symbol
<i>Ribes montigenum</i>	gooseberry currant	McClatchie	RIMO2
<i>Ribes oxycanthoides</i>	Canadian gooseberry	L.	RIOX
<i>Rosa nutkana</i>	Nootka rose	K. Presl	RONU
<i>Rosa woodsii</i>	Woods' rose	Lindl.	ROWO
<i>Rubus idaeus</i>	American red raspberry	L.	RUID
<i>Rubus parviflorus</i>	thimbleberry	Nutt.	RUPA
<i>Rudbeckia occidentalis</i>	western coneflower	Nutt.	RUOC2
<i>Salix bebbiana</i>	Bebb willow	Sarg.	SABE2
<i>Salix boothii</i>	Booth's willow	Dorn	SABO2
<i>Salix drummondiana</i>	Drummond's willow	Barratt ex Hook.	SADR
<i>Salix exigua</i>	narrowleaf willow	Nutt.	SAEX
<i>Salix glauca</i>	grayleaf willow	L.	SAGL
<i>Salix lasiandra</i>	>>Salix lucida ssp. lasiandra	Benth.	SALA5
<i>Salix lutea</i>	yellow willow	Nutt.	SALU2
<i>Salix scouleriana</i>	Scouler's willow	Barratt ex Hook.	SASC
<i>Salix wolfii</i>	Wolf's willow	Bebb	SAWO
<i>Sambucus cerulea</i>	>>Sambucus nigra ssp. cerulea	Raf.	SACE3
<i>Saxifraga odontoloma</i>	brook saxifrage	Piper	SAOD2
<i>Scrophularia lanceolata</i>	lanceleaf figwort	Pursh	SCLA
<i>Sedum debile</i>	orpine stonecrop	S. Wats.	SEDE
<i>Sedum lanceolatum</i>	spearleaf stonecrop	Torr.	SELA
<i>Senecio crassulus</i>	thickleaf ragwort	Gray	SECR
<i>Senecio integerrimus</i>	lambstongue ragwort	Nutt.	SEIN2
<i>Senecio multilobatus</i>	>>Packera multilobata	Torr. & Gray ex Gray	SEMU3
<i>Senecio serra</i>	tall ragwort	Hook.	SESE2
<i>Senecio streptanthifolius</i>	>>Packera streptanthifolia	Greene	SEST3
<i>Senecio triangularis</i>	arrowleaf ragwort	Hook.	SETR
<i>Shepherdia canadensis</i>	russet buffaloberry	(L.) Nutt.	SHCA
<i>Sidalcea oregana</i>	Oregon checkerbloom	(Nutt. ex Torr. & Gray) Gray	SIOR
<i>Silene menziesii</i>	Menzies' campion	Hook.	SIME
<i>Silene oregana</i>	Oregon silene	S. Watson	SIOR3
<i>Smilacina racemosa</i>	>>Maianthemum racemosum ssp. amplexicaule	(L.) Desf. var. amplexicaulis (Nutt.) S. Wats.	SMRAA

Genus/Species	Common name	Authority	Symbol
<i>Smilacina stellata</i>	>>Maianthemum stellatum	(L.) Desf.	SMST
<i>Solidago multiradiata</i>	Rocky Mountain goldenrod	Ait.	SOMU
<i>Solidago nana</i>	baby goldenrod	Nutt.	SONA
<i>Sorbus scopulina</i>	Greene's mountain ash	Greene	SOSC2
<i>Stellaria jamesiana</i>	>>Pseudostellaria jamesiana	Torr.	STJA3
<i>Stellaria nitens</i>	shiny chickweed	Nutt.	STNI
<i>Stipa lettermanii</i>	>>Achnatherum lettermanii	Vasey	STLE4
<i>Stipa nelsonii</i>	>>Achnatherum nelsonii ssp. nelsonii	Scribn.	STNE3
<i>Symphoricarpos oreophilus</i>	mountain snowberry	Gray	SYOR2
<i>Synthyris pinnatifida</i>	featherleaf kittentails	S. Wats.	SYPI
<i>Taraxacum officinale</i>	common dandelion	G.H. Weber ex Wiggers	TAOF
<i>Thalictrum fendleri</i>	Fendler's meadow-rue	Engelm. ex Gray	THFE
<i>Thlaspi montanum</i>	alpine pennycress	L.	THMO5
<i>Tragopogon dubius</i>	yellow salsify	Scop.	TRDU
<i>Trifolium longipes</i>	longstalk clover	Nutt.	TRLO
<i>Trifolium pratense</i>	red clover	L.	TRPR2
<i>Trifolium repens</i>	white clover	L.	TRRE3
<i>Trisetum spicatum</i>	spike trisetum	(L.) K. Richt.	TRSP2
<i>Triteleia grandiflora</i>	largeflower triteleia	Lindl.	TRGR7
<i>Urtica dioica</i>	stinging nettle	L.	URDI
<i>Vaccinium membranaceum</i>	thinleaf huckleberry	Dougl. ex Torr.	VAME
<i>Valeriana edulis</i>	tobacco root	Nutt. ex Torr. & Gray	VAED
<i>Valeriana occidentalis</i>	western valerian	Heller	VAOC2
<i>Veratrum californicum</i>	California false hellebore	Dur.	VECA2
<i>Verbascum thapsus</i>	common mullein	L.	VETH
<i>Veronica biloba</i>	twolobe speedwell	L.	VEBI2
<i>Veronica serpyllifolia</i>	thymeleaf speedwell	L.	VESE
<i>Vicia americana</i>	American vetch	Muhl. ex Willd.	VIAM
<i>Viguiera multiflora</i>	showy goldeneye	(Nutt.) S.F. Blake var. <i>nevadensis</i> (A. Nelson) S.F. Blake	VIMUN
<i>Viola adunca</i>	hookedspur violet	Sm.	VIAD
<i>Viola nephrophylla</i>	northern bog violet	Greene	VINE
<i>Viola nuttallii/vallicola</i>	>>Viola vallicola var. vallicola	Pursh ssp. vallicola (A. Nels.) Taylor & MacBryde	VINUV2

Genus/Species	Common name	Authority	Symbol
Viola purpurea	goosefoot violet	Kellogg	VIPU4
Wyethia amplexicaulis	mule-ears	(Nutt.) Nutt.	WYAM
Zigadenus elegans	mountain deathcamas	Pursh	ZIEL2
Zigadenus paniculatus	foothill deathcamas	(Nutt.) S. Wats.	ZIPA2

APPENDIX C
SYNOPTIC TABLE

A part (the first ten vegetation units) of the advanced combined synoptic table with phi coefficient (Fisher's exact test, $p < 0.01$) and constancy (frequency) percentage. The vegetation units are in columns, each row represents one species. The last three rows contain average positive fidelity, No. of faithful species with $\phi \geq 40$ and species richness for each vegetation unit

Group No. No. of relevés	1 3	2 4	3 6	4 3	5 5	6 5	7 5	8 6	9 5	10 3
<i>Galium boreale</i>	62.5 ¹⁰⁰	---	---	---	---	---	---	---	---	---
<i>Potentilla gracilis</i>	49.9 ¹⁰⁰	---	---	---	---	---	---	---	---	---
<i>Epilobium ciliatum</i>	48.5 ⁶⁷	---	---	---	---	---	---	---	---	---
<i>Potentilla fruticosa</i>	47.5 ⁶⁷	---	---	---	---	---	---	---	---	---
<i>Stellaria nitens</i>	44.2 ¹⁰⁰	---	18.7 ⁵⁰	---	---	---	---	---	---	---
<i>Salix bebbiana</i>	42.4 ³³	---	---	---	---	---	---	---	---	---
<i>Mertensia ciliata</i>	40.5 ¹⁰⁰	---	16.5 ⁵⁰	---	---	---	---	---	---	---
<i>Zigadenus elegans</i>	---	86.2 ¹⁰⁰	---	---	---	---	---	---	---	---
<i>Aconitum columbianum</i>	---	69.6 ¹⁰⁰	20.3 ³³	---	---	---	---	---	---	---
<i>Senecio triangularis</i>	38.1 ⁶⁷	59.7 ¹⁰⁰	38.1 ⁶⁷	---	---	---	---	---	---	---
<i>Arnica amplexicaulis</i>	---	58.8 ⁵⁰	---	---	---	---	---	---	---	---
<i>Castilleja rhexifolia</i>	---	43.3 ⁷⁵	---	37.9 ⁶⁷	---	---	---	27.2 ⁵⁰	---	---
<i>Ranunculus jovis</i>	---	---	---	70.8 ⁶⁷	---	---	---	---	---	---
<i>Senecio crassulus</i>	---	---	---	42.6 ¹⁰⁰	32.6 ⁸⁰	---	---	---	---	---
<i>Ribes montigenum</i>	---	---	16.9 ⁵⁰	41.2 ¹⁰⁰	---	---	---	25.0 ⁶⁷	---	---
<i>Ligusticum porteri</i>	---	---	---	---	89.2 ⁸⁰	---	---	---	---	---
<i>Poa bolanderi</i>	---	---	---	---	81.0 ¹⁰⁰	---	---	---	---	---
<i>Rubus parviflorus</i>	---	---	---	---	---	---	77.0 ⁶⁰	---	---	---
<i>Goodyera oblongifolia</i>	---	---	---	---	---	---	64.1 ⁶⁰	---	---	---
<i>Sorbus scopulina</i>	---	---	---	---	24.0 ⁴⁰	---	52.5 ⁸⁰	---	---	---
<i>Chimaphila umbellata</i>	---	---	---	---	---	---	50.6 ⁴⁰	---	---	---
<i>Rubus idaeus</i>	---	---	---	---	---	---	---	70.2 ⁵⁰	---	---
<i>Carex geyeri</i>	---	---	---	---	---	---	32.0 ⁶⁰	56.8 ¹⁰⁰	---	---
<i>Poa cusickii</i>	---	---	---	---	---	---	---	53.5 ⁵⁰	---	---
<i>Thlaspi montanum</i>	---	---	---	38.8 ⁶⁷	---	---	---	---	47.6 ⁸⁰	---
<i>Calochortus nuttallii</i>	---	---	---	---	---	---	---	---	---	81.0 ¹⁰⁰
<i>Juniperus scopulorum</i>	---	---	---	---	---	---	---	---	---	76.7 ¹⁰⁰
<i>Lomatium grayi</i>	---	---	---	---	---	---	---	---	---	65.7 ⁶⁷
<i>Shepherdia canadensis</i>	---	---	---	---	---	---	36.0 ⁶⁰	---	22.4 ⁴⁰	---
<i>Lupinus argenteus</i>	---	---	---	27.3 ⁶⁷	34.1 ⁸⁰	---	---	---	---	---
<i>Carex pachystachya</i>	37.4 ⁶⁷	---	---	---	---	---	---	---	---	---
<i>Urtica dioica</i>	34.4 ⁶⁷	---	---	---	---	18.4 ⁴⁰	---	---	---	---
<i>Erigeron eatonii</i>	---	---	---	32.1 ⁶⁷	---	---	---	---	28.3 ⁶⁰	---
<i>Arenaria congesta</i>	---	---	---	---	---	---	---	33.7 ⁶⁷	---	---
<i>Cystopteris fragilis</i>	---	---	---	---	---	---	---	36.8 ⁵⁰	---	---
<i>Pyrola secunda</i>	---	---	---	50.5 ¹⁰⁰	27.9 ⁶⁰	---	50.5 ¹⁰⁰	---	---	---
<i>Angelica arguta</i>	45.0 ⁶⁷	69.6 ¹⁰⁰	---	---	---	---	---	---	---	---
<i>Salix wolfii</i>	---	61.7 ⁷⁵	---	---	---	---	---	---	---	---
<i>Pedicularis groenlandica</i>	---	61.7 ⁷⁵	---	---	---	---	---	---	---	---
<i>Habenaria dilatata</i>	---	54.3 ¹⁰⁰	---	---	---	---	---	---	---	---
<i>Aster occidentalis</i>	---	46.4 ¹⁰⁰	---	46.4 ¹⁰⁰	---	---	---	---	---	---
<i>Mimulus guttatus</i>	---	40.6 ⁷⁵	---	---	---	---	---	---	---	---
<i>Juncus ensifolius</i>	---	39.9 ⁵⁰	---	---	---	---	---	---	---	---
<i>Lonicera utahensis</i>	---	---	---	57.0 ¹⁰⁰	32.1 ⁶⁰	---	44.5 ⁸⁰	---	---	---

Group No. No. of relevés	1 3	2 4	3 6	4 3	5 5	6 5	7 5	8 6	9 5	10 3
<i>Arnica latifolia</i>	---	---	--- ³³	51.8 ¹⁰⁰	---	---	40.3 ⁸⁰	32.6 ⁶⁷	---	---
<i>Hieracium albiflorum</i>	---	---	---	---	64.1 ¹⁰⁰	--- ²⁰	50.4 ⁸⁰	---	---	---
<i>Arnica cordifolia</i>	---	---	27.0 ⁶⁷	--- ⁶⁷	23.7 ⁶⁰	---	44.0 ¹⁰⁰	27.0 ⁶⁷	---	---
<i>Juncus parryi</i>	---	---	---	---	---	---	---	63.9 ⁸³	---	---
<i>Anemone multifida</i>	---	---	---	---	---	---	---	---	49.8 ⁸⁰	---
<i>Artemisia arbuscula</i>	---	---	---	---	---	---	---	---	---	47.9 ¹⁰⁰
<i>Zigadenus paniculatus</i>	---	---	---	---	---	---	---	---	---	45.0 ⁶⁷
<i>Ceanothus velutinus</i>	---	---	---	---	---	---	---	---	---	43.9 ⁶⁷
<i>Equisetum arvense</i>	34.4 ⁶⁷	---	---	---	---	---	---	---	---	---
<i>Lonicera involucrata</i>	42.8 ¹⁰⁰	42.8 ¹⁰⁰	26.1 ⁶⁷	---	---	---	---	---	---	---
<i>Taraxacum officinale</i>	41.3 ¹⁰⁰	---	---	---	---	---	---	---	---	---
<i>Lathyrus pauciflorus</i>	---	---	---	---	---	--- ²⁰	---	---	---	---
<i>Salix drummondiana</i>	27.8 ⁶⁷	32.1 ⁷⁵	---	---	---	---	---	---	---	---
<i>Veratrum californicum</i>	---	42.8 ¹⁰⁰	---	---	---	---	---	---	---	---
<i>Symphoricarpos oreophilus</i>	---	---	---	---	---	---	---	---	---	---
<i>Osmorhiza chilensis</i>	---	---	18.8 ¹⁰⁰	---	18.8 ¹⁰⁰	18.8 ¹⁰⁰	---	---	---	---
<i>Berberis repens</i>	---	---	---	---	---	---	---	---	---	28.3 ¹⁰⁰
<i>Pedicularis racemosa</i>	---	26.7 ¹⁰⁰	26.7 ¹⁰⁰	26.7 ¹⁰⁰	26.7 ¹⁰⁰	---	26.7 ¹⁰⁰	20.4 ⁸³	26.7 ¹⁰⁰	---
<i>Thalictrum fendleri</i>	---	---	18.0 ¹⁰⁰	---	---	18.0 ¹⁰⁰	---	---	18.0 ¹⁰⁰	---
<i>Paxistima myrsinites</i>	---	---	18.0 ⁸³	24.1 ¹⁰⁰	---	---	24.1 ¹⁰⁰	24.1 ¹⁰⁰	---	---
Average positive fidelity	40.4	44.4	23.9	40.7	33.0	26.2	37.7	30.4	28.9	47.3
No. of faithful spp, phi≥40	11	14	0	7	3	0	9	4	2	8
Species richness	74	63	71	53	57	72	56	90	49	30

APPENDIX D

ANALYSIS OF THE SYNOPTIC TABLE

The list of diagnostic species for forest vegetation units (1-20) and non-forest vegetation units (21-34). Threshold values: faithful species, $\phi \geq 40$; constant species, frequency ≥ 60 %; dominant species, cover ≥ 5 %. A digit after a dominant species indicates percentage of relevés in which the species occurs with cover ≥ 5 %. Brief habitat characteristics are in Table 4.1, 4.2.

Forest vegetation units

Unit No. 1

Number of relevés: 3

Faithful species: *Galium boreale* 62.5, *Potentilla gracilis* 49.9, *Epilobium ciliatum* 48.5, *Potentilla fruticosa* 47.5, *Angelica arguta* 45.0, *Stellaria nitens* 44.2, *Lonicera involucrata* 42.8, *Salix bebbiana* 42.4, *Taraxacum officinale* 41.3, *Salix boothii* 41.1, *Mertensia ciliata* 40.5

Constant species: *Abies lasiocarpa* 100, *Thalictrum fendleri* 100, *Symphoricarpos oreophilus* 100, *Rudbeckia occidentalis* 100, *Picea engelmannii* 100, *Osmorhiza chilensis* 100, *Geranium richardsonii* 100, *Urtica dioica* 67, *Stellaria jamesiana* 67, *Smilacina stellata* 67, *Senecio triangularis* 67, *Senecio serra* 67, *Salix drummondiana* 67, *Rosa woodsii* 67, *Ribes lacustre* 67, *Polemonium foliosissimum* 67, *Pedicularis racemosa* 67, *Mitella pentandra* 67, *Geum macrophyllum* 67, *Equisetum arvense* 67, *Elymus glaucus* 67, *Collinsia parviflora* 67, *Carex pachystachya* 67, *Aster engelmannii* 67, *Agastache urticifolia* 67, *Achillea millefolium* 67

Dominant species: *Picea engelmannii* 100, *Abies lasiocarpa* 67, *Salix boothii* 33, *Ribes lacustre* 33, *Populus tremuloides* 33, *Osmorhiza chilensis* 33, *Equisetum arvense* 33, *Cornus sericea* 33

Group No. 2

Number of relevés: 4

Faithful species: *Zigadenus elegans* 86.2, *Angelica arguta* 69.6, *Aconitum columbianum* 69.6, *Salix wolfii* 61.7, *Pedicularis groenlandica* 61.7, *Senecio triangularis* 59.7, *Arnica amplexicaulis* 58.8, *Habenaria dilatata* 54.3, *Aster occidentalis* 46.4, *Castilleja rhexifolia* 43.3, *Veratrum californicum* 42.8, *Lonicera involucrata* 42.8, *Salix boothii* 41.1, *Mimulus guttatus* 40.6, *Juncus ensifolius* 39.9

Constant species: *Abies lasiocarpa* 100, *Picea engelmannii* 100, *Pedicularis racemosa* 100, *Geranium richardsonii* 100, *Thalictrum fendleri* 75, *Salix drummondiana* 75, *Ligusticum filicinum* 75, *Frasera speciosa* 75, *Achillea millefolium* 75

Dominant species: *Picea engelmannii* 100, *Abies lasiocarpa* 75, *Veratrum californicum* 50, *Salix wolfii* 50, *Geranium richardsonii* 50, *Arnica amplexicaulis* 50, *Senecio triangularis* 25, *Salix boothii* 25, *Carex microptera* 25

Group No. 3

Number of relevés: 6

Faithful species:

Constant species: *Abies lasiocarpa* 100, *Thalictrum fendleri* 100, *Stellaria jamesiana* 100, *Picea engelmannii* 100, *Pedicularis racemosa* 100, *Osmorhiza chilensis* 100, *Hackelia micrantha* 100, *Geranium viscosissimum* 100, *Erythronium grandiflorum* 100, *Aster engelmannii* 100, *Aquilegia caerulea* 100, *Rudbeckia occidentalis* 83, *Paxistima myrsinites* 83, *Ligusticum filicinum* 83, *Hydrophyllum capitatum* 83, *Carex rossii* 83, *Achillea millefolium* 83, *Senecio triangularis* 67, *Potentilla glandulosa* 67, *Lonicera involucrata* 67, *Geranium richardsonii* 67, *Bromus carinatus* 67, *Arnica cordifolia* 67

Dominant species: *Abies lasiocarpa* 100, *Picea engelmannii* 100, *Rudbeckia occidentalis* 50, *Arnica cordifolia* 33, *Thalictrum fendleri* 17, *Pedicularis racemosa* 17, *Ligusticum filicinum* 17

Group No. 4

Number of relevés: 3

Faithful species: *Ranunculus jovis* 70.8, *Lonicera utahensis* 57.0, *Arnica latifolia* 51.8, *Pyrola secunda* 50.5, *Aster occidentalis* 46.4, *Senecio crassulus* 42.6, *Ribes montigenum* 41.2

Constant species: *Abies lasiocarpa* 100, *Thalictrum fendleri* 100, *Stellaria jamesiana* 100, *Picea engelmannii* 100, *Pedicularis racemosa* 100, *Paxistima myrsinites* 100, *Osmorhiza chilensis* 100, *Erythronium grandiflorum* 100, *Aster engelmannii* 100, *Aquilegia caerulea* 100, *Valeriana occidentalis* 67, *Thlaspi montanum* 67, *Sedum debile* 67, *Lupinus argenteus* 67, *Ligusticum filicinum* 67, *Geranium viscosissimum* 67, *Frasera speciosa* 67, *Erigeron eatonii* 67, *Claytonia lanceolata* 67, *Castilleja rhexifolia* 67, *Carex rossii* 67, *Arnica cordifolia* 67, *Achillea millefolium* 67

Dominant species: *Picea engelmannii* 100, *Arnica latifolia* 67, *Abies lasiocarpa* 33

Group No. 5

Number of relevés: 5

Faithful species: *Ligusticum porteri* 89.2, *Poa bolanderi* 81.0, *Hieracium albiflorum* 64.1

Constant species: *Abies lasiocarpa* 100, *Viola nuttallii/vallicola* 100, *Stellaria jamesiana* 100, *Populus tremuloides* 100, *Picea engelmannii* 100, *Pedicularis racemosa* 100, *Osmorhiza chilensis* 100, *Ligusticum filicinum* 100, *Hydrophyllum capitatum* 100, *Erigeron speciosus* 100, *Carex rossii* 100, *Aster engelmannii* 100, *Achillea millefolium* 100, *Abies lasiocarpa* 100, *Senecio serra* 80, *Senecio crassulus* 80, *Pseudotsuga menziesii* 80, *Lupinus argenteus* 80, *Hackelia micrantha* 80, *Claytonia lanceolata* 80, *Aquilegia caerulea* 80, *Symphoricarpos oreophilus* 60, *Pyrola secunda* 60, *Potentilla glandulosa* 60, *Paxistima myrsinites* 60, *Osmorhiza occidentalis* 60, *Lonicera utahensis* 60, *Bromus carinatus* 60, *Arnica cordifolia* 60

Dominant species: *Abies lasiocarpa* 100, *Picea engelmannii* 100, *Osmorhiza chilensis* 100, *Ligusticum filicinum* 40, *Populus tremuloides* 20, *Pedicularis racemosa* 20

Group No. 6

Number of relevés: 5

Faithful species:

Constant species: *Abies lasiocarpa* 100, *Thalictrum fendleri* 100, *Rudbeckia occidentalis* 100, *Populus tremuloides* 100, *Osmorhiza chilensis* 100, *Hydrophyllum capitatum* 100, *Hackelia micrantha* 100, *Delphinium nuttallianum* 100, *Carex rossii* 100, *Agastache urticifolia* 100, *Viola nuttallii/vallicola* 80, *Symphoricarpos oreophilus* 80, *Stellaria nitens* 80, *Stellaria jamesiana* 80, *Smilacina stellata* 80, *Potentilla glandulosa* 80, *Polemonium foliosissimum* 80, *Picea engelmannii* 80, *Phacelia hastata* 80, *Geranium viscosissimum* 80, *Erythronium grandiflorum* 80, *Elymus glaucus* 80, *Collomia linearis* 80, *Bromus carinatus* 80, *Aster engelmannii* 80, *Achillea millefolium* 80, *Senecio serra* 60, *Ribes lacustre* 60, *Pseudotsuga menziesii* 60, *Paxistima myrsinites* 60, *Galium bifolium* 60, *Erigeron speciosus* 60, *Carex hoodii* 60

Dominant species: *Abies lasiocarpa* 100, *Populus tremuloides* 100, *Osmorhiza chilensis* 80, *Picea engelmannii* 60, *Rudbeckia occidentalis* 40, *Pseudotsuga menziesii* 20, *Thalictrum fendleri* 20

Group No. 7

Number of relevés: 5

Faithful species: *Rubus parviflorus* 77.0, *Goodyera oblongifolia* 64.1, *Sorbus scopulina* 52.5, *Chimaphila umbellata* 50.6, *Pyrola secunda* 50.5, *Hieracium albiflorum* 50.4, *Lonicera utahensis* 44.5, *Arnica cordifolia* 44.0, *Arnica latifolia* 40.3

Constant species: *Abies lasiocarpa* 100, *Pseudotsuga menziesii* 100, *Picea engelmannii* 100, *Pedicularis racemosa* 100, *Paxistima myrsinites* 100, *Aster engelmannii* 100, *Osmorhiza chilensis* 80, *Frasera speciosa* 80, *Aquilegia caerulea* 80, *Thalictrum fendleri* 60, *Shepherdia canadensis* 60, *Sambucus cerulea* 60, *Pinus flexilis* 60, *Mitella pentandra* 60, *Carex rossii* 60, *Carex geyeri* 60

Dominant species: *Pseudotsuga menziesii* 100, *Abies lasiocarpa* 100, *Picea engelmannii* 60, *Carex geyeri* 40, *Arnica cordifolia* 40, *Pyrola secunda* 20, *Osmorhiza chilensis* 20, *Arnica latifolia* 20

Group No. 8

Number of relevés: 6

Faithful species: *Rubus idaeus* 70.2, *Juncus parryi* 63.9, *Carex geyeri* 56.8, *Poa cusickii* 53.5

Constant species: *Abies lasiocarpa* 100, *Picea engelmannii* 100, *Penstemon leonardii* 100, *Paxistima myrsinites* 100, *Aster engelmannii* 100, *Achillea millefolium* 100, *Carex geyeri* 100, *Symphoricarpos oreophilus* 83, *Pedicularis racemosa* 83, *Leucopoa kingii* 83, *Stellaria jamesiana* 67, *Ribes montigenum* 67, *Potentilla glandulosa* 67, *Populus tremuloides* 67, *Hieracium cynoglossoides* 67, *Arnica latifolia* 67, *Arnica cordifolia* 67, *Arenaria congesta* 6, *Pseudotsuga menziesii* 67

Dominant species: *Picea engelmannii* 100, *Carex geyeri* 67, *Abies lasiocarpa* 67, *Carex rossii* 17

Group No. 9

Number of relevés: 5

Faithful species: *Anemone multifida* 49.8, *Thlaspi montanum* 47.6

Constant species: *Abies lasiocarpa* 100, *Thalictrum fendleri* 100, *Pinus flexilis* 100, *Picea engelmannii* 100, *Penstemon leonardii* 100, *Pedicularis racemosa* 100, *Lomatium graveolens* 100, *Erythronium grandiflorum* 100, *Aster engelmannii* 100, *Ribes lacustre* 80, *Poa fendleriana* 80, *Leucopoa kingii* 80, *Claytonia lanceolata* 80, *Pseudotsuga menziesii* 60, *Paxistima myrsinites* 60, *Osmorhiza chilensis* 60, *Frasera speciosa* 60, *Erigeron eatonii* 60, *Castilleja miniata* 60, *Berberis repens* 60

Dominant species: *Picea engelmannii* 100, *Pinus flexilis* 80, *Pseudotsuga menziesii* 40, *Pedicularis racemosa* 20, *Aster engelmannii* 20

Group No. 10

Number of relevés: 3

Faithful species: *Calochortus nuttallii* 81.0, *Juniperus scopulorum* 76.7, *Lomatium grayi* 65.7, *Artemisia arbuscula* 47.9, *Zigadenus paniculatus* 45.0, *Ceanothus velutinus* 43.9, *Artemisia tridentata* 42.4, *Comandra umbellata* 41.0

Constant species: *Juniperus scopulorum* 100, *Elymus spicatus* 100, *Berberis repens* 100, *Balsamorhiza sagittata* 100, *Symphoricarpos oreophilus* 67, *Senecio integerrimus* 67, *Sambucus cerulea* 67, *Penstemon leonardii* 67, *Lomatium graveolens* 67, *Crepis acuminata* 67

Dominant species: *Artemisia arbuscula* 100, *Juniperus scopulorum* 67, *Elymus spicatus* 67, *Balsamorhiza sagittata* 33

Group No. 11

Number of relevés: 3

Faithful species: *Cercocarpus ledifolius* 66.7, *Mertensia oblongifolia* 60.5, *Crepis acuminata* 59.7, *Viola purpurea* 59.6, *Chrysothamnus viscidiflorus* 58.3, *Ceanothus velutinus* 43.9, *Artemisia tridentata* 42.4

Constant species: *Cercocarpus ledifolius* 100, *Symphoricarpos oreophilus* 100, *Elymus spicatus* 100, *Berberis repens* 100, *Balsamorhiza sagittata* 100, *Achillea millefolium* 100, *Thalictrum fendleri* 67, *Senecio integerrimus* 67, *Pseudotsuga menziesii* 67, *Prunus virginiana* 67, *Pinus flexilis* 67, *Penstemon leonardii* 67, *Paxistima myrsinites* 67, *Helianthella uniflora* 67, *Aster perelegans* 67

Dominant species: *Symphoricarpos oreophilus* 100, *Cercocarpus ledifolius* 100, *Balsamorhiza sagittata* 100, *Elymus spicatus* 67, *Chrysothamnus viscidiflorus* 33, *Aster perelegans* 33, *Artemisia tridentata* 33

Group No. 12

Number of relevés: 4

Faithful species: *Petradoria pumila* 86.3, *Cercocarpus ledifolius* 66.7, *Solidago nana* 65.1, *Aster ascendens* 49.9, *Artemisia tridentata* 42.4, *Comandra umbellata* 41.0

Constant species: *Cercocarpus ledifolius* 100, *Symphoricarpos oreophilus* 100, *Leucopoa kingii* 100, *Eriogonum umbellatum* 100, *Elymus spicatus* 100, *Berberis repens* 100, *Sedum debile* 75, *Pseudotsuga menziesii* 75, *Pinus flexilis* 75, *Penstemon leonardii* 75, *Machaeranthera commixta* 75, *Erigeron speciosus* 75, *Balsamorhiza sagittata* 75

Dominant species: *Cercocarpus ledifolius* 100, *Elymus spicatus* 75, *Artemisia arbuscula* 50, *Solidago nana* 25, *Prunus virginiana* 25, *Berberis repens* 25, *Artemisia tridentata* 25

Group No. 13

Number of relevés: 6

Faithful species: *Acer grandidentatum* 53.0, *Arnica cordifolia* 44.0, *Smilacina racemosa* 42.1, *Viola adunca* 40.2

Constant species: *Abies lasiocarpa* 100, *Thalictrum fendleri* 100, *Symphoricarpos oreophilus* 100, *Stellaria jamesiana* 100, *Pseudotsuga menziesii* 100, *Osmorhiza chilensis* 100, *Hydrophyllum capitatum* 100, *Hackelia micrantha* 100, *Elymus glaucus* 100, *Agastache urticifolia* 100, *Stellaria nitens* 83, *Carex rossii* 83, *Berberis repens* 83, *Viola nuttallii/vallicola* 67, *Mitella pentandra* 67, *Collinsia parviflora* 67, *Claytonia lanceolata* 67, *Aster engelmannii* 67, *Amelanchier alnifolia* 67

Dominant species: *Pseudotsuga menziesii* 100, *Thalictrum fendleri* 83, *Osmorhiza chilensis* 83, *Abies lasiocarpa* 50, *Berberis repens* 33, *Symphoricarpos oreophilus* 17, *Populus tremuloides* 17, *Elymus glaucus* 17, *Arnica cordifolia* 17

Group No. 14

Number of relevés: 5

Faithful species: *Linanthastrum nuttallii* 51.9, *Aster glaucodes* 47.9, *Acer glabrum* 46.5

Constant species: *Symphoricarpos oreophilus* 100, *Pseudotsuga menziesii* 100, *Pinus flexilis* 100, *Paxistima myrsinites* 100, *Berberis repens* 100, *Pedicularis racemosa* 80, *Leucopoa kingii* 80, *Aster engelmannii* 80, *Abies lasiocarpa* 60, *Picea engelmannii* 60, *Juniperus communis* 60

Dominant species: *Pseudotsuga menziesii* 100, *Pinus flexilis* 100, *Berberis repens* 60, *Linanthastrum nuttallii* 40, *Juniperus communis* 40, *Leucopoa kingii* 20, *Aster perelegans* 20, *Aster engelmannii* 20, *Abies lasiocarpa* 20

Group No. 15

Number of relevés: 3

Faithful species: *Astragalus tenellus* 57.2, *Bromus anomalus* 44.5, *Shepherdia canadensis* 40.5

Constant species: *Thalictrum fendleri* 100, *Symphoricarpos oreophilus* 100, *Pseudotsuga menziesii* 100, *Pinus flexilis* 100, *Phacelia hastata* 100, *Leucopoa kingii* 100, *Helianthella uniflora* 100, *Erigeron speciosus* 100, *Aster engelmannii* 100, *Stipa nelsonii* 67, *Stipa lettermanii* 67, *Stellaria jamesiana* 67, *Ribes montigenum* 67, *Penstemon leonardii* 67, *Penstemon cyananthus* 67, *Pedicularis racemosa* 67, *Osmorhiza chilensis* 67, *Machaeranthera commixta* 67, *Lomatium graveolens* 67, *Hackelia micrantha* 67, *Geranium viscosissimum* 67, *Frasera speciosa* 67, *Castilleja applegatei* 67, *Berberis repens* 67, *Balsamorhiza sagittata* 67, *Aster glaucodes* 67, *Achillea millefolium* 67

Dominant species: *Pseudotsuga menziesii* 100, *Pinus flexilis* 67, *Lomatium graveolens* 67, *Symphoricarpos oreophilus* 33, *Leucopoa kingii* 33, *Eriogonum umbellatum* 33, *Balsamorhiza sagittata* 33, *Aster engelmannii* 33

Group No. 16

Number of relevés: 4

Faithful species: *Poa leptocoma* 49.4, *Lupinus argenteus* 44.4, *Artemisia ludoviciana* 43.8

Constant species: *Abies lasiocarpa* 100, *Symphoricarpos oreophilus* 100, *Stellaria jamesiana* 100, *Populus tremuloides* 100, *Picea engelmannii* 100, *Osmorhiza occidentalis* 100, *Hydrophyllum capitatum* 100, *Helianthella uniflora* 100, *Hackelia micrantha* 100, *Erigeron speciosus* 100, *Bromus carinatus* 100, *Aster engelmannii* 100, *Senecio serra* 75, *Phacelia hastata* 75, *Ligusticum filicinum* 75, *Galium bifolium* 75, *Elymus trachycaulus* 75, *Carex hoodii* 75, *Artemisia spiciformis* 75, *Agastache urticifolia* 75, *Achillea millefolium* 75

Dominant species: *Populus tremuloides* 100, *Abies lasiocarpa* 50, *Symphoricarpos oreophilus* 25, *Pseudotsuga menziesii* 25, *Picea engelmannii* 25, *Carex rossii* 25, *Carex geyeri* 25, *Bromus carinatus* 25, *Artemisia ludoviciana* 25, *Agastache urticifolia* 25

Group No. 17

Number of relevés: 6

Faithful species: *Valeriana occidentalis* 45.5, *Tragopogon dubius* 41.6, *Lathyrus pauciflorus* 40.0

Constant species: *Thalictrum fendleri* 100, *Senecio serra* 100, *Rudbeckia occidentalis* 100, *Populus tremuloides* 100, *Osmorhiza occidentalis* 100, *Osmorhiza chilensis* 100, *Elymus glaucus* 100, *Delphinium occidentale* 100, *Bromus carinatus* 83, *Agastache urticifolia* 83, *Abies lasiocarpa* 83, *Viola nuttallii/vallicola* 83, *Symphoricarpos oreophilus* 83, *Hackelia micrantha* 83, *Erigeron speciosus* 83, *Carex hoodii* 83, *Achillea millefolium* 83, *Stellaria jamesiana* 67, *Sidalcea oregana* 67, *Sambucus cerulea* 67, *Mertensia ciliata* 67, *Heracleum lanatum* 67, *Galium bifolium* 67, *Collinsia parviflora* 67

Dominant species: *Rudbeckia occidentalis* 100, *Populus tremuloides* 100, *Thalictrum fendleri* 67, *Senecio serra* 67, *Abies lasiocarpa* 67, *Lathyrus pauciflorus* 50, *Bromus carinatus* 50, *Elymus glaucus* 33, *Agastache urticifolia* 33, *Veratrum californicum* 17, *Osmorhiza occidentalis* 17, *Osmorhiza chilensis* 17, *Mertensia ciliata* 17, *Lathyrus lanszwertii* 17, *Heracleum lanatum* 17

Group No. 18

Number of relevés: 10

Faithful species:

Constant species: *Thalictrum fendleri* 100, *Symphoricarpos oreophilus* 100, *Senecio serra* 100, *Rudbeckia occidentalis* 100, *Populus tremuloides* 100, *Hackelia micrantha* 100, *Elymus glaucus* 100, *Carex hoodii* 100, *Agastache urticifolia* 100, *Osmorhiza chilensis* 90, *Delphinium occidentale* 90, *Bromus carinatus* 90, *Aster engelmannii* 90, *Achillea millefolium* 90, *Osmorhiza occidentalis* 80, *Lathyrus pauciflorus* 80, *Abies lasiocarpa* 70, *Stellaria jamesiana* 60, *Phacelia hastata* 60, *Nemophila breviflora* 60, *Galium bifolium* 60, *Elymus trachycaulus* 60, *Delphinium nuttallianum* 60, *Collomia linearis* 60

Dominant species: *Populus tremuloides* 100, *Symphoricarpos oreophilus* 80, *Senecio serra* 80, *Thalictrum fendleri* 60, *Agastache urticifolia* 60, *Rudbeckia occidentalis* 50, *Lathyrus pauciflorus* 40, *Elymus glaucus* 40, *Bromus carinatus* 40, *Carex hoodii* 20, *Elymus trachycaulus* 10

Group No. 19

Number of relevés: 4

Faithful species: *Scrophularia lanceolata* 55.0, *Valeriana occidentalis* 45.5, *Lathyrus pauciflorus* 40.0

Constant species: *Thalictrum fendleri* 100, *Symphoricarpos oreophilus* 100, *Senecio serra* 100, *Rudbeckia occidentalis* 100, *Populus tremuloides* 100, *Phacelia hastata* 100, *Osmorhiza occidentalis* 100, *Osmorhiza chilensis* 100, *Geranium viscosissimum* 100, *Delphinium occidentale* 100, *Bromus carinatus* 100, *Agastache urticifolia* 100, *Galium bifolium* 75, *Elymus trachycaulus* 75, *Elymus glaucus* 75, *Carex hoodii* 75

Dominant species: *Symphoricarpos oreophilus* 100, *Populus tremuloides* 100, *Rudbeckia occidentalis* 50, *Bromus carinatus* 50, *Senecio serra* 25, *Mertensia ciliata* 25, *Agastache urticifolia* 25, *Abies lasiocarpa* 25

Group No. 20

Number of relevés: 2

Faithful species: *Cynoglossum officinale* 70.9, *Dactylis glomerata* 63.8, *Allium bisceptrum* 63.8, *Ranunculus orthorhynchus* 61.6, *Bromus ciliatus* 60.3, *Carex pachystachya* 58.7, *Elymus cinereus* 55.1, *Sidalcea oregana* 55.0, *Urtica dioica* 54.3, *Arnica chamissonis* 54.3, *Geum macrophyllum* 51.4, *Phleum pratense* 48.5, *Smilacina stellata* 46.7, *Trifolium repens* 46.4, *Veratrum californicum* 42.8, *Poa reflexa* 41.9, *Taraxacum officinale* 41.3, *Lathyrus pauciflorus* 40.0

Constant species: *Thalictrum fendleri* 100, *Symphoricarpos oreophilus* 100, *Senecio serra* 100, *Rudbeckia occidentalis* 100, *Populus tremuloides* 100, *Polemonium foliosissimum* 100, *Osmorhiza occidentalis* 100, *Osmorhiza chilensis* 100, *Hackelia micrantha* 100, *Elymus glaucus* 100, *Delphinium occidentale* 100, *Carex hoodii* 100, *Agastache urticifolia* 100, *Achillea millefolium* 100, *Veratrum californicum* 100, *Carex pachystachya* 100, *Elymus cinereus* 100

Dominant species: *Thalictrum fendleri* 100, *Populus tremuloides* 100, *Osmorhiza chilensis* 100, *Veratrum californicum* 50, *Osmorhiza occidentalis* 50, *Elymus cinereus* 50, *Carex pachystachya* 50

Non-forested vegetation unit

Unit No. 21

Number of relevés: 4

Faithful species: *Chrysothamnus viscidiflorus* 42.4, *Elymus cinereus* 40.0

Constant species: *Symphoricarpos oreophilus* 100, *Potentilla glandulosa* 100, *Hackelia micrantha* 100, *Bromus carinatus* 100, *Artemisia spiciformis* 100, *Elymus cinereus* 75, *Senecio serra* 75, *Geranium viscosissimum* 75, *Carex hoodii* 75, *Agastache urticifolia* 75

Dominant species: *Symphoricarpos oreophilus* 100, *Artemisia spiciformis* 100, *Bromus carinatus* 50, *Elymus cinereus* 50, *Senecio serra* 25, *Chrysothamnus viscidiflorus* 25, *Agastache urticifolia* 25

Group No. 22

Number of relevés: 5

Faithful species: *Poa arnowiae* 45.3, *Eriogonum heracleoides* 44.3, *Penstemon cyananthus* 44.2

Constant species: *Symphoricarpos oreophilus* 100, *Stipa nelsonii* 100, *Eriogonum umbellatum* 100, *Artemisia spiciformis* 100, *Stipa lettermanii* 80, *Potentilla glandulosa* 80, *Geranium viscosissimum* 80, *Bromus carinatus* 80, *Agastache urticifolia* 80, *Polemonium foliosissimum* 60, *Machaeranthera commixta* 60, *Hieracium cynoglossoides* 60, *Helianthella uniflora* 60, *Hackelia micrantha* 60, *Elymus trachycaulus* 60, *Delphinium nuttallianum* 60, *Balsamorhiza sagittata* 60, *Aster perelegans* 60, *Achillea millefolium* 60

Dominant species: *Artemisia spiciformis* 100, *Symphoricarpos oreophilus* 80, *Eriogonum umbellatum* 60, *Stipa lettermanii* 20, *Potentilla glandulosa* 20, *Bromus carinatus* 20, *Balsamorhiza sagittata* 20

Group No. 23

Number of relevés: 4

Faithful species:

Constant species: *Symphoricarpos oreophilus* 100, *Stipa nelsonii* 100, *Stipa lettermanii* 100, *Rudbeckia occidentalis* 100, *Potentilla glandulosa* 100, *Hackelia micrantha* 100, *Geranium viscosissimum* 100, *Eriogonum umbellatum* 100, *Elymus trachycaulus* 100, *Bromus carinatus* 100, *Artemisia spiciformis* 100, *Agastache urticifolia* 100, *Stellaria jamesiana* 75, *Senecio serra* 75, *Osmorhiza occidentalis* 75, *Erigeron speciosus* 75

Dominant species: *Agastache urticifolia* 100, *Rudbeckia occidentalis* 75, *Potentilla glandulosa* 50, *Artemisia spiciformis* 50, *Geranium viscosissimum* 50, *Symphoricarpos oreophilus* 25, *Lupinus sericeus* 25, *Bromus carinatus* 75, *Eriogonum umbellatum* 25, *Elymus trachycaulus* 25

Group No. 24

Number of relevés: 3

Faithful species: *Wyethia amplexicaulis* 89.1, *Poa bulbosa* 85.1, *Antennaria parvifolia* 80.1, *Poa secunda* 63.6, *Artemisia arbuscula* 47.9, *Koeleria macrantha* 46.4, *Hieracium cynoglossoides* 43.0, *Comandra umbellata* 41.0

Constant species: *Symphoricarpos oreophilus* 100, *Stipa lettermanii* 100, *Eriogonum umbellatum* 100, *Artemisia spiciformis* 100

Dominant species: *Wyethia amplexicaulis* 67, *Artemisia arbuscula* 67, *Symphoricarpos oreophilus* 33, *Artemisia spiciformis* 33

Group No. 25

Number of relevés: 4

Faithful species: *Carex multcostata* 86.3, *Ranunculus adoneus* 86.2, *Polygonum douglasii* 45.6

Constant species: *Stipa nelsonii* 100, *Rudbeckia occidentalis* 100, *Osmorhiza occidentalis* 100, *Ligusticum filicinum* 100, *Hackelia micrantha* 100, *Delphinium occidentale* 100, *Bromus carinatus* 100, *Senecio crassulus* 75, *Osmorhiza chilensis* 75, *Geranium viscosissimum* 75, *Galium bifolium* 75, *Collinsia parviflora* 75, *Aster integrifolius* 75, *Agastache urticifolia* 75

Dominant species: *Rudbeckia occidentalis* 75, *Ranunculus adoneus* 75, *Ligusticum filicinum* 50, *Delphinium occidentale* 25, *Stipa nelsonii* 25, *Bromus carinatus* 25, *Senecio crassulus* 25

Group No. 26

Number of relevés: 4

Faithful species: *Aster integrifolius* 52.2, *Heuchera parvifolia* 51.3, *Elymus lanceolatus* 47.2

Constant species: *Osmorhiza occidentalis* 100, *Hackelia micrantha* 100, *Geranium viscosissimum* 100, *Elymus trachycaulus* 100, *Agastache urticifolia* 100, *Achillea millefolium* 100, *Stipa nelsonii* 75, *Senecio crassulus* 75, *Potentilla glandulosa* 75, *Lomatium graveolens* 75, *Ligusticum filicinum* 75, *Helianthella uniflora* 75, *Erythronium grandiflorum* 75, *Eriogonum umbellatum* 75, *Erigeron speciosus* 75, *Bromus carinatus* 75, *Aster engelmannii* 75

Dominant species: *Agastache urticifolia* 50, *Rudbeckia occidentalis* 25, *Potentilla gracilis* 25, *Osmorhiza occidentalis* 25, *Lomatium graveolens* 25, *Helianthella uniflora* 25, *Geranium viscosissimum* 25, *Bromus carinatus* 25, *Artemisia spiciformis* 25

Group No. 27

Number of relevés: 3

Faithful species: *Clematis occidentalis* 67.3, *Linum kingii* 64.1, *Linum lewisii* 59.3, *Penstemon compactus* 58.9, *Linanthastrum nuttallii* 51.9

Constant species: *Symphoricarpos oreophilus* 100, *Penstemon leonardii* 100, *Lomatium graveolens* 100, *Leucopoa kingii* 100, *Aster engelmannii* 100, *Phacelia hastata* 67, *Helianthella uniflora* 67, *Berberis repens* 67, *Aster glaucodes* 67

Dominant species: *Lomatium graveolens* 33, *Linanthastrum nuttallii* 33, *Balsamorhiza sagittata* 33, *Aster glaucodes* 33, *Berberis repens* 33

Group No. 28

Number of relevés: 6

Faithful species: *Hymenoxys acaulis* 91.1, *Phlox hoodii* 91.0, *Synthyris pinnatifida* 73.6, *Linum kingii* 64.1, *Phlox pulvinata* 63.4, *Anemone multifida* 63.4, *Linum lewisii* 59.3, *Sedum lanceolatum* 57.2, *Lesquerella multiceps* 57.2, *Valeriana edulis* 49.3, *Penstemon compactus* 48.2, *Castilleja applegatei* 46.6, *Monardella odoratissima* 45.4, *Zigadenus paniculatus* 45.0, *Erigeron eatonii* 41.6, *Elymus elymoides* 40.3

Constant species: *Penstemon leonardii* 100, *Lomatium graveolens* 100, *Picea engelmannii* 83, *Solidago multiradiata* 67, *Ivesia gordonii* 67, *Elymus trachycaulus* 67, *Aster engelmannii* 67, *Arenaria congesta* 67, *Achillea millefolium* 67

Dominant species: *Linum lewisii* 50, *Lomatium graveolens* 17, *Linum kingii* 17

Group No. 29

Number of relevés: 5

Faithful species: *Apocynum androsaemifolium* 62.7, *Juncus parryi* 61.2, *Epilobium canum* 58.8, *Solidago multiradiata* 57.9, *Arenaria congesta* 53.4, *Viguiera multiflora* 48.4, *Ivesia gordonii* 45.1, *Cystopteris fragilis* 44.9, *Sedum debile* 39.8

Constant species: *Potentilla glandulosa* 100, *Penstemon leonardii* 100, *Eriogonum umbellatum* 100, *Achillea millefolium* 100, *Stipa lettermanii* 80, *Pinus flexilis* 80, *Erigeron speciosus* 80, *Carex rossii* 80, *Leucopoa kingii* 60, *Elymus trachycaulus* 60, *Castilleja applegatei* 60, *Bromus carinatus* 60, *Berberis repens* 60

Dominant species: *Juncus parryi* 20

Group No. 30

Number of relevés: 5

Faithful species: *Monardella odoratissima* 55.6, *Eriogonum caespitosum* 52.1, *Penstemon humilis* 48.1, *Ivesia gordonii* 45.1, *Artemisia dracunculus* 44.2, *Heuchera parvifolia* 40.5

Constant species: *Leucopoa kingii* 100, *Symphoricarpos oreophilus* 80, *Stipa lettermanii* 80, *Penstemon leonardii* 80, *Machaeranthera commixta* 80, *Eriogonum umbellatum* 80, *Potentilla glandulosa* 60, *Phacelia hastata* 60, *Helianthella uniflora* 60, *Geranium viscosissimum* 60, *Erigeron speciosus* 60, *Agastache urticifolia* 60

Dominant species: *Monardella odoratissima* 80, *Eriogonum umbellatum* 40, *Artemisia dracunculus* 20

Group No. 31

Number of relevés: 6

Faithful species: *Artemisia tridentata* 42.4, *Comandra umbellata* 41.0

Constant species: *Artemisia tridentata* 100, *Symphoricarpos oreophilus* 100, *Elymus spicatus* 100, *Balsamorhiza sagittata* 100, *Penstemon leonardii* 83, *Eriogonum umbellatum* 83, *Berberis repens* 83, *Machaeranthera commixta* 67, *Crepis acuminata* 67

Dominant species: *Artemisia tridentata* 100, *Elymus spicatus* 67, *Balsamorhiza sagittata* 67, *Symphoricarpos oreophilus* 17, *Poa secunda* 17, *Chrysothamnus viscidiflorus* 17, *Ceanothus velutinus* 17, *Artemisia arbuscula* 17, *Berberis repens* 17

Group No. 32

Number of relevés: 3

Faithful species: *Carex rostrata* 69.6, *Carex nebrascensis* 69.6, *Polygonum bistortoides* 57.2, *Salix wolfii* 54.4, *Pedicularis groenlandica* 54.4, *Juncus ensifolius* 54.4, *Saxifraga odontoloma* 52.8, *Juncus balticus* 46.1, *Salix drummondiana* 45.1, *Juncus mertensianus* 42.4, *Deschampsia cespitosa* 42.4, *Carex raynoldsii* 42.4, *Betula glandulosa* 42.4, *Salix boothii* 41.1

Constant species: *Geranium richardsonii* 100, *Picea engelmannii* 67, *Mimulus guttatus* 67, *Lonicera involucrata* 67, *Habenaria dilatata* 67

Dominant species: *Salix boothii* 100, *Carex nebrascensis* 100, *Salix drummondiana* 67, *Juncus balticus* 67, *Carex rostrata* 33, *Betula glandulosa* 33

Group No. 33

Number of relevés: 4

Faithful species: *Betula occidentalis* 81.0, *Salix lutea* 81.0, *Juncus balticus* 71.2, *Carex rostrata* 69.6, *Carex nebrascensis* 69.6, *Salix lasiandra* 66.7, *Salix exigua* 66.0, *Cornus sericea* 58.1, *Habenaria dilatata* 54.3, *Equisetum arvense* 54.3, *Phleum pratense* 48.5, *Salix drummondiana* 45.1, *Ranunculus orthorhynchus* 44.9, *Veratrum californicum* 42.8, *Salix boothii* 41.1, *Mimulus guttatus* 40.6

Constant species: *Geranium richardsonii* 100

Dominant species: *Salix lasiandra* 100, *Salix boothii* 100, *Carex rostrata* 100, *Carex nebrascensis* 100, *Veratrum californicum* 75, *Salix drummondiana* 75, *Juncus balticus* 75, *Salix lutea* 25, *Equisetum arvense* 25, *Betula occidentalis* 25

Group No. 34

Number of relevés: 2

Faithful species: *Fragaria virginiana* 78.8, *Ranunculus macounii* 60.3, *Trifolium pratense* 56.6, *Equisetum arvense* 54.3, *Arnica chamissonis* 54.3, *Geum macrophyllum* 51.4, *Smilacina stellata* 46.7, *Castilleja miniata* 46.0, *Salix drummondiana* 45.1, *Veratrum californicum* 42.8, *Lonicera involucrata* 42.8, *Poa reflexa* 41.9, *Taraxacum officinale* 41.3, *Salix boothii* 41.1

Constant species: *Thalictrum fendleri* 100, *Rudbeckia occidentalis* 100, *Osmorhiza chilensis* 100, *Geranium richardsonii* 100, *Achillea millefolium* 100

Dominant species: *Salix drummondiana* 100, *Salix boothii* 100, *Equisetum arvense* 100, *Salix lasiandra* 50, *Ranunculus macounii* 50, *Geranium richardsonii* 50, *Betula occidentalis* 50

CURRICULUM VITAE

Antonin Kusbach**Personal information**

Department of Wildland Resources and Ecology Center
 Utah State University
 Email: tony.kusbach@usu.edu

Education

Ph.D. 2006	Ecology	Utah State University
B.S. 1982	Forestry	Mendel University, Brno, Czech Republic

Research interests

Forest ecosystem classification
 Application of forest ecosystem classification in ecosystem management

Grants and funding

Kusbach, A., Shaw, J.D. Long, J.N. Forest ecosystem classification in the Intermountain West. Forest Inventory and Analysis Program, USDA Forest Service. 2010-2013.

Awards and honors

North American Forest Ecology Workshop 2009, the best student presentation in section "Ecosystem Classification". <http://www.nafew2009.org/>

Utah State University, College of Natural Resources, Ecology Center Graduate Research Support Award 2009-2010. <http://www.usu.edu/ecology/>

Utah State University, College of Natural Resources, Ecology Center Graduate Research Support Award 2007-2008.

Utah State University, Graduate Student Senate, Enhancement Award 2008

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